# SEAMAP Reef Fish Video Survey: Relative Indices of Abundance of Red Snapper July 2012 

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SEAMAP Reef Fish Video Survey:
Relative Indices of Abundance of Red Snapper
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## Introduction

The primary objective of the annual Southeast Area Monitoring and Assessment Program (SEAMAP) reef fish video survey is to provide an index of the relative abundances of fish species associated with topographic features (e.g reefs, banks, and ledges) located on the continental shelf of the Gulf of Mexico (GOM) from Brownsville, TX to the Dry Tortugas, FL (Figures 1 and 2). Secondary objectives include quantification of habitat types sampled (video and side-scan), and collection of environmental data throughout the survey. Because the survey is conducted on topographic features the species assemblages targeted are typically classified as reef fish (e.g. red snapper, Lutjanus campechanus), but occasionally fish more commonly associated with pelagic environments are observed (e.g. hammerhead shark, Sphyrna lewini). The survey has been executed from 1992-1997, 2001-2002, and 2004-2011 and historically takes place from May August. The 2001 survey was abbreviated due to ship scheduling, during which, the only sites that were completed were located in the western Gulf of Mexico. Types of data collected on the survey include diversity, abundance (minimum count), fish length, habitat type, habitat coverage, and bottom topography. The size of fish sampled with the video gear is species specific however red snapper sampled over the history of the survey had fork lengths ranging from $146-917 \mathrm{~mm}$, and mean annual fork lengths ranging from $370.6-593.1 \mathrm{~mm}$. Age and reproductive data cannot be collected with the camera gear but beginning with the 2012 survey, a vertical line component will be coupled with the video drops to collect hard parts, fin clips, and gonads.

## Methods

## Sampling design

Total reef area available to select survey sites from is approximately $1771 \mathrm{~km}^{2}$, of which $1244 \mathrm{~km}^{2}$ is located in the eastern GOM and $527 \mathrm{~km}^{2}$ in the western GOM. The large size of the survey area necessitates a two-stage sampling design to minimize travel times between stations. The first-stage uses stratified random sampling to select blocks that are 10 minutes of latitude by 10 minutes of longitude in dimension (Figures 1 and 2). The block strata were defined by geographic region (4 regions: South Florida, Northeast Gulf, Louisiana-Texas Shelf, and South Texas), and by total reef habitat area contained in the block (blocks $\leq 20 \mathrm{~km}^{2}$ reef, block $>20 \mathrm{~km}^{2}$ reef). There are a total of 7 strata. A 0.1 by 0.1 mile grid is then overlaid onto the reef area contained within a given block and the ultimate sampling sites (second stage units) are randomly selected from that grid.


Figure 1. SEAMAP reef fish video survey sample blocks located in the eastern Gulf of Mexico.


Figure 2. SEAMAP reef fish video survey sample blocks located in the western Gulf of Mexico.

## Data reduction

Various limitations either in design, implementation, or performance of gear causes limitations in calculating minimum counts and are therefore dropped from the design-based indices development and analysis as follows. In 1992, each fish was counted every time it came into view over the entire record time and the total of all these counts was the maximum count. Maximum count methodologies are not preferred and the 1992 video tapes were destroyed during Hurricane Katrina and cannot be re-viewed, so 1992 data is excluded from analyses (unknown number of stations). The 2001 survey was abbreviated due to ship scheduling, during which, the only sites that were completed were located in the western GOM. Because of the spatial imbalance associated with data gathered in 2001, that entire year has been dropped ( 80 total sites). Stratum 1 (South Florida) and stratum 7 (S. Texas) are blocks that contain very little reef and were not consistently chosen for sampling and were also dropped ( 184 total sites). Occasionally tapes are unable to be read (i.e. organisms cannot be identified to species) for the following reasons including: 1) camera views are more than $50 \%$ obstructed, 2) sub-optimal lighting conditions, 3) increased backlighting, 4) increased turbidity, 5) cameras out of focus, 6) cameras failed to film. In all of these cases the station is flagged as 'XX' in the data set and dropped (190 total sites). Sites that did not receive a stratum assignment are also dropped (62). By these criteria the data set is reduced 4744 down to 4228 sites analyzed.

## Gear and deployment

The SEAMAP reef fish survey has employed several camcorders in underwater housings since 1992. Sony VX2000 DCR digital camcorders mounted in Gates PD150M underwater housings were used from 2002 to 2005 and Sony PD170 camcorders during the years 2006 and 2007. In 2008 a stereo video camera system was developed and assembled at the NMFS Mississippi Laboratories Stennis Space Center Facility and has been used in all subsequent surveys. The stereo video unit consists of a digital stereo still camera head, digital video camera, CPU, and hard drive mounted in an aluminum housing. All of the camcorder housings we have used were rated to a maximum depth of 150 meters while the stereo camera housings are rated to 600 meters. Stereo cameras are mounted orthogonally at a height of 50 cm above the bottom of the pod and the array is baited with squid during deployment.

At each sampling site the stereo video unit is deployed for 40 minutes total, however the cameras and CPU delay filming for 5 minutes to allow for descent to the bottom, and settling of suspended sediment following impact. Once turned on, the cameras film for approximately 30 minutes before shutting off and retrieval of the array. During camera deployment the vessel drifts away from the site and a CTD cast executed, collecting water depth, temperature, conductivity, and transmissivity from the surface to the maximum depth. Seabird units are the standard onboard NOAA vessels however the model employed was vessel/cruise dependent.

## Video tape viewing

One video tape from each station is selected for viewing out of four possible. If all four video cameras face reef fish habitat and are in focus, tape selection is random. Videos are viewed for twenty minutes starting from the time when the view clears from suspended sediment. Viewers identify, and enumerate all species to the lowest taxonomic level during the 20 minute viewable segment. From 1993-2007 the time when each fish entered and left the field of view was recorded a procedure referred to as time in - time out (TITO) and from these data a minimum count was calculated. The minimum count is the maximum number of individuals of a selected
taxon in the field of view at one instance. Each 20 minute video is evaluated to determine the highest minimum count observed during a 20 minute recording. The 2008-2011 digital video allows the viewer to record a frame number or time stamp of the image when the maximum number of individuals of a species occurred, along with the number of taxon identified in the image but does not use the TITO method. Both the TITO and current viewing procedure result in the minimum count estimator of relative abundance. Minimum count methodology is preferred because it prevents counting the same fish more than once.

## Fish length measurement

Beginning in 1995 fish lengths were measured from video using lasers attached on the camera system with known geometry. However, the frequency of hitting targets with the laser is low and precluded estimating size frequency distributions. Additionally, the same fish can be measured more than once at a given station. So, the lengths measured provide the range of sizes observed. The stereo cameras used in 2008-2010 allow size estimation from fish images. The Vision Measurement System (Geometrics Inc.) was used to estimate size of red snapper. We estimated a length frequency distribution by weighting station length frequencies by station Minimum Counts (Figure 30, 32).

## Model based indices

Delta-lognormal modeling methods were used to estimate relative abundance indices for red snapper (Lo et al. 1992). The main advantage of using this method is allowance for the probability of zero catch (Ortiz et al. 2000). The index computed by this method is a mathematical combination of yearly abundance estimates from two distinct generalized linear models: a binomial (logistic) model which describes proportion of positive abundance values (i.e. presence/absence) and a lognormal model which describes variability in only the nonzero abundance data (Lo et al. 1992).

The delta-lognormal index of relative abundance $\left(I_{y}\right)$ as described by Lo et al. (1992) was estimated as:
(2) $\quad I_{y}=c_{y} p_{y}$,
where $c_{y}$ is the estimate of mean CPUE for positive catches only for year $y$, and $p_{y}$ is the estimate of mean probability of occurrence during year $y$. Both $c_{y}$ and $p_{y}$ were estimated using generalized linear models. Data used to estimate abundance for positive catches (c) and probability of occurrence $(p)$ were assumed to have a lognormal distribution and a binomial distribution, respectively, and modeled using the following equations:
(3) $\ln (c)=X \beta+\varepsilon$
and

$$
\begin{equation*}
p=\frac{e^{\mathrm{X}^{\beta}+\varepsilon}}{1+e^{\mathrm{X} \beta+\varepsilon}}, \tag{4}
\end{equation*}
$$

respectively, where $c$ is a vector of the positive catch data, $p$ is a vector of the presence/absence data, $X$ is the design matrix for main effects, $\beta$ is the parameter vector for main effects, and $\varepsilon$ is a vector of independent normally distributed errors with expectation zero and variance $\sigma^{2}$. Therefore, $c_{y}$ and $p_{y}$ were estimated as least-squares means for each year along with their corresponding standard errors, $\operatorname{SE}\left(c_{y}\right)$ and $\operatorname{SE}\left(p_{y}\right)$, respectively. From these estimates, $I_{y}$ was calculated, as in equation (1), and its variance calculated as:

$$
\begin{equation*}
V\left(I_{y}\right) \approx V\left(c_{y}\right) p_{y}^{2}+c_{y}^{2} V\left(p_{y}\right)+2 c_{y} p_{y} \operatorname{Cov}(c, p), \tag{5}
\end{equation*}
$$

where:

$$
\begin{equation*}
\left.\operatorname{Cov}(c, p) \approx \rho_{\mathrm{c}, \mathrm{p}} \mid \operatorname{SE}\left(c_{y}\right) \operatorname{SE}\left(p_{y}\right)\right], \tag{6}
\end{equation*}
$$

and $\rho_{\mathrm{c}, \mathrm{p}}$ denotes correlation of $c$ and $p$ among years.
The submodels of the delta-lognormal model were built using a backward selection procedure based on type 3 analyses with an inclusion level of significance of $\alpha=0.05$. Binomial submodel performance was evaluated using AIC, while the performance of the lognormal submodel was evaluated based on analyses of residual scatter and QQ plots in addition to AIC. Variables that could be included in the submodels were: Year (1987-2011).

## Design based indices

A delta-lognormal modeling approach (Lo et al., 1992) was used to develop abundance indices. Independent variables used in the model were year, region and depth. Region is divided into east and west at 89.15 west longitude. The GENMOD procedure in SAS (v.9.2) was used to conduct separate forward stepwise regressions on the binomial and lognormal sub-models to determine which variables to retain for use in fitting the delta lognormal model. Only variables that reduced model deviance by at least $1 \%$ with a type 3 analysis level of significance of $\alpha=0.05$ were retained. The GLIMMIX and MIXED procedures in SAS (v. 9.2) were used to develop the binomial and lognormal sub-models, respectively. A backward selection procedure was used to determine which variables retained from the GENMOD procedure were to be included into each final sub-model based on a type 3 analyses with a level of significance for inclusion of $\alpha=0.01$. Year was including in all terminal models regardless of significance, while region and depth were retained in both the binomial and lognormal sub-models. The estimates from each model were weighted using the stratum area, and separate covariance structures were developed for each survey year. For the binomial models, a logistic-type mixed model was employed.

## Results

Red snapper were observed at banks in both the western and eastern GOM (Figures 3 16), and the spatial distributions observed are highly reflective of the reef sampling universe used to select sampling sites (Figures 1-2). Gaps in habitat level information exist in central Florida, Mississippi river delta region, and portions of the Texas coast. In most years the survey shows good coverage in the defined sampling universe, and coverage improved through time as the sampling universe expanded and more sites were added to the survey. Reef blocks from coastal Texas are often not selected for sampling due to small spatial coverage of reef, and frequent high winds and rough sea states during the spring/early summer sampling season.

Design based analysis retained year, region and depth in the binomial and log-normal GOM-wide sub-model. Design based red snapper proportion positives ranged from 0.1 (1993) to 0.41 (2010) with a reported value of 0.35 in 2011 (Table 1, Figure 17), while standardized index of abundance ranged from 0.5 (1995) to 1.88 (2010), and reported a value of 1.84 in 2011 (Table 2, Figure 18). Model based analysis shows GOM wide standardized index trends are similar to the design based runs and ranged from 0.29 (1994) to 2.35 (2011)(Table 12, Figure 43).

Design based analysis retained year, region and depth in the binomial and log-normal eastGOM sub-model. Design based east-GOM red snapper proportion positives ranged from 0.027 (1994) to 0.309 (2011) (Table 1, Figure 17), and the standardized index of abundance ranged from 0.24 (1994) to 2.1 (2011) (Table 2, Figure 18). Model based east-GOM standardized index values trended similarly to design based runs and ranged from 0.04 (1994) to 3.23 (2011) (Table 14, Figure 45). Annual mean fork lengths for east GOM red snapper have ranged from 370.6 to 514.8 mm in the east in 2009 and 2010 respectively (Table 15).

Design based analysis retained year, region and depth in the binomial and log-normal west-GOM sub-model. Design based red snapper proportion positive ranged from 0.15 (1993) to 0.57 (2010) with a reported value of 0.41 in 2011 (Table 1, Figure 17), while standardized index values ranged from 0.59 (1996) to 2.03 (2010) (Table 2, Figure 18). Model based analysis of west GOM standardized index values show similar trends as the design based runs and ranged from 0.53 (1994) to 2.23 (2011) (Table 13, Figure 44). Annual mean fork lengths for west GOM red snapper have ranged from 200 to 917 mm in the east in 2005 and 2004 respectively (Table 15).

Design and model based output for proportion positives, lo-index, and standardized index values from 2010 and 2011 are at historical highs for the survey. Gulf wide and west GOM values for proportion positives, lo-index, and standardized index output show a slight decrease from 2010 to 2011, while east GOM values have continued to increase since 2008. There appears to be some evidence of year classes cycling through as age 3 size fish move onto the reef, for instance 2004 is noted as a strong year class and 2007 we see a spike in min-count in both the east and west data (other strong years classes include 2000 and 2006).

## Literature cited

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Lo, N. C. H., L.D. Jacobson, and J.L. Squire. 1992. Indices of relative abundance from fish spotter data based on delta-lognormal models. Can. J. Fish. Aquat. Sci. 49: 2515-1526.

Figure 3. Spatial distribution of red snapper observed and associated min-count values during the 1993 reef fish video survey.


Figure 4. Spatial distribution of red snapper observed and associated min-count values during the 1994 reef fish video survey.


Figure 5. Spatial distribution of red snapper observed and associated min-count values during the 1995 reef fish video survey.


Figure 6. Spatial distribution of red snapper observed and associated min-count values during the 1996 reef fish video survey.


Figure 7. Spatial distribution of red snapper observed and associated min-count values during the 1997 reef fish video survey.


Figure 8. Spatial distribution of red snapper observed and associated min-count values during the 2002 reef fish video survey.


Figure 9. Spatial distribution of red snapper observed and associated min-count values during the 2004 reef fish video survey.


Figure 10. Spatial distribution of red snapper observed and associated min-count values during the 2005 reef fish video survey.


Figure 11. Spatial distribution of red snapper observed and associated min-count values during the 2006 reef fish video survey.


Figure 12. Spatial distribution of red snapper observed and associated min-count values during the 2007 reef fish video survey.


Figure 13. Spatial distribution of red snapper observed and associated min-count values during the 2008 reef fish video survey.


Figure 14. Spatial distribution of red snapper observed and associated min-count values during the 2009 reef fish video survey.


Figure 15. Spatial distribution of red snapper observed and associated min-count values during the 2010 reef fish video survey.


Figure 16. Spatial distribution of red snapper observed and associated min-count values during the 2011 reef fish video survey.


Table 1. Proportion of stations capturing red snapper (proportion positive) by region and year for the design based model.

| Year | GOM wide | East GOM | West GOM |
| :---: | :---: | :---: | :---: |
| 1993 | 0.1040936 | 0.0526316 | 0.1555556 |
| 1994 | 0.1284742 | 0.0266667 | 0.2325581 |
| 1995 | 0.1467432 | 0.0529801 | 0.2405063 |
| 1996 | 0.1500716 | 0.0583942 | 0.2339181 |
| 1997 | 0.2859206 | 0.0337838 | 0.5413534 |
| 2002 | 0.2683614 | 0.1552795 | 0.3814433 |
| 2004 | 0.2537324 | 0.2013423 | 0.3061224 |
| 2005 | 0.2908075 | 0.2125984 | 0.3823529 |
| 2006 | 0.1472674 | 0.0932836 | 0.2058824 |
| 2007 | 0.2747524 | 0.1677419 | 0.3924051 |
| 2008 | 0.2011923 | 0.0994152 | 0.3114754 |
| 2009 | 0.2575986 | 0.1585366 | 0.3674699 |
| 2010 | 0.4155844 | 0.3045685 | 0.5714286 |
| 2011 | 0.3512597 | 0.3096774 | 0.4150943 |

Figure 17. Comparison of the east, west, and GOM-wide proportion of positive red snapper catch sites for the design based model.


Table 2. Lo and standardized index values by region and year for the design based model.

|  | Lo index |  |  | Standardized index |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | GOM wide | East <br> GOM | West GOM | GOM wide | East GOM | West GOM |
| 1993 | 0.20967 | 0.11488 | 0.40422 | 0.51427 | 0.4397 | 0.61146 |
| 1994 | 0.23845 | 0.06248 | 0.53307 | 0.58487 | 0.23916 | 0.80638 |
| 1995 | 0.20462 | 0.09399 | 0.39582 | 0.5019 | 0.35977 | 0.59876 |
| 1996 | 0.24199 | 0.11507 | 0.39036 | 0.59355 | 0.44043 | 0.5905 |
| 1997 | 0.52027 | 0.16181 | 1.0229 | 1.27611 | 0.61932 | 1.54734 |
| 2002 | 0.51478 | 0.39263 | 0.70069 | 1.26265 | 1.50278 | 1.05993 |
| 2004 | 0.54023 | 0.41162 | 0.60728 | 1.32506 | 1.57546 | 0.91864 |
| 2005 | 0.50115 | 0.36563 | 0.7256 | 1.22921 | 1.39945 | 1.09762 |
| 2006 | 0.22005 | 0.1762 | 0.28464 | 0.53973 | 0.67442 | 0.43057 |
| 2007 | 0.41428 | 0.26069 | 0.79635 | 1.01614 | 0.99781 | 1.20463 |
| 2008 | 0.23233 | 0.1815 | 0.40457 | 0.56985 | 0.69469 | 0.612 |
| 2009 | 0.3526 | 0.23474 | 0.6521 | 0.86485 | 0.89847 | 0.98643 |
| 2010 | 0.76841 | 0.53619 | 1.34125 | 1.88474 | 2.05228 | 2.02892 |
| 2011 | 0.74897 | 0.5503 | 0.99611 | 1.83707 | 2.10627 | 1.50682 |

Figure 18. Comparison of the east, west, and GOM-wide standardized indices of red snapper abundance.


Table 3. Iteration history (a), fit statistics (b), type III tests (c), and over-dispersion diagnostics of the GLIMMIX binomial on proportion positives for the GOM-wide model.
a

| Iteration History |  |  |  |
| ---: | ---: | ---: | ---: | ---: |
| Iteration | Evaluations | -2 Res Log Like | Criterion |
| 1 | 1 | 21956.42410392 | 0.00000000 |

b

| Fit Statistics |  |
| :--- | :--- |
| -2 Res Log Likelihood | 21956.4 |
| AIC (smaller is better) | 21984.4 |
| AICC (smaller is better) | 21984.5 |
| BIC (smaller is better) | 22073.9 |

c

| Type 3 Tests of Fixed Effects |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: |
| Num |  |  |  |  |  |  |  | Den |  |  |
| Effect | DF | DF | Chi-Square | F Value | Pr $>$ ChiSq | Pr $>F$ |  |  |  |  |
| year | 13 | 1466 | 128.02 | 9.80 | $<.0001$ | $<.0001$ |  |  |  |  |
| zone | 1 | 4178 | 70.94 | 70.94 | $<.0001$ | $<.0001$ |  |  |  |  |
| depth | 1 | 3411 | 120.16 | 120.16 | $<.0001$ | $<.0001$ |  |  |  |  |

d

| Description | Value |
| :--- | ---: |
| Deviance | 741.5532 |
| Scaled Deviance | 3572.6273 |
| Pearson Chi-Square | 788.0838 |
| Scaled Pearson Chi-Square | 3796.8002 |
| Extra-Dispersion Scale | 0.2076 |

Figure 19. GOM-wide observed versus proportion positive for design based simulation.
Delta lognormal mincount for red snapper Diagnostic plots: 1) Obs vs Pred Proport Posit


Figure 20. GOM-wide chi-square residuals of proportion positive design based model.
Delta lognormal mincount for red snapper
Chisq Residuals proportion positive


Table 4. GOM-wide red snapper lo and standardized index of abundance values by year design based model.

| Year | $N$ | LoIndex | StdIndex | SE | CV | LCL | UCL |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1993 | 159 | 0.20967 | 0.51427 | 0.079286 | 0.37815 | 0.24754 | 1.0684 |
| 1994 | 125 | 0.23845 | 0.58487 | 0.092332 | 0.38722 | 0.27695 | 1.23514 |
| 1995 | 230 | 0.20462 | 0.5019 | 0.051326 | 0.25083 | 0.30624 | 0.82257 |
| 1996 | 322 | 0.24199 | 0.59355 | 0.051993 | 0.21485 | 0.3881 | 0.90777 |
| 1997 | 297 | 0.52027 | 1.27611 | 0.084398 | 0.16222 | 0.92449 | 1.76148 |
| 2002 | 258 | 0.51478 | 1.26265 | 0.080064 | 0.15553 | 0.92682 | 1.72016 |
| 2004 | 198 | 0.54023 | 1.32506 | 0.095156 | 0.17614 | 0.93413 | 1.87959 |
| 2005 | 407 | 0.50115 | 1.22921 | 0.066427 | 0.13255 | 0.94405 | 1.6005 |
| 2006 | 418 | 0.22005 | 0.53973 | 0.045024 | 0.20461 | 0.35997 | 0.80924 |
| 2007 | 489 | 0.41428 | 1.01614 | 0.056484 | 0.13634 | 0.77459 | 1.33302 |
| 2008 | 309 | 0.23233 | 0.56985 | 0.05134 | 0.22098 | 0.36821 | 0.88189 |
| 2009 | 430 | 0.3526 | 0.86485 | 0.05222 | 0.1481 | 0.64417 | 1.16114 |
| 2010 | 329 | 0.76841 | 1.88474 | 0.091725 | 0.11937 | 1.48571 | 2.39094 |
| 2011 | 440 | 0.74897 | 1.83707 | 0.08251 | 0.11016 | 1.47478 | 2.28836 |

Table 5. Fit statistics (a), and type III tests (b) of the GLM on positive catches for the GOM-wide design based model.
a $\qquad$

| Fit Statistics |  |
| :--- | :--- |
| -2 Res Log Likelihood | 2507.8 |
| AIC (smaller is better) | 2509.8 |
| AICC (smaller is better) | 2509.8 |
| BIC (smaller is better) | 2514.6 |

b

| Type 3 Tests of Fixed Effects |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Effect | $\begin{gathered} \mathrm{Num} \\ D F \end{gathered}$ | $\begin{gathered} \text { Den } \\ D F \end{gathered}$ | F Value | Pr $>$ F |
| year | 13 | 954 | 3.05 | 0.0002 |
| zone | 1 | 954 | 0.03 | 0.8563 |
| depth | 1 | 954 | 12.06 | 0.0005 |

Figure 21. GOM-wide observed versus standardized mincount for design based model.
Delta lognormal mincount for red snapper Observed and Standardized mincount (95\% Cl)

## STDmincount



$$
\begin{array}{lll}
\text { PLOT } & \leftrightarrow \text { STDmincount } & \Leftrightarrow \leftrightarrow \text { LCI } \\
& \Leftrightarrow \forall \text { UCl } & \mapsto \text { obsmincount }
\end{array}
$$

Figure 22. GOM-wide observed versus predicted mincount of positive data for design based model.

Delta lognormal mincount for red snapper Diagnostic plots: 2) Obs vs Pred mincount of Posit only MINCOUNT


Figure 23. GOM-wide observed versus predicted mincount for design based model.
Delta lognormal mincount for red snapper
Diagnostic plots: 3) Obs vs Pred mincount Input units


Figure 24. GOM wide residuals of positive mincounts by year for design based model.
Delta lognormal mincount for red snapper Residuals positive mincounts * Year


Figure 25 GOM-wide residuals distribution from positive mincount design based model.
Delta lognormal mincount for red snapper
Residuals positive mincount Distribution


Figure 26 GOM-wide qqplot of residuals of positive mincounts from design based model.
Delta lognormal mincount for red snapper QQplot Residuals Positive mincount rates


Table 6. Iteration history (a), fit statistics (b), type III tests (c), and over-dispersion diagnostics of the GLIMMIX binomial on proportion positives for the east GOM model.
a

| Iteration History |  |  |  |
| ---: | ---: | ---: | ---: |
| Iteration | Evaluations | -2 Res Log Like | Criterion |
| 1 | 1 | 13797.30604244 | 0.00000000 |

b

| Fit Statistics |  |
| :--- | :---: |
| -2 Res Log Likelihood | 13797.3 |
| AIC (smaller is better) | 13825.3 |
| AICC (smaller is better) | 13825.5 |
| BIC (smaller is better) | 13907.9 |

c

| Type 3 Tests of Fixed Effects |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Effect | DF | Den |  |  |  |  |
| year | 13 | 875 | 80.77 | 6.16 | $<.0001$ | $<.0001$ |
| depth | 1 | 1930 | 113.95 | 113.95 | $<.0001$ | $<.0001$ |

d

| Description | Value |
| :--- | ---: |
| Deviance | 684.1846 |
| Scaled Deviance | 1990.4576 |
| Pearson Chi-Square | 807.0360 |
| Scaled Pearson Chi-Square | 2347.8620 |
| Extra-Dispersion Scale | 0.3437 |

Figure 27. Observed versus predicted proportion positive from east GOM design based model.
Delta lognormal mincount for red snapper
Diagnostic plots: 1) Obs vs Pred Proport Posit


Figure 28. Chi-square residuals of proportion positives from east GOM design based model.
Delta lognormal mincount for red snapper Chisq Residuals proportion positive


Table 7. East GOM red snapper lo and standardized index of abundance by year for design based model.

| Year | $N$ | Lolndex | StdIndex | SE | CV | LCL | UCL |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1993 | 114 | 0.11488 | 0.4397 | 0.064842 | 0.56443 | 0.15359 | 1.25874 |
| 1994 | 75 | 0.06248 | 0.23916 | 0.058254 | 0.9323 | 0.04917 | 1.16316 |
| 1995 | 151 | 0.09399 | 0.35977 | 0.041984 | 0.44667 | 0.1533 | 0.84428 |
| 1996 | 137 | 0.11507 | 0.44043 | 0.04905 | 0.42627 | 0.19452 | 0.9972 |
| 1997 | 148 | 0.16181 | 0.61932 | 0.087341 | 0.53978 | 0.22527 | 1.70268 |
| 2002 | 161 | 0.39263 | 1.50278 | 0.095485 | 0.24319 | 0.93044 | 2.42719 |
| 2004 | 149 | 0.41162 | 1.57546 | 0.090826 | 0.22066 | 1.01864 | 2.43665 |
| 2005 | 254 | 0.36563 | 1.39945 | 0.065864 | 0.18014 | 0.9789 | 2.00067 |
| 2006 | 268 | 0.1762 | 0.67442 | 0.045942 | 0.26073 | 0.40381 | 1.12638 |
| 2007 | 310 | 0.26069 | 0.99781 | 0.050659 | 0.19432 | 0.67892 | 1.46647 |
| 2008 | 171 | 0.1815 | 0.69469 | 0.061778 | 0.34038 | 0.35827 | 1.34702 |
| 2009 | 246 | 0.23474 | 0.89847 | 0.050221 | 0.21394 | 0.58851 | 1.37169 |
| 2010 | 197 | 0.53619 | 2.05228 | 0.088657 | 0.16535 | 1.4777 | 2.85027 |
| 2011 | 310 | 0.5503 | 2.10627 | 0.080644 | 0.14654 | 1.57363 | 2.8192 |

Table 8. Fit statistics (a), and type III tests (b) of the GLM on positive catches for the east GOM design based model.
a $\qquad$

| Fit Statistics |  |
| :--- | :--- |
| -2 Res Log Likelihood | 1033.2 |
| AIC (smaller is better) | 1035.2 |
| AICC (smaller is better) | 1035.2 |
| BIC (smaller is better) | 1039.2 |

b

| Type 3 Tests of Fixed Effects |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Effect | $\begin{gathered} N u m \\ D F \end{gathered}$ | $\begin{gathered} \text { Den } \\ D F \end{gathered}$ | F Value | Pr $>$ F |
| year | 13 | 412 | 2.11 | 0.0128 |
| depth | 1 | 412 | 4.59 | 0.0327 |

Figure 29. Observed and standardized mincounts from east GOM design based model.
Delta lognormal mincount for red snapper Observed and Standardized mincount (95\% Cl)


Figure 30. Observed versus predicted mincounts from east GOM design based model.
Delta lognormal mincount for red snapper
Diagnostic plots: 2) Obs vs Pred mincount of Posit only
MINCOUNT


PLOT $\bullet$ MINCOUNT $\leftrightarrow-\Delta \otimes$ bc_cpu

Figure 31. Observed versus predicted mincounts from east GOM design based models
Delta lognormal mincount for red snapper Diagnostic plots: 3) Obs vs Pred mincount Input units


$$
\text { PLOT } \quad \cdots \text { MINCOUNT } \quad \leftrightarrow-\diamond \diamond \text { index }
$$

Figure 32 Residuals of positive mincounts for east GOM design based model.
Delta lognormal mincount for red snapper Residuals positive mincounts * Year


Figure 33. Positive mincount distribution from east GOM design based model.
Delta lognormal mincount for red snapper
Residuals positive mincount Distribution


Figure 34. QQ plot of positive mincounts from east GOM design based model.
Delta lognormal mincount for red snapper QQplot Residuals Positive mincount rates


Table 9. Iteration history (a), fit statistics (b), type III tests (c), and over-dispersion diagnostics of the GLIMMIX binomial on proportion positives for the west GOM model.
a

| Iteration History |  |  |  |  |
| ---: | ---: | ---: | ---: | :---: |
| Iteration | Evaluations | -2 Res Log Like | Criterion |  |
| 1 | 1 | 7112.76125721 | 0.00000000 |  |

b

| Fit Statistics |  |
| :--- | :--- |
| -2 Res Log Likelihood | 7112.8 |
| AIC (smaller is better) | 7140.8 |
| AICC (smaller is better) | 7141.0 |
| BIC (smaller is better) | 7215.5 |

c
Type 3 Tests of Fixed Effects

|  | Num | Den |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Effect | $D F$ | $D F$ | Chi-Square | F Value | Pr $>$ ChiSq | $\operatorname{Pr}>F$ |
| year | 13 | 463 | 71.29 | 5.39 | $<.0001$ | $<.0001$ |
| depth | 1 | 1191 | 5.72 | 5.72 | 0.0167 | 0.0169 |

d

| Description | Value |
| :--- | ---: |
| Deviance | 819.0246 |
| Scaled Deviance | 1751.3653 |
| Pearson Chi-Square | 684.1130 |
| Scaled Pearson Chi-Square | 1462.8763 |
| Extra-Dispersion Scale | 0.4676 |

Figure 35. Observed versus predicted proportion positive from west GOM design based model.
Delta lognormal mincount for red snapper
Diagnostic plots: 1) Obs vs Pred Proport Posit


Figure 36. Chi-square residuals of proportion postives of west GOM design based model.
Delta lognormal mincount for red snapper Chisq Residuals proportion positive


Table 10. West GOM red snapper Lo and standardized index of abundance by year for design based model.

| Year | $N$ | Lolndex | StdIndex | SE | $C V$ | LCL | $U C L$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1993 | 45 | 0.40422 | 0.61146 | 0.18653 | 0.46146 | 0.25394 | 1.47233 |
| 1994 | 43 | 0.53307 | 0.80638 | 0.20841 | 0.39096 | 0.37928 | 1.71443 |
| 1995 | 79 | 0.39582 | 0.59876 | 0.11036 | 0.2788 | 0.34643 | 1.03491 |
| 1996 | 171 | 0.39036 | 0.5905 | 0.09221 | 0.23621 | 0.37053 | 0.94105 |
| 1997 | 133 | 1.0229 | 1.54734 | 0.14198 | 0.13881 | 1.1738 | 2.03974 |
| 2002 | 97 | 0.70069 | 1.05993 | 0.13077 | 0.18663 | 0.73208 | 1.53461 |
| 2004 | 49 | 0.60728 | 0.91864 | 0.17917 | 0.29504 | 0.51547 | 1.63714 |
| 2005 | 136 | 0.7256 | 1.09762 | 0.12643 | 0.17424 | 0.77668 | 1.55118 |
| 2006 | 136 | 0.28464 | 0.43057 | 0.08142 | 0.28605 | 0.24573 | 0.75446 |
| 2007 | 158 | 0.79635 | 1.20463 | 0.13151 | 0.16514 | 0.86773 | 1.67235 |
| 2008 | 122 | 0.40457 | 0.612 | 0.09745 | 0.24087 | 0.38061 | 0.98407 |
| 2009 | 166 | 0.6521 | 0.98643 | 0.11332 | 0.17378 | 0.69863 | 1.39279 |
| 2010 | 98 | 1.34125 | 2.02892 | 0.2309 | 0.17215 | 1.44154 | 2.85564 |
| 2011 | 106 | 0.99611 | 1.50682 | 0.16627 | 0.16692 | 1.08162 | 2.09918 |

Table 11. Fit statistics (a), and type III tests (b) of the GLM on positive catches for the west GOM design based model.
a

| Fit Statistics |  |
| :--- | :--- |
| -2 Res Log Likelihood | 1360.0 |
| AIC (smaller is better) | 1362.0 |
| AICC (smaller is better) | 1362.0 |
| BIC (smaller is better) | 1366.2 |

b

| Type 3 Tests of Fixed Effects |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Num | Den |  |  |  |
| Effect | DF | DF | F Value | Pr > F |
| year | 13 | 524 | 2.28 | 0.0064 |
| depth | 1 | 524 | 5.76 | 0.0168 |

Figure 37. Observed and standardized mincounts from west GOM design based model.
Delta lognormal mincount for red snapper Observed and Standardized mincount ( $95 \%$ CI)
STDmincount


$$
\begin{array}{lll}
\text { PLOT } & \leftrightarrow \text { STDmincount } & \forall \Leftrightarrow \text { LCI } \\
& \Leftrightarrow \Leftarrow \text { UCI } & \leftrightarrow \text { obsmincount }
\end{array}
$$

Figure 38. Observed versus predicted mincounts of positive data from west GOM design based model.

Delta lognormal mincount for red snapper Diagnostic plots: 2) Obs vs Pred mincount of Posit only MINCOUNT


Figure 39. Observed versus predicted mincounts of west GOM design based model.
Delta lognormal mincount for red snapper Diagnostic plots: 3) Obs vs Pred mincount Input units


## PLOT $\because$ MINCOUNT $\quad \leftrightarrow-\diamond \diamond$ index

Figure 40. Residuals of positive mincounts for west GOM design based model.
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Figure 42. QQ plot of positive mincounts for west GOM design based model.
Delta lognormal mincount for red snapper QQplot Residuals Positive mincount rates


Table 12. GOM wide red snapper Lo and standardized index of abundance by year for model based runs.

| Survey <br> Year | Frequency | $N$ | Lo <br> Index | Scaled <br> Index | $C V$ | LCL | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.08929 | 168 | 0.07873 | 0.36992 | 0.70235 | 0.10425 | 1.31259 |
| 1994 | 0.08451 | 142 | 0.07169 | 0.33683 | 0.84613 | 0.07747 | 1.46442 |
| 1995 | 0.10266 | 263 | 0.06264 | 0.29435 | 0.67252 | 0.08675 | 0.99872 |
| 1996 | 0.15528 | 322 | 0.07598 | 0.35701 | 0.57243 | 0.12310 | 1.03537 |
| 1997 | 0.25926 | 297 | 0.19372 | 0.91024 | 0.41834 | 0.40770 | 2.03222 |
| 2001 | 0.23750 | 80 | 0.24015 | 1.12841 | 0.48343 | 0.45124 | 2.82180 |
| 2002 | 0.23596 | 267 | 0.22808 | 1.07168 | 0.38202 | 0.51225 | 2.24207 |
| 2004 | 0.22500 | 200 | 0.22032 | 1.03523 | 0.40298 | 0.47654 | 2.24893 |
| 2005 | 0.25791 | 411 | 0.26298 | 1.23565 | 0.34377 | 0.63328 | 2.41100 |
| 2006 | 0.12679 | 418 | 0.11511 | 0.54089 | 0.47013 | 0.22129 | 1.32206 |
| 2007 | 0.23313 | 489 | 0.24803 | 1.16544 | 0.34914 | 0.59144 | 2.29651 |
| 2008 | 0.16587 | 416 | 0.16659 | 0.78275 | 0.41933 | 0.34998 | 1.75064 |
| 2009 | 0.21535 | 469 | 0.25851 | 1.21468 | 0.35083 | 0.61452 | 2.40097 |
| 2010 | 0.33623 | 345 | 0.46882 | 2.20285 | 0.30357 | 1.21646 | 3.98906 |
| 2011 | 0.30769 | 455 | 0.50100 | 2.35407 | 0.28695 | 1.34119 | 4.13189 |

Figure 43. Model based GOM-wide standardized versus observed mincounts.


| PLOT | $\Leftrightarrow$ STDmincount | $\cdots-\operatorname{LCl}$ |
| ---: | :--- | :--- | :--- |
|  | -- UCI | $\cdots$ obsmincount |

Table 13. West GOM red snapper Lo and standardized index of abundance by year for model based runs.

| Survey <br> Year | Frequency | $N$ | Lo <br> Index | Scaled <br> Index | $C V$ | LCL | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.17021 | 47 | 0.32502 | 0.55434 | 0.46715 | 0.22797 | 1.34800 |
| 1994 | 0.22222 | 45 | 0.40985 | 0.69902 | 0.41710 | 0.31378 | 1.55724 |
| 1995 | 0.22892 | 83 | 0.31509 | 0.53741 | 0.29894 | 0.29936 | 0.96476 |
| 1996 | 0.23392 | 171 | 0.32834 | 0.55999 | 0.24919 | 0.34276 | 0.91490 |
| 1997 | 0.54135 | 133 | 0.88256 | 1.50526 | 0.15520 | 1.10561 | 2.04936 |
| 2001 | 0.28889 | 45 | 0.41267 | 0.70384 | 0.35269 | 0.35486 | 1.39599 |
| 2002 | 0.38144 | 97 | 0.54674 | 0.93250 | 0.22054 | 0.60306 | 1.44191 |
| 2004 | 0.30000 | 50 | 0.55731 | 0.95052 | 0.29913 | 0.52927 | 1.70702 |
| 2005 | 0.38235 | 136 | 0.66606 | 1.13601 | 0.17575 | 0.80148 | 1.61017 |
| 2006 | 0.20588 | 136 | 0.27406 | 0.46743 | 0.29992 | 0.25990 | 0.84069 |
| 2007 | 0.39241 | 158 | 0.76710 | 1.30833 | 0.16188 | 0.94846 | 1.80474 |
| 2008 | 0.28467 | 137 | 0.40156 | 0.68489 | 0.23416 | 0.43146 | 1.08718 |
| 2009 | 0.35028 | 177 | 0.64564 | 1.10117 | 0.17019 | 0.78538 | 1.54394 |
| 2010 | 0.53333 | 105 | 1.30970 | 2.23376 | 0.18087 | 1.56025 | 3.19800 |
| 2011 | 0.40367 | 109 | 0.95308 | 1.62553 | 0.16231 | 1.17741 | 2.24421 |

Figure 44. Model based west GOM standardized versus observed mincounts.


## PLOT $\Leftarrow \diamond$ STDmincount

 UCITable 14. East GOM red snapper Lo and standardized index of abundance by year for model based runs.

| Survey <br> Year | Frequency | $N$ | Lndex | Scaled <br> Index | $C V$ | $L C L$ | $U C L$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.05785 | 121 | 0.01032 | 0.14507 | 2.28759 | 0.00970 | 2.17045 |
| 1994 | 0.02062 | 97 | 0.00288 | 0.04046 | 6.26370 | 0.00087 | 1.89067 |
| 1995 | 0.04444 | 180 | 0.00302 | 0.04240 | 4.16360 | 0.00140 | 1.28472 |
| 1996 | 0.06623 | 151 | 0.00720 | 0.10113 | 2.56188 | 0.00588 | 1.73942 |
| 1997 | 0.03049 | 164 | 0.00861 | 0.12102 | 2.71622 | 0.00655 | 2.23463 |
| 2002 | 0.15294 | 170 | 0.07411 | 1.04138 | 0.80026 | 0.25499 | 4.25291 |
| 2004 | 0.20000 | 150 | 0.07826 | 1.09977 | 0.77984 | 0.27708 | 4.36508 |
| 2005 | 0.19636 | 275 | 0.10491 | 1.47423 | 0.67608 | 0.43216 | 5.02901 |
| 2006 | 0.08865 | 282 | 0.04482 | 0.62979 | 0.92859 | 0.13010 | 3.04874 |
| 2007 | 0.15710 | 331 | 0.07977 | 1.12090 | 0.72005 | 0.30775 | 4.08259 |
| 2008 | 0.10753 | 279 | 0.07523 | 1.05708 | 0.80929 | 0.25563 | 4.37123 |
| 2009 | 0.13356 | 292 | 0.10527 | 1.47920 | 0.69914 | 0.41887 | 5.22366 |
| 2010 | 0.25000 | 240 | 0.17141 | 2.40863 | 0.61045 | 0.78150 | 7.42349 |
| 2011 | 0.27746 | 346 | 0.23050 | 3.23892 | 0.56357 | 1.13300 | 9.25916 |

Figure 45. Model based east GOM standardized versus observed mincounts.


## PLOT STDmincount -- - LCl <br> UCI <br> -•• obsmincount

Figure 46. Red snapper length frequency of fish measured from video with lasers in 1995.


Figure 47. Red snapper length frequency of fish measured from video with lasers in 1996.


Figure 48. Red snapper length frequency of fish measured from video with lasers in 1997.


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Figure 54. Red snapper length frequency of fish measured from video with lasers in 2007.


Figure 55. Red snapper length frequency of fish measured with stereo cameras in 2008.


Figure 56. Red snapper length frequency distribution (weighted by minimum counts at each site) from fish measured with stereo cameras in 2008.


Figure 57. Red snapper length frequency of fish measured with stereo cameras in 2009.


Figure 58. Red snapper length frequency distribution (weighted by minimum counts at each site) from fish measured with stereo cameras in 2009.


Figure 59. Red snapper length frequency of fish measured with stereo cameras in 2010.


Figure 60. Red snapper length frequency distribution (weighted by minimum counts at each site) from fish measured with stereo cameras in 2010.


Figure 61. Red snapper length frequency of fish measured with stereo cameras in 2011.


Figure 62. Red snapper length frequency distribution (weighted by minimum counts at each site) from fish measured with stereo cameras in 2011.


Table 15. Red snapper lengths (fork lengths in mm ) measured by laser from video tapes (1995-2007) and by stereo still cameras (2008-2010).

| East Gulf |  |  |  |  |  | West Gulf |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | N | Minimum | Maximum | Mean | SE | N | Minimum | Maximum | Mean | SE |
| 1995 | 0 | - | - | - | - | 9 | 430.0 | 766.0 | 582.33 | 38.09 |
| 1996 | 0 | - | - | - | - | 110 | 214.0 | 860.0 | 449.62 | 10.64 |
| 1997 | 134 | 236.0 | 758.0 | 406.52 | 9.56 | 0 | - | - | - | - |
| 2001 | 0 | - | - | - | - | 13 | 400.0 | 725.0 | 593.15 | 27.19 |
| 2002 | 0 | - | - | - | - | 195 | 245.0 | 917.0 | 506.83 | 8.16 |
| 2004 | 1044 | 207.0 | 915.0 | 417.97 | 3.25 | 0 | - | - | - | - |
| 2005 | 259 | 146.0 | 790.0 | 476.37 | 6.69 | 191 | 200.0 | 733.0 | 446.62 | 7.59 |
| 2006 | 103 | 183.0 | 752.0 | 412.33 | 14.90 | 63 | 276.0 | 668.0 | 442.63 | 13.16 |
| 2007 | 389 | 190.0 | 874.0 | 386.77 | 5.06 | 273 | 213.0 | 868.0 | 443.60 | 8.13 |
| 2008 | 24 | 284.2 | 834.0 | 459.93 | 27.59 | 23 | 287.7 | 721.9 | 470.06 | 26.34 |
| 2009 | 27 | 275.0 | 583.0 | 370.63 | 13.48 | 120 | 253.0 | 545.0 | 375.46 | 6.34 |
| 2010 | 72 | 290.6 | 798.6 | 514.81 | 13.03 | 71 | 271.4 | 742.5 | 415.60 | 11.91 |
| 2011 | 122 | 294.69 | 865.55 | 512.07 | 9.64 | 45 | 205.69 | 735.59 | 468.08 | 31.34 |

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Figure 51. Red snapper length frequency of fish measured from video with lasers in 2004.
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