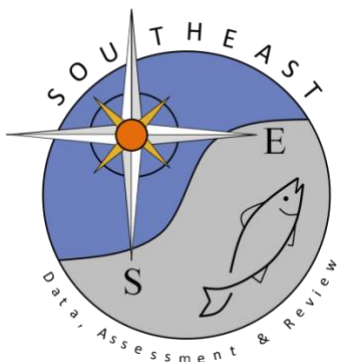


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Changes in Reef Fish Community Structure Following the Deepwater Horizon Oil Spill

Justin P. Lewis^{1*}, Joseph H. Tarnecki¹, Steven B. Garner¹, David D. Chagaris^{1,2} & William F. Patterson III¹

Large-scale anthropogenic disturbances can have direct and indirect effects on marine communities, with direct effects often taking the form of widespread injury or mortality and indirect effects manifesting as changes in food web structure. Here, we report a time series that captures both direct and indirect effects of the *Deepwater Horizon* Oil Spill (DWH) on northern Gulf of Mexico (nGoM) reef fish communities. We observed significant changes in community structure immediately following the DWH, with a 38% decline in species richness and 26% decline in Shannon-Weiner diversity. Initial shifts were driven by widespread declines across a range of trophic guilds, with subsequent recovery unevenly distributed among guilds and taxa. For example, densities of small demersal invertivores, small demersal browsers, generalist carnivores, and piscivores remained persistently low with little indication of recovery seven years after the DWH. Initial declines among these guilds occurred prior to the arrival of the now-widespread, invasive lionfish (*Pterois* spp.), but their lack of recovery suggests lionfish predation may be affecting recovery. Factors affecting persistently low densities of generalist carnivores and piscivores are not well understood but warrant further study given the myriad ecosystem services provided by nGoM reef fishes.

The nature, frequency, and intensity of disturbance are important drivers of community structure^{1–3}, and it is well established that evolutionary history⁴, historical disturbance regimes^{2,4}, and the prior state of a community^{2,5,6} affect its response⁷. Although natural disturbances can be important for maintaining diverse, resilient species assemblages⁸, research focused on the impacts of chronic anthropogenic stressors on biodiversity has revealed that even specious communities, presumed to be resilient, can respond unpredictably to natural and anthropogenic disturbances^{9,10}. This is particularly true in marine systems which have experienced impacts from human activities for centuries and are severely degraded as a result^{11,12}. Numerous examples exist of long-term community shifts from estuarine¹³, coral reef^{14,15}, and continental shelf^{16,17} systems, and it is not uncommon for communities to remain unaffected by localized or moderate disturbances⁵, only to exhibit a non-linear response following a series of disturbances¹⁸ or a single event of sufficient scale or intensity¹⁹.

The 2010 *Deepwater Horizon* Oil Spill (DWH) was the epitome of a large-scale, anthropogenic disturbance capable of producing substantial community-level impacts. Over an 87-day period, approximately 4.9 million gallons of oil²⁰ was released into northern Gulf of Mexico (nGoM) at a depth of ~1,500 m producing a surface slick of ~40,000 km² at its maximum extent²¹. Between 4 and 14% of the total discharge was transported to the benthos by contaminated marine snow^{21–24}, thus exposing numerous pelagic and benthic communities to toxic polycyclic aromatic hydrocarbons (PAHs)²⁵ as well as emulsifying dispersants²⁶. Both lethal and sublethal effects (e.g., compromised immune²⁷ and endocrine function²⁸, developmental abnormalities²⁹, reduced growth³⁰, and impaired olfaction³¹ and predator avoidance^{32,33}) of oil exposure have been well documented for numerous taxa^{34–37}, and negative effects at the organismal level had the clear potential to elicit effects on community structure through bottom-up³⁸ or top-down mechanisms^{39,40}.

Much of the effort to document the community-level responses to the DWH was focused on monitoring coastal habitats that provide critical nurseries for several marine taxa⁴¹, are widely studied by community ecologists⁴², and whose proximity favored the rapid collection of critical baseline data⁴³. Despite extensive shoreline oiling⁴⁴, impacts were mostly relegated to heavily oiled, coastal sites in Louisiana where significant vegetation loss occurred along the marsh edge⁴⁵. Inshore communities, particularly nekton and fish assemblages, in areas of

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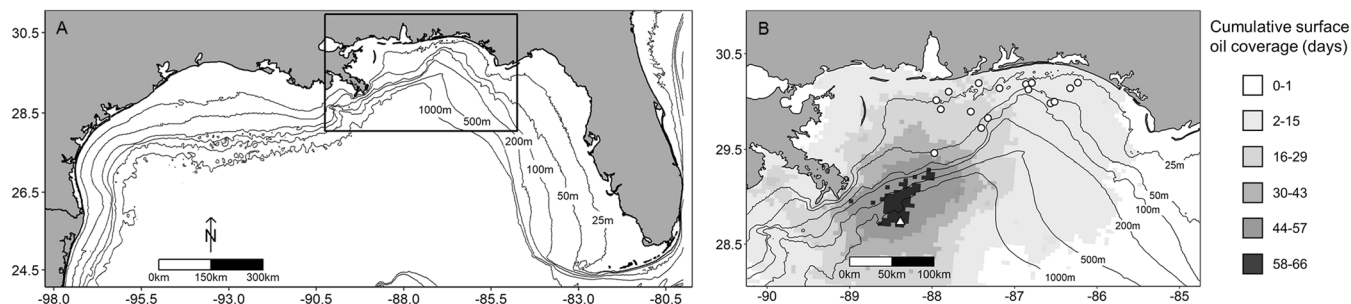


Figure 1. Map of the northern Gulf of Mexico (A) and sampling region (B). Panel B shows the location of the Deepwater Horizon blowout at the Mississippi Canyon-252 wellhead (triangle), and the natural reefs surveyed from 2009 to 2010 (circles). The shaded area represents the cumulative surface oil coverage in days. Maps were produced in R version 3.5.1⁷⁷.

limited exposure showed little indication of negative effects higher than the organismal level^{46–49}. Multiple assessments of community-level effects on the nGoM continental shelf also indicated no impacts or a resilient recovery. For example, changes in plankton communities were relatively brief and isolated to the active spill period^{50–52}; post-DWH meiofauna diversity and abundance were comparable to pre-spill estimates⁵³; and macroinfauna abundance and diversity showed no signs of impacts⁵⁴.

The lack of discernable DWH effects on community structure have been documented repeatedly but mostly for communities dominated by short-lived taxa^{45,55,56} more likely to be resilient to such a disturbance⁵⁷. Considerably less is known about the impact to and response of neritic fish communities, including reef fish assemblages in the nGoM. These fish communities are diverse^{58,59}, including numerous short-lived, small demersal species, as well as long-lived, fisheries species likely to be resistant to additional sources of mortality^{60,61}. Key to the persistence of many fisheries species is the maintenance of multiple year classes of relatively old, mature individuals (i.e., the storage effect⁶⁰) that experience low natural mortality, which minimizes biomass loss between periodically strong year classes⁶¹. Thus, compensatory population increases following DWH-induced mortality would likely be slow.

Resistance is also conferred through their high vagility and generalist diets which would allow individuals to disperse to avoid acute mortality⁶² while exploiting locally abundant prey^{63,64}. However, evidence of sub-lethal PAH exposure suggests negative effects reached well beyond the footprint of the surface slick^{37,65}, and recent ecosystem simulations suggest mortality resulting from acute resource limitation was perhaps more severe and widespread than exposure mortality⁶⁶. Indeed, changes in feeding ecology, trophic position, and condition of red snapper (*Lutjanus campechanus*) were documented following the DWH^{63,67} and provide empirical support for a shift in resource availability with possible negative effects on population productivity⁶⁷. Similar shifts in trophic position and pathways were observed in other reef fish species concomitant with petrocarbon cycling through the nGoM food web^{68,69}. However, to date few data have been presented linking declines in reef fish abundance or community shifts that followed either direct effects of the DWH or indirect effects via food web impacts⁷⁰.

The lack of information regarding community-level responses following the DWH may stem, in part, from the lack of pre-spill baseline data for many taxa and communities. Here, we analyze a time-series of nGoM reef fish community data that span an eight-year period starting before the DWH. We test whether significant changes occurred in reef fish community structure following the DWH, and if so whether changes occurred evenly across trophic guilds or were more concentrated within specific guilds. Results are presented in the context of acute, direct and chronic, indirect effects of the DWH, as well as factors that may affect the resiliency of reef fish communities.

Methods

Site description and data collection. Fish communities were surveyed at 16 natural reefs (Fig. 1) in the nGoM with a VideoRay Pro4 mini remotely operated vehicle (ROV) during 2009–2017. Reefs were randomly selected from a series of sites surveyed in 2008–2009⁷¹, encompassed a depth range of 17–72 m, and were distributed over an 8,000 km² area of the continental shelf. Sites were representative of the morphologically variable hard bottom habitat in the region and included low relief ledges, rocky ridges, rock rubble mounds, and flat limestone block reefs⁷². The epibenthic communities are dominated by coralline algae, soft corals (e.g., black corals, gorgonians, and octocorals), and sponges with limited coverage by azooxanthellate, ahermatypic corals⁷². Although impacts to epibenthic species were observed west of our sampling region⁷³, sedimentation of contaminated marine snow was patchy²⁴, and we observed no signs of habitat degradation (e.g., oiling or injured/deceased coral colonies). Thus, monitoring of the epibenthic community was not undertaken.

In total, 250 ROV surveys were completed between 2009 and 2017. The pre-DWH portion of our time series consisted of 26 surveys across 11 sites and occurred from the summer of 2009 to the spring of 2010. All 2010 surveys classified as pre-DWH occurred prior to surface oil entering our sampling region; surveys grouped in the 2010 time bin were sampled in November. During 2011–2013, sites were typically surveyed in the spring, summer, and fall, but later in the time series funding constraints limited sampling to summer months. Each ROV survey consisted of 3 to 4 orthogonal, 25-m transects either 1 or 2 m off the bottom, depending on visibility^{70,71}.

The ROV was equipped with a 570-line color camera with wide-angle (116°) lens for real-time viewing, twin forward projecting LEDs (3,600 lumens), and rated to a depth of 170 m. An additional forward-facing, high definition (1080p at 60–120 fps) camera (GoPro Hero 2, 3, or 4) was mounted at a 45° angle to the ROV's float block above the internal camera to record high definition video of the reef fish community.

Video samples were analyzed on a high-resolution monitor; observed fishes were identified to the lowest taxonomic level possible and enumerated. Counts were summed across transects to estimate the total abundance for each site. For active, schooling species prone to double counting (e.g., scads, herrings, etc.), the total abundance across all transects was superseded by an estimate of the minimum number of individuals within a school obtained while maneuvering between transect locations. Species abundances were converted to densities by dividing by the total area sampled following the methods of Patterson *et al.*⁷¹. Results presented below only include those taxa identified to the level of species with the exception of purple reeffish (*Chromis scotti*) and dusky damselfish (*Stegastes fuscus*), which are difficult to distinguish on video footage and therefore combined into a single group, damselfish.

Community analysis. Permutational multivariate analysis of variance (PERMANOVA) models were computed in PRIMER (v7) to test for temporal changes in reef fish community structure following the DWH. To reduce the influence of abundant species, taxa-specific densities were $\log(x + 1)$ transformed. A resemblance matrix was then computed based on the Bray-Curtis dissimilarity, with the inclusion of a dummy species at a density of 1. The dummy species was included because Bray-Curtis can behave erratically if few species are shared between sites⁷⁴, which is important to consider when evaluating environmental impacts on community structure. The PERMANOVA model had a three-factor hierarchical design with site nested within month nested within year. The nested factors, site and month, were treated as random factors while year was treated as a fixed factor. Our model also included two covariates, depth and longitude, which were z-score transformed. The reasons for this approach were: (1) each covariate represents a gradient along which reef fish community structure naturally varies^{58,75}; (2) cluster analysis of pre-DWH community structure did not identify groups that might justify the use of discrete categorical factors to evaluate depth or longitude effects; (3) the use of covariates as opposed to fixed factors permitted the inclusion of the entire data set; and, (4) at this scale there is not a clear relationship between straight line distance from the well head and impacts. Changes in community structure were also evaluated using common community indices of species richness (*S*), diversity (Shannon-Weiner *H'*), and evenness (Pielou's *J'*). All indices were calculated using the *vegan* package⁷⁶ in R⁷⁷. Temporal changes were evaluated with linear mixed effects models (LMMs) using the *lme4* package⁷⁸ followed by Dunnett's multiple comparisons using the *multcomp* package⁷⁹.

Trophic guild and species-specific trends. We evaluated temporal changes in density for nine trophic guilds: herbivores, small demersal browsers, large demersal browsers, small demersal invertivores, large demersal invertivores, generalist carnivores, piscivores, reef planktivores, and pelagic planktivores, with small versus large indicating species generally smaller versus larger than 200 mm total length. Species were assigned to guilds based on dietary data, both from the literature (see Appendix A) and recent analyses, and densities were summed by guild for each ROV sample. In the case of small demersal invertivores, we excluded tomtate (*Haemulon flavolineatum*) from guild-level estimates because densities of this schooling grunt (Haemulidae) were highly variable and often an order of magnitude larger than other guild members. Their inclusion obscured the more general, guild-level pattern (Fig. S1). The species-specific analyses included taxa ($n = 52$) whose relative frequency of occurrence was $>5\%$ before or after DWH and for which sufficient data were available for model convergence.

Temporal trends in trophic guilds and species were assessed by computing standardized density indices with generalized linear mixed effects models (GLMM) following the delta approach^{80,81}. This approach consists of two models, one to model the probability of observing zero individuals (hereafter, presence/absence) and a second to model the density given a guild or species was observed⁸¹. The product of the two sub-models was then used as the standardized density index for each guild or taxon. GLMMs included year as a factor and the repeated measures design was specified by including a random intercept parameter for each site. Longitude and depth were also included as covariates following z-score transformation. Least-squares means were calculated for each sub-model as the annual average from a reference grid of predictions across factor levels (i.e., years). Monte Carlo simulations were computed to estimate an annual density index and confidence intervals following the methods in Chagaris *et al.*⁸². Briefly, the product of the least-squares mean standard error and 10,000 random normal deviates $X \sim N(\mu = 0, \sigma = 1)$ was added to the least-squares mean estimate of annual density. Error deviates of the log-normal model were adjusted when the log-normal and binomial least-squares mean were correlated (Pearson's correlation p -value ≤ 0.05). Values of each Monte Carlo simulation were then back-transformed into their original measurement units to obtain a distribution of density values.

The results from binomial and log-normal models were also evaluated separately to infer whether temporal differences resulted from a significant change in presence/absence or non-zero abundance. Multiple comparisons performed using Dunnett's method, as described above. For guilds and species observed prior to the DWH, comparisons were made between the pre- and post-DWH time periods with non-zero density estimates. For species only observed during the post-DWH time period, comparisons were made between the first and subsequent years with non-zero density estimates.

Results

Species composition. Our ROV dataset included 138 species from 43 families. The highest densities were observed for grunts and snappers (Lutjanidae) reflecting the fact either tomtate or vermilion snapper (*Rhomboplites aurorubens*) was the most abundant species in a given year (standardized density range 27–136 and 26–195 individuals 1000 km², respectively) (Supplemental Table S1). Approximately 43% percent of reef fish species were distributed among five other families: Serranidae (15.2%), Carangidae (7.2%), Sciaenidae (5.8%), Sparidae (5.8%), and Pomacentridae (5.1%).

Source	df	SS	MS	Pseudo-F	P-value
Depth	1	8.15×10^4	8.15×10^4	41.69	<0.01*
Longitude	1	3.13×10^4	3.13×10^4	16.09	<0.01*
Year	8	4.28×10^4	5.35×10^4	1.76	<0.01*
Month/Year	26	5.80×10^4	2.23×10^3	1.20	0.02*
Site/Month/Year	206	3.77×10^5	1.83×10^3	1.20	0.17*
Residuals	7	1.07×10^4	1.53×10^3		
Total	249	6.01×10^5			

Table 1. Permutational multivariate analysis of variance results based on Bray-Curtis dissimilarity. Significant differences ($\alpha = 0.05$) denoted with an asterisk (*).

	Pre-DWH	2010	2011	2012	2013	2014	2015	2016	2017
Pre-DWH	63.0	—	—	—	—	—	—	—	—
2010	65.8*	61.6	—	—	—	—	—	—	—
2011	66.6*	65.8	67.5	—	—	—	—	—	—
2012	68.1	67.9	69.2	70.5	—	—	—	—	—
2013	67.1*	67.9*	68.7*	69.5	67.0	—	—	—	—
2014	69.5*	69.5*	70.0	70.9	68.8	71.2	—	—	—
2015	65.9*	69.1*	70.1*	70.3*	65.8*	68.5	60.8	—	—
2016	67.6*	65.8*	69.1*	69.3*	66.6*	67.9	63.3	62.0	—
2017	66.0*	65.3*	68.2*	68.1	65.3*	67.1	61.9*	60.5	60.7

Table 2. Post-hoc pairwise comparisons of community structure among years based on Bray-Curtis dissimilarity. Values along and below the diagonal represent within and between year dissimilarities, respectively. Significant differences ($\alpha = 0.05$) denoted with an asterisk (*).

Community analysis. PERMANOVA results indicated community structure significantly differed among years (Table 1). Both covariates and the random effect of month within year were also statistically significant, while the random effect of site within month within year was not significant. Of the eight pairwise comparisons between pre- and post-DWH periods, significant differences in community structure were observed in 2010–2011 and 2013–2017 (Table 2). Pairwise comparisons from the post-DWH portion of the time series (i.e., 2010–2017) also indicated significant interannual differences in community structure. However, significant differences during the post-DWH period were more common for comparisons separated by one or more years.

All three community indices (S , H' , and J'), as well as total fish density, declined following the DWH (Fig. 2) with a significant decline observed for S (Table S2). Species richness showed an upward trend in 2011–2015, going from 14.1 to 21.3 species per site and remained comparable to our pre-DWH baseline of 16.4 in 2016 and 2017. However, H' and J' continued to decline and were significantly lower than pre-DWH estimates in 2012, 2013, and 2016 (Table S2). The 2010 decline in total fish density, though not statistically significant, was substantial and represented a decline of ~62%. From 2011 onward, fish density showed a slight positive trend.

Trophic guild and species-specific trends. Following the DWH, densities of all eight trophic guilds observed prior to the oil spill declined (Fig. 3). The magnitude of these declines ranged from 35 to 96% and four of the eight trophic guilds reached their lowest densities in 2010. Although we did not observe significant changes in guild presence/absence (Supplemental Table S3), the initial declines in herbivore, small and large demersal browser, small and large demersal invertivore, generalist carnivore, and piscivore densities were associated with significantly lower abundances when present (Supplemental Table S4). The 52 species for which species-specific trends were evaluated reflect this general pattern (Supplemental Table S1). Forty-six species were observed prior to the DWH and 43 declined between our pre-DWH baseline and 2010. For 29 species, these initial declines reflected either a complete absence, significant change in presence/absence, (Supplemental Table S5), or significantly lower densities when present (Supplemental Table S6).

Guilds comprised of small-bodied species that forage on benthic prey showed the largest declines immediately following the DWH. Densities of herbivores, small demersal browsers, and small demersal invertivores declined by 96%, 87%, and 82%, and remained persistently low through much of the time series (Fig. 3). The decline in herbivore density almost entirely reflected doctorfish (*Acanthurus chirurgus*) abundance, while the decline in small demersal browsers resulted from lower densities of the cocoa damselfish (*Stegastes variabilis*) and seaweed blenny (*Parablennius marmoratus*) (Fig. 4). The trend displayed by small demersal invertivores was driven by species like the slippery dick (*Halichoeres bivittatus*) and cubbyu (*Paraques umbrosus*) and differed from that of the tomtate, which has a looser association with the reef structure (Fig. 4).

Declines among large-bodied species reliant on benthic production were also evident several years following the DWH, although these declines were less severe (Fig. 3). The blue angelfish (*Holocanthus bermudensis*) was the most abundant large demersal browser and drove guild-level trends (Fig. 4). The density of large demersal invertivores reflected red porgy (*Pagrus pagrus*) and gray triggerfish (*Balistes capricus*) abundances. However, each

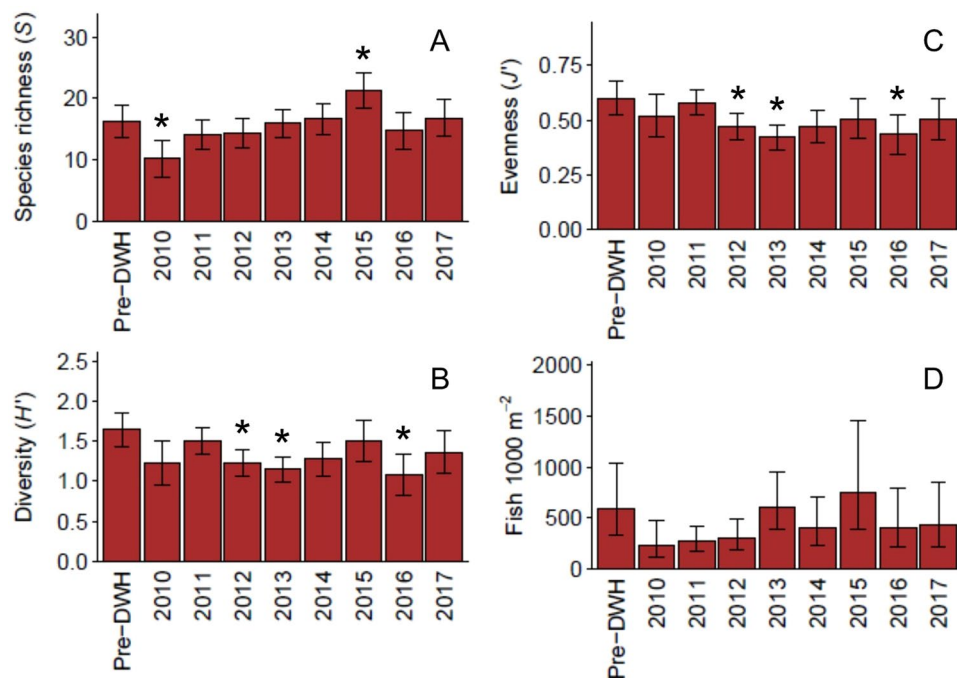


Figure 2. Standardized community indices or total fish density ($\pm 95\%$ CIs) estimated from generalized linear models with corresponding 95% CIs generated through Monte Carlo simulations. An asterisk (*) denotes a significant pairwise difference between pre- and post-DWH time points (Supplemental Table S2).

species clearly displayed disparate patterns. The initial decline in red porgy was followed by an increase in 2011 and subsequent decline (Fig. 4). Conversely, gray triggerfish density was more variable and temporal changes were not associated with a significant difference in presence/absence or density when present.

Guilds representing higher trophic level consumers showed trends more similar to the small-bodied demersal guilds. Densities of generalist carnivores and piscivores declined by 64% and 73% in 2010 and densities remained persistently low thereafter (Fig. 3). For both guilds, these trends were driven by large-bodied, fisheries species (Fig. 5). Red snapper, gray snapper (*L. griseus*), and red grouper (*Epinephelus morio*) declined by 69%, 85%, and 70% following the DWH, and low densities persisted through 2017. The three most abundant piscivores, scamp (*Mycteroperca phenax*), gag (*M. microlepis*), and sandbar shark (*Carcharhinus plumbeus*), all declined after the spill and both scamp and gag remained at densities below pre-DWH baseline values. Densities of smaller bodied generalist carnivores, [e.g., bank seabass (*Centropristis ocyurus*) and belted sandfish (*Serranus subligarius*) (Fig. 5)] either displayed no change or failed to recover following the DWH. The one exception was the invasive lionfish (*Pterois* spp.) which was first observed in 2011 and rapidly increased through 2017 (Fig. 5).

Unlike guilds that rely on benthic forage, reef planktivore densities remained unchanged following the DWH (Fig. 3). Similarly, the downward trend through 2012 and subsequent increase were not associated with a significant difference in presence/absence or density when present (Supplemental Tables S3 and S4). The vermilion snapper was the most abundant reef planktivore and displayed a similar temporal pattern (Fig. 6). Of the nine other reef planktivores, four displayed a significant decline in presence/absence or density when present (Supplemental Table S3). However, even species that experienced declines of >90% in 2010 generally increased to pre-DWH densities [e.g., yellowtail reeffish (*Chromis enchrysur*) and damselfish (Fig. 6)]. Pelagic planktivores were not observed during our pre-DWH surveys and infrequently observed thereafter. No pelagic planktivores met our selection criteria for species-specific analysis (see Methods).

Discussion

Our results indicate reef fish communities exhibited clear signs of negative impacts following the DWH with significant shifts in community structure and declines in species richness, diversity, evenness, and total fish density. This change in community structure was unique in that it that concomitant declines were observed for all eight trophic guilds. At no other point were similar, synchronous declines present nor were significant pairwise differences in community structure evident between successive years. The species composition of the more abundant fishes was similar before and after the spill suggesting declines in species richness resulted from an absence of rare species and changes in community structure, species diversity, evenness, and total fish density resulted from shifts in relative abundances. Declines in species richness did not persist and has remained similar to pre-DWH richness. However, lower estimates of diversity and evenness were evident several years post-spill reflecting lower densities of small demersal browsers, small demersal invertivores, generalist carnivores, and piscivores. These effects are similar to those observed among deep-sea benthic communities where declines in megafauna⁸³, macrofauna⁸⁴, meiofauna⁸⁵, and foraminifera⁸⁶ abundance and diversity followed the DWH. Although the most severe

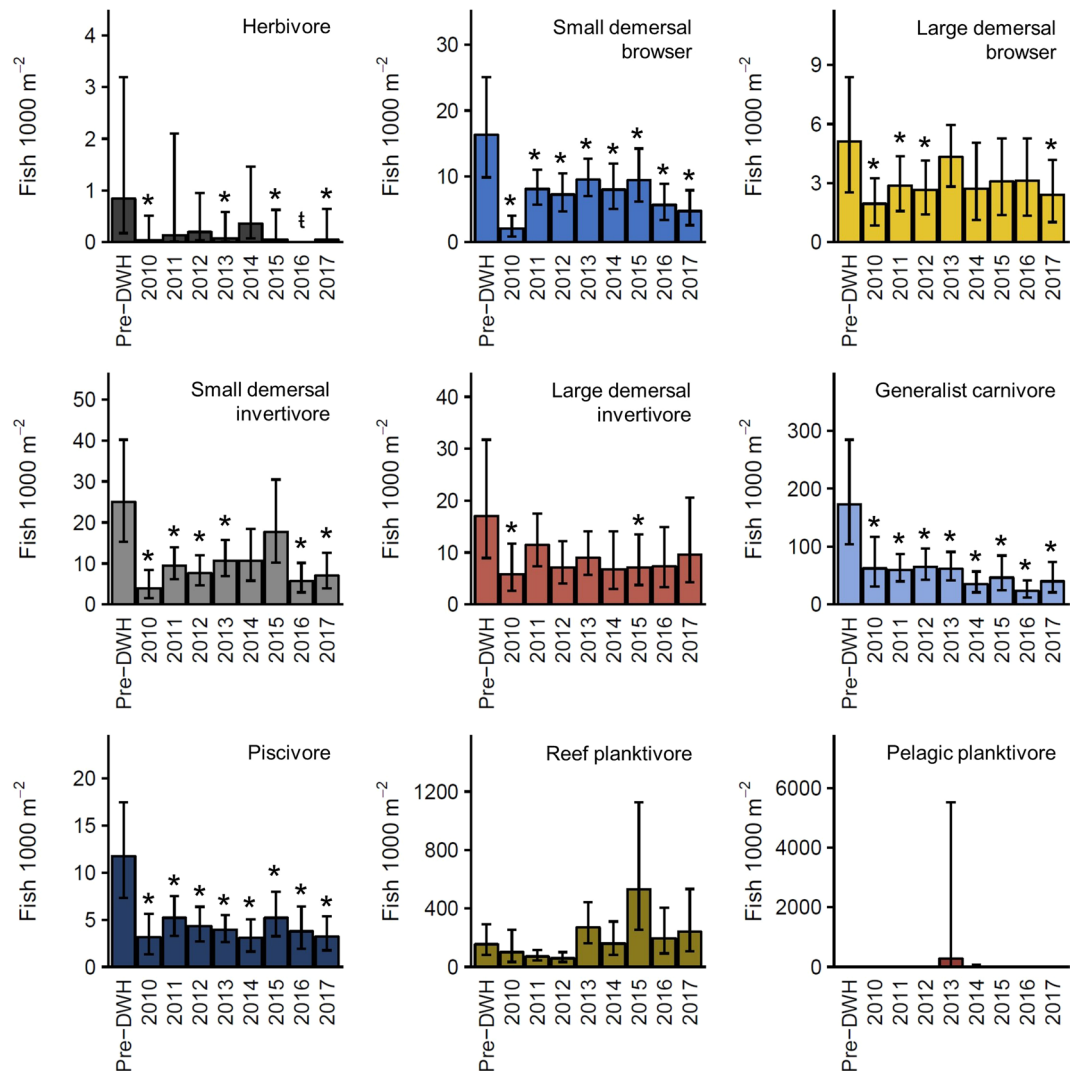


Figure 3. Standardized trophic guild densities ($\pm 95\%$ CIs) estimated from generalized linear mixed effects models with corresponding 95% CIs generated through Monte Carlo simulations. A unique color was assigned to each trophic guild and this color scheme is used in subsequent figures. A stroked t (t) indicates a significant pairwise difference in presence/absence between pre-DWH and post-DWH time bins (Supplemental Table S3) or complete absence. An asterisk (*) denotes a significant difference in density when present (Supplemental Table S4). Trophic guild names are provided in the top right corner of each panel.

impacts to deep-sea benthic communities were in immediate vicinity of the wellhead⁸⁴, the negative responses observed among decapod crustacean communities on deep bank formations⁸⁴ and fish communities on artificial reef⁷⁰ 100 s of km away suggests acute exposure at distant sites was of sufficient intensity to produce observable shifts.

We report declines across a range of reef fish taxa regardless of vagility, trophic position, or diet, which seem improbable without DWH-induced mortality. However, a primary challenge is identifying the extent to which declines resulted from mortality or emigration. Small fishes that live in close proximity to the reef matrix (e.g., small demersal browsers and invertivores) experienced the largest initial declines, which likely reflects a high incident of acute mortality. These species can have limited ($<10 \text{ m}^2$) home ranges⁸⁷, are heavily reliant on local resource pools, and would incur the highest cost associated with emigration. These traits not only increase the probability of exposure-related mortality but also mortality associated with resource limitation⁶⁶. Impacts on pelagic production^{88,89}, increased trophic position^{68,69}, and greater reliance on benthic resources⁶³ suggest small demersal reef fishes experienced increased resource competition and higher predation immediately following the DWH. The effects of resource limitation were possibly exacerbated by sub-lethal exposure³⁷ which can result in physiological stress leading to impaired predator avoidance and foraging ability^{32,33,90}. This inference is further supported by the fact that initial declines among large-bodied reef fishes (e.g., large demersal invertivores, some generalist carnivores, and piscivores) were less severe. These species are quite mobile, exhibit varying degrees of site fidelity on natural reefs⁹¹, and their movements can be affected by large-scale disturbances^{62,92}. Additionally, a large area (maximum = $290,000 \text{ km}^2$) of the continental shelf was closed to harvest for a few months during the

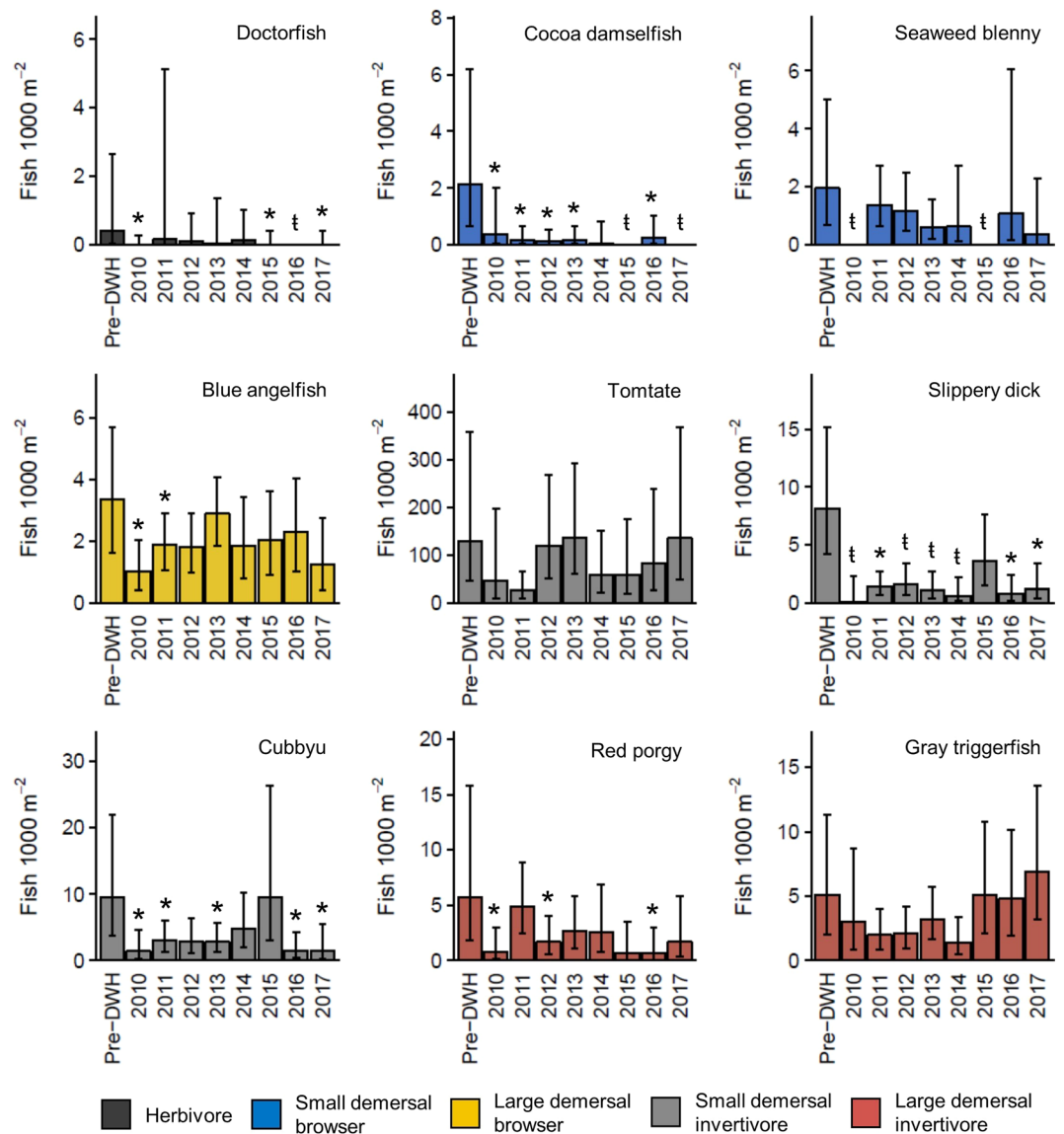


Figure 4. Standardized densities of representative species from the herbivore, small demersal browser, large demersal browser, small demersal invertivore, and large demersal invertivore, trophic guilds. Estimates were derived from generalized linear mixed effects models using the delta approach. Corresponding 95% CIs were generated through Monte Carlo simulations. A stroked t (t) indicates a significant pairwise difference in presence/absence between pre-DWH and post-DWH time bins (Supplemental Table S5) or complete absence. An asterisk (*) denotes a significant difference in density when present (Supplemental Table S6). Species names are provided in upper right corner of each panel.

spill, perhaps reducing fishing-related mortality, albeit temporarily. Emigration following the DWH also may have occurred if the cost of exposure or resource limitation outweighed the benefits of staying^{93,94}. For some individuals, exposure may have been sufficient to elicit movement from affected areas. For others, the response was perhaps a shift in foraging behavior^{63,68}, followed by emigration as resources became more scarce. Individuals present after the DWH probably consisted of few residents that survived acute exposure and resource limitation along with new immigrants seeking more favorable conditions⁹³. The fact that guilds comprised of mobile reef species (i.e., generalist carnivores and piscivores) showed little indication of recovery suggests a large number of individuals were either permanently displaced, perished (either from starvation or exposure⁶⁶), or basal resource pools remained insufficient to support pre-DWH densities.

The effects of natural disturbances within the region (e.g., hurricanes and hypoxia) are often insufficient to produce an observable change in large, mobile reef fish abundance^{95,96}. Thus, it appears the acute impact of the DWH was more severe compared to large-scale natural disturbances typical of our study area. Initial community-wide declines are, however, notably similar to the effects of harmful algal blooms that seasonally occur along the West Florida Shelf (WFS)^{97–99}. The 1971 red tide event is the most well-documented case of the impacts on reef fish communities and subsequent recovery. Exposure resulted in the near extirpation of reef

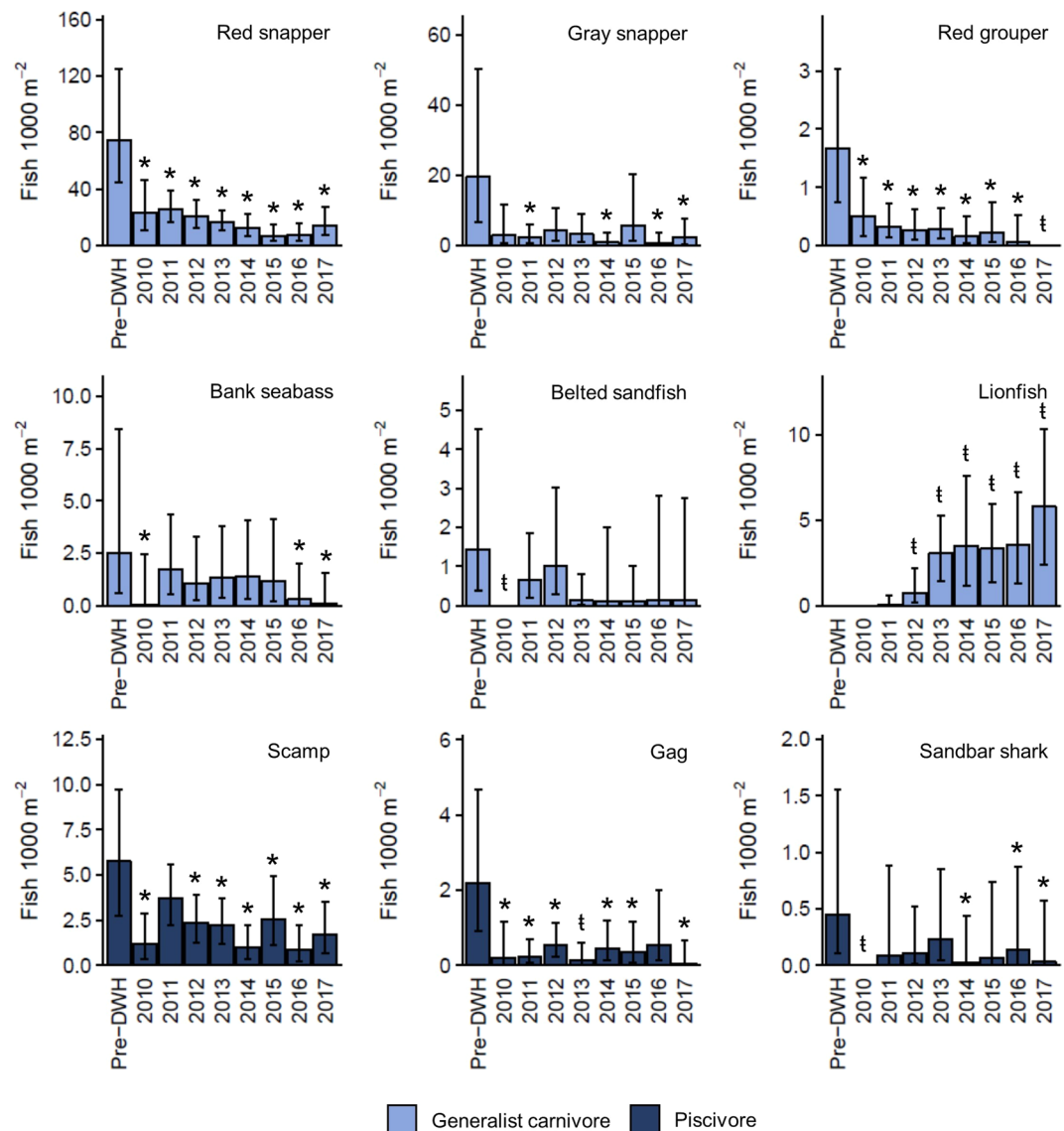


Figure 5. Standardized densities of representative species from the generalist carnivore and piscivore trophic guilds. Density estimates were derived from generalized linear mixed effects models using the delta approach. Corresponding 95% CIs were generated through Monte Carlo simulations. A stroked *t* (*t*) indicates a significant pairwise difference in presence/absence (Supplemental Table S5) or complete absence. An asterisk (*) denotes a significant difference in density when present (Supplemental Table S6). Species names are provided in upper right corner of each panel.

fishes across a >1,500 km² stretch of the WFS⁹⁷. Conspicuous¹⁰⁰ and indiscriminate reef fish mortality produced clear declines in species richness, but the most pronounced effects were changes in species relative abundance⁹⁷. Recovery followed a predictable pattern of succession initiated by the arrival of small demersal species that recruit directly to reefs, followed by a peak in abundance among early pioneers, and subsequent decline in abundance superior competitors arrived; increases among large mobile fishes that do not recruit directly to reefs occurred later⁹⁷. Full recovery took several years, but the community that developed was nearly identical to that observed prior to the 1971 red tide. The same general pattern of community level impacts, succession, and recovery were also observed on WFS artificial reefs following a 2005 red tide event⁹⁸, corroborating observations by Smith⁹⁷ and others that documented successional patterns among GoM and Caribbean reef fish communities^{101,102}. Although the rate of succession can vary among sites^{103,104}, after seven years of post-DWH monitoring no such pattern emerged despite similarities between community members in the aforementioned studies.

A clear difference between the DWH and natural disturbances was the potential for chronic PAH exposure, which can have long-term, higher-order impacts even at sub-lethal levels¹⁰⁵. While sedimentation of contaminated marine snow was the primary vector transporting oil and dispersants to the benthos, this phenomenon was patchy and mostly concentrated off the continental shelf and west of our sampling area^{106,107}. As a result, sediment concentrations of total petroleum and polycyclic aromatic hydrocarbons (PAHs) were typical of background

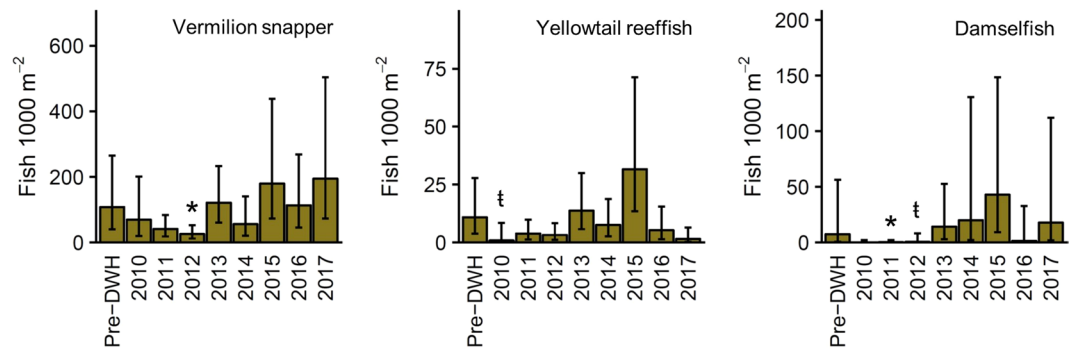


Figure 6. Standardized densities of reef planktivores estimated from generalized linear mixed effects models using the delta approach. Corresponding 95% CIs were generated through Monte Carlo simulations. A stroked t (t) indicates a significant pairwise difference in presence/absence between pre-DWH and post-DWH time bins (Table S5) or complete absence. An asterisk (*) denotes a significant difference density when present (Table S6). Species names are provided in upper right corner of each panel.

levels within our study area in August 2010⁵⁴. Nonetheless, examination of reef fish tissue samples indicated elevated levels of PAH concentrations in liver and PAH metabolites in bile persisted for years following the DWH^{37,65,108}, thus providing evidence of continual reef fish exposure to toxic petroleum compounds for some time after the spill despite uncertainties about the mechanism of exposure.

The fact that reef fish density remained low for a number of years following the DWH suggests indirect food web effects have played an important role as well⁶⁶. Consumption of swarming zooplankton by red snapper declined markedly following the DWH, a possible reflection of an initial reduction in pelagic production⁶⁶. The reduction in zooplankton was balanced by increased foraging on demersal fishes and invertebrates⁶³. This increased reliance on benthic resources continued over weeks and months following DWH, as indicated by declines in $\delta^{34}\text{S}$ ^{63,68}. The associated enrichment of $\delta^{15}\text{N}$ persisted for several years following the DWH^{63,68}, and similar long-term shifts in $\delta^{15}\text{N}$ were observed for gray triggerfish, tomtate, red porgy, and vermilion snapper concurrent with declines in $\delta^{13}\text{C}$, indicative of petrocarbon cycling through the food web^{68,109}. The 2010 shifts in trophic position and community-wide declines that extended into 2011 also pre-date the arrival and rapid expansion of the invasive lionfish and are in general agreement with previous reports that community-wide impacts following the DWH occurred prior to the nGoM lionfish invasion⁷⁰.

Although the role of food web effects resulting from a shift in resource availability is apparent, the success of the invasive lionfish may also be an important factor suppressing community recovery. Numerous studies have documented community-level effects of lionfish^{110,111} via predation^{112,113} and competition^{114,115}. Impacts are typically most evident among small demersal reef fishes that recruit directly to reef habitats⁷⁰ and vulnerable to predation both as adults^{112,116} and newly-settled recruits¹¹³. The stunted recovery of small demersal invertivores and browsers, despite their capacity for rapid recolonization following mass mortality events⁹⁷ and the continued declines among native predators, provides a clear indication that the success of the invasive lionfish is affecting the response among native, small demersal fishes. How lionfish may be affecting the recovery of fisheries species has not been evaluated, but competition between lionfish and native predators (e.g., groupers, snappers, and jacks) may affect population productivity¹¹⁷. Taxonomic resolution of dietary data for fisheries species is often poor and presents challenges when attempting to capture competitive interactions in systems with diverse species assemblages¹¹⁷. Diet information for the majority of small demersal reef fish is also lacking, and the potential food web effects emanating from low densities of small demersal species remains unknown.

The changes in community structure, particularly the persistently low densities among certain groups, provides a clear indication of lasting, community-wide impacts. The available evidence suggests initial declines in 2010 likely reflected both mortality and emigration resulting from exposure and resource limitation. Mortality due to direct or indirect effects of the spill likely drove initial declines of small demersal species, while large-bodied consumers were more likely to be permanently displaced or to suffer delayed mortality. Community-wide declines into 2011 were indicative of protracted resource limitation. However, the lack of recovery in small demersal reef fishes from 2012 onward may, to some extent, reflect top-down pressure from lionfish. How community shifts have altered the flow of energy to higher trophic level fisheries species or impacted system resilience remains uncertain.

A clear challenge moving forward is identifying the underlying mechanisms driving these patterns, estimating the relative impacts of individual stressors (e.g., exposure, resource limitation, lionfish invasion, fisheries harvest, and food web effects), and developing management strategies to facilitate recovery. Efforts along these lines appear particularly relevant considering the dramatic declines observed for higher trophic position consumers that not only serve important ecological roles¹¹⁸ but also provide numerous economic¹¹⁹ and cultural¹²⁰ benefits. Many of the ecosystems services we ascribe to reef fish assemblages in the nGoM are inextricably linked to fisheries harvest^{119,120}, and fishery-dependent data and assessment of fishery stocks should enable tracking resilience among those species. However, continued funding for fishery-independent surveys, such as the ROV work that forms that basis of the analyses presented herein, is critical to assess the long-term effects of the DWH, lionfish invasion, and the potential for resiliency in the nGoM ecosystem.

Data availability

Data are available through the Gulf of Mexico Research Initiative Information & Data Cooperative (GRIIDC) at <https://data.gulfresearchinitiative.org> under DOI 10.7266/N72J68SF and 10.7266/n7-n4j3-0a26. The remaining portion of the time series can be obtained from the corresponding author upon request.

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References

- Hall, A. R. *et al.* Diversity–disturbance relationships: frequency and intensity interact. *Biol. Lett.* **8**, 768–771 (2012).
- Davies, K. W., Svejcar, T. J. & Bates, J. D. Interaction of historical and nonhistorical disturbances maintains native plant communities. *Ecol. Appl.* **19**, 1536–1545 (2009).
- Svensson, J. R., Lindegarth, M. & Pavia, H. Equal rates of disturbance cause different patterns of diversity. *Ecology* **90**, 496–505 (2009).
- Waples, R. S., Beechie, T. & Pess, G. R. Evolutionary history, habitat disturbance regimes, and anthropogenic changes: what do these mean for resilience of Pacific salmon populations? *Ecol. Soc.* **14**, art3 (2009).
- Ebeling, A. W., Laur, D. R. & Rowley, R. J. Severe storm disturbances and reversal of community structure in a southern California kelp forest. *Mar. Biol.* **84**, 287–294 (1985).
- Suding, K. N., Gross, K. L. & Houseman, G. R. Alternative states and positive feedbacks in restoration ecology. *Trends Ecol. Evol.* **19**, 46–53 (2004).
- Reice, S. R. Nonequilibrium dynamics of biological community structure. *Am. Sci.* **82**, 424–434 (1994).
- Ebeling, A. W. & Hixon, M. A. Tropical and temperate reef fishes: comparison of community structures. In *The Ecology of Fishes on Coral Reefs* (ed. Sale, P. F.) 509–563 (Elsevier), <https://doi.org/10.1016/B978-0-08-092551-6.50023-4> (1991).
- Mumby, P. J. & Steneck, R. S. Coral reef management and conservation in light of rapidly evolving ecological paradigms. *Trends Ecol. Evol.* **23**, 555–563 (2008).
- Scheffer, M. & Carpenter, S. R. Catastrophic regime shifts in ecosystems: linking theory to observation. *Trends Ecol. Evol.* **18**, 648–656 (2003).
- Lotze, H. K. *et al.* Depletion, degradation, and recovery potential of estuaries and coastal seas. *Science* **312**, 1806–1809 (2006).
- Jackson, J. B. C. *et al.* Historical overfishing and the recent collapse of coastal ecosystems. *Science* **293**, 629–637 (2001).
- Kemp, W. *et al.* Eutrophication of Chesapeake Bay: historical trends and ecological interactions. *Mar. Ecol. Prog. Ser.* **303**, 1–29 (2005).
- Elmhirst, T., Connolly, S. R. & Hughes, T. P. Connectivity, regime shifts and the resilience of coral reefs. *Coral Reefs* **28**, 949–957 (2009).
- Hughes, T. P., Graham, N. A. J., Jackson, J. B. C., Mumby, P. J. & Steneck, R. S. Rising to the challenge of sustaining coral reef resilience. *Trends Ecol. Evol.* **25**, 633–642 (2010).
- Fogarty, M. J. & Murawski, S. A. Large-scale disturbance and the structure of marine systems: Fishery impacts on Georges Bank. *Ecol. Appl.* **8**, 6–22 (1998).
- Steneck, R. S., Vavriner, J. & Leland, A. V. Accelerating trophic-level dysfunction in kelp forest ecosystems of the western North Atlantic. *Ecosystems* **7**, 323–332 (2004).
- Hughes, T. P. Catastrophes, phase shifts, and large-scale degradation of a Caribbean coral reef. *Science* **265**, 1547–1551 (1994).
- Wernberg, T. *et al.* Climate-driven regime shift of a temperate marine ecosystem. *Science* **353**, 169–172 (2016).
- McNutt, M. K. *et al.* Review of flow rate estimates of the Deepwater Horizon oil spill. *Proc. Natl. Acad. Sci.* **109**, 20260–20267 (2012).
- Deepwater Horizon Natural Resource Damage Assessment Trustees. *Deepwater Horizon Oil Spill: Final Programmatic Damage Assessment and Restoration Plan and Final Programmatic Environmental Impact Statement*. (2016).
- Valentine, D. L. *et al.* Fallout plume of submerged oil from Deepwater. *Horizon. Proc. Natl. Acad. Sci.* **111**, 15906–15911 (2014).
- Chanton, J. *et al.* Using natural abundance radiocarbon to trace the flux of petrocarbon to the seafloor following the deepwater horizon oil spill. *Environ. Sci. Technol.* **49**, 847–854 (2015).
- Passow, U. & Ziervogel, K. Marine snow sedimented oil released during the Deepwater Horizon spill. *Oceanography* **29**, 118–125 (2016).
- Hylland, K. Polycyclic aromatic hydrocarbon (PAH) ecotoxicology in marine ecosystems. *J. Toxicol. Environ. Heal. Part A* **69**, 109–123 (2006).
- Hemmer, M. J., Barron, M. G. & Greene, R. M. Comparative toxicity of eight oil dispersants, Louisiana sweet crude oil (LSC), and chemically dispersed LSC to two aquatic test species. *Environ. Toxicol. Chem.* **30**, 2244–2252 (2011).
- Holladay, S. D. *et al.* Benzo[a]pyrene-induced hypocellularity of the pronephros in tilapia (*Oreochromis niloticus*) is accompanied by alterations in stromal and parenchymal cells and by enhanced immune cell apoptosis. *Vet. Immunol. Immunopathol.* **64**, 69–82 (1998).
- Navas, J. M. & Segner, H. Antiestrogenicity of β -naphthoflavone and PAHs in cultured rainbow trout hepatocytes: evidence for a role of the arylhydrocarbon receptor. *Aquat. Toxicol.* **51**, 79–92 (2000).
- Rhodes, S., Farwell, A., Mark Hewitt, L., MacKinnon, M. & George Dixon, D. The effects of dimethylated and alkylated polycyclic aromatic hydrocarbons on the embryonic development of the Japanese medaka. *Ecotoxicol. Environ. Saf.* **60**, 247–258 (2005).
- Gilliers, C., Claireaux, G., Galois, R., Loizeau, V. & Pape, O. Le. Influence of hydrocarbons exposure on survival, growth and condition of juvenile flatfish: a mesocosm experiment. *J. Life Sci.* **4**, 113–122 (2012).
- Schlenker, L. S. *et al.* Damsels in distress: oil exposure modifies behavior and olfaction in bicolor damselfish (*Stegastes partitus*). *Environ. Sci. Technol.* **53**, 10993–11001 (2019).
- Stieglitz, J. D., Mager, E. M., Hoenig, R. H., Benetti, D. D. & Grosell, M. Impacts of Deepwater Horizon crude oil exposure on adult mahi-mahi (*Coryphaena hippurus*) swim performance. *Environ. Toxicol. Chem.* **35**, 2613–2622 (2016).
- Farr, J. A. Impairment of antipredator behavior in *Palaemonetes pugio* by exposure to sublethal doses of parathion. *Trans. Am. Fish. Soc.* **106**, 37–41 (1977).
- Whitehead, A. *et al.* Genomic and physiological footprint of the Deepwater Horizon oil spill on resident marsh fishes. *Proc. Natl. Acad. Sci.* **109**, 20298–20302 (2012).
- Barron, M. G. Ecological impacts of the Deepwater Horizon oil spill: implications for immunotoxicity. *Toxicol. Pathol.* **40**, 315–320 (2012).
- Schwacke, L. H. *et al.* Health of common bottlenose dolphins (*Tursiops truncatus*) in Barataria Bay, Louisiana, following the Deepwater Horizon oil spill. *Environ. Sci. Technol.* **48**, 93–103 (2014).
- Murawski, S. A., Hogarth, W. T., Peebles, E. B. & Barbeiri, L. Prevalence of external skin lesions and polycyclic aromatic hydrocarbon concentrations in Gulf of Mexico fishes, post-Deepwater Horizon. *Trans. Am. Fish. Soc.* **143**, 1084–1097 (2014).
- Moreno, R., Jover, L., Diez, C., Sardà, F. & Sanpera, C. Ten years after the Prestige oil spill: seabird trophic ecology as indicator of long-term effects on the coastal marine ecosystem. *PLoS One* **8**, 1–10 (2013).

39. Bascompte, J., Melián, C. J. & Sala, E. Interaction strength combinations and the overfishing. *Proc. Natl. Acad. Sci. USA* **102**, 5443–5447 (2005).
40. Fleeger, J. W., Carman, K. R. & Nisbet, R. M. Indirect effects of contaminants in aquatic ecosystems. *Sci. Total Environ.* **317**, 207–233 (2003).
41. Beck, M. W. *et al.* The role of nearshore ecosystems as fish and shellfish nurseries. *Issues Ecol.* **2003**, 1–12 (2003).
42. Larkum, A. W. D., Orth, R. J. & Duarte, C. M. *Seagrasses: Biology, Ecology, and Conservation*. (Springer Netherlands (2006).
43. Able, K. W. *et al.* Fish assemblages in Louisiana salt marshes: effects of the Macondo oil spill. *Estuaries and Coasts* **38**, 1385–1398 (2015).
44. Michel, J. *et al.* Extent and degree of shoreline oiling: Deepwater Horizon oil spill, Gulf of Mexico, USA. *Plos One* **8**, e65087 (2013).
45. Silliman, B. R. *et al.* Degradation and resilience in Louisiana salt marshes after the BP-Deepwater Horizon oil spill. *Proc. Natl. Acad. Sci.* **109**, 11234–11239 (2012).
46. Fodrie, F. J. & Heck, K. L. Response of coastal fishes to the Gulf of Mexico oil disaster. *Plos One* **6** (2011).
47. Moody, R. M., Cebrian, J. & Heck, K. L. Interannual recruitment dynamics for resident and transient marsh species: evidence for a lack of impact by the Macondo oil spill. *Plos One* **8**, 1–11 (2013).
48. Schaefer, J., Frazier, N. & Barr, J. Dynamics of near-coastal fish assemblages following the Deepwater Horizon oil spill in the northern Gulf of Mexico. *Trans. Am. Fish. Soc.* **145**, 108–119 (2016).
49. Fodrie, F. J. *et al.* Integrating organismal and population responses of estuarine fishes in macondo spill research. *Bioscience* **64**, 778–788 (2014).
50. Parsons, M. L., Morrison, W., Rabalais, N. N., Turner, R. E. & Tyre, K. N. Phytoplankton and the Macondo oil spill: A comparison of the 2010 phytoplankton assemblage to baseline conditions on the Louisiana shelf. *Environ. Pollut.* **207**, 152–160 (2015).
51. Carassou, L., Hernandez, F. J. & Graham, W. M. Change and recovery of coastal mesozooplankton community structure during the Deepwater Horizon oil spill. *Environ. Res. Lett.* **9**, 124003 (2014).
52. Abbriano, R. *et al.* Deepwater Horizon oil spill: a review of the planktonic response. *Oceanography* **24**, 294–301 (2011).
53. Landers, S. C. *et al.* Meiofauna and trace metals from sediment collections in Florida after the deepwater horizon oil spill. *Gulf Mex. Sci.* **32**, 1–10 (2014).
54. Cooksey, C. *et al.* *Ecological Condition of Coastal Ocean Waters along the U.S. Continental Shelf of Northeastern Gulf of Mexico: 2010* (2014).
55. Roth, A. M. F. & Baltz, D. M. Short-term effects of an Oil spill on marsh-edge fishes and decapod crustaceans. *Estuaries and Coasts* **32**, 565–572 (2009).
56. Paperno, R. *et al.* The disruption and recovery of fish communities in the Indian River Lagoon, Florida, following two hurricanes in 2004. *Estuaries and Coasts* **29**, 1004–1010 (2006).
57. Soto, L. A., Botello, A. V., Licea-Durán, S., Lizárraga-Partida, M. L. & Yáñez-Arancibia, A. The environmental legacy of the Ixtoc-I oil spill in Campeche Sound, southwestern Gulf of Mexico. *Front. Mar. Sci.* **1**, 1–9 (2014).
58. Murawski, S. A., Peebles, E. B., Gracia, A., Tunnell, J. W. & Armenteros, M. Comparative abundance, species composition, and demographics of continental shelf fish assemblages throughout the Gulf of Mexico. *Mar. Coast. Fish.* **10**, 325–346 (2018).
59. Garner, S. B., Boswell, K. M., Lewis, J. P., Tarnecki, J. H. & Patterson, W. F. III. Effect of reef morphology and depth on fish community and trophic structure in the northcentral Gulf of Mexico. *Estuar. Coast. Shelf Sci.* **230**, 106423 (2019).
60. Warner, R. R. & Chesson, P. L. Coexistence mediated by recruitment fluctuations: a field guide to the storage effect. *Am. Nat.* **125**, 769–787 (1985).
61. Winemiller, K. O. & Rose, K. A. Patterns of life history diversification in North American fishes: implications for population regulation. *Can. J. Fish. Aquat. Sci.* **49**, 2196–2218 (1992).
62. Watterson, J. C., Patterson, W. F. III., Shipp, R. L. & Cowan, J. H. Movement of red snapper, *Lutjanus campechanus*, in the north central Gulf of Mexico: potential effects of hurricanes. *Gulf Mex. Sci.* **16**, 92–104 (1998).
63. Tarnecki, J. H. & Patterson, W. F. III. Changes in red snapper diet and trophic ecology following the Deepwater Horizon oil spill. *Mar. Coast. Fish.* **7**, 135–147 (2015).
64. McCrawley, J. R. & Cowan, J. H. J. Seasonal and size specific diet and prey demand of Red Snapper on Alabama artificial reefs: implications for management. In *Red Snapper (Lutjanus campechanus) Ecology and Fisheries in the Gulf of Mexico* (eds. Patterson III., W. F., Cowan, J. H. J., Fitzhugh, G. R. & Nieland, D. L.) 71–96 (American Fisheries Society (2007).
65. Snyder, S. M., Pulster, E. L., Wetzel, D. L. & Murawski, S. A. PAH exposure in Gulf of Mexico demersal fishes, post-Deepwater Horizon. *Environ. Sci. Technol.* **49**, 8786–8795 (2015).
66. Ainsworth, C. H. *et al.* Impacts of the Deepwater Horizon oil spill evaluated using an end-to-end ecosystem model. *Plos One* **13**, 1–21 (2018).
67. Herdter, E. S., Chambers, D. P., Stallings, C. D. & Murawski, S. A. Did the Deepwater Horizon oil spill affect growth of red snapper in the Gulf of Mexico? *Fish. Res.* **191**, 60–68 (2017).
68. Patterson, W. F. III. *et al.* The utility of stable and radioisotopes in fish tissues as biogeochemical tracers of marine oil spill food web effects. In *Scenarios and Responses to Future Deep Oil Spills* (eds. Murawski, S. A. *et al.*) 219–238 (Springer International Publishing). https://doi.org/10.1007/978-3-030-12963-7_13 (2020).
69. Norberg, M. J. The ecology of tomtate, *Haemulon aurolineatum*, in the northern Gulf of Mexico and effects of the Deepwater Horizon oil spill. (University of South Alabama (2015).
70. Dahl, K. A., Patterson, W. F. III. & Snyder, R. Experimental assessment of lionfish removals to mitigate reef fish community shifts on northern Gulf of Mexico artificial reefs. *Mar. Ecol. Prog. Ser.* **558**, 207–221 (2016).
71. Patterson, W. F. III., Tarnecki, J. H., Addis, D. T. & Barbieri, L. R. Reef fish community structure at natural versus artificial reefs in the northern Gulf of Mexico. *Proc. 66th Gulf Caribb. Fish. Institute/Gulf Caribb. Fish. Inst.* (2014).
72. Thompson, M. J., Schroeder, W. W., Phillips, N. W. & Graham, B. D. *Ecology of Live Bottom Habitats of the Northeastern Gulf of Mexico: A Community Profile Ecology of Live Bottom Habitats of the Northeastern Gulf of Mexico: A Community Profile*. (1999).
73. Silva, M., Etnoyer, P. J. & MacDonald, I. R. Coral injuries observed at mesophotic reefs after the Deepwater Horizon oil discharge. *Deep. Res. Part II Top. Stud. Oceanogr.* **129**, 96–107 (2016).
74. Clarke, K. R., Somerfield, P. J. & Chapman, M. G. On resemblance measures for ecological studies, including taxonomic dissimilarities and a zero-adjusted Bray-Curtis coefficient for denuded assemblages. *J. Exp. Mar. Bio. Ecol.* **330**, 55–80 (2006).
75. Semmler, R. F., Hoot, W. C. & Reaka, M. L. Are mesophotic coral ecosystems distinct communities and can they serve as refugia for shallow reefs? *Coral Reefs* **36**, 433–444 (2017).
76. Oksanen, J. F. *et al.* *vegan: Community Ecology Package*. (2018).
77. Team, R. core. R: A language and environment for statistical computing. (2018).
78. Bates, D., Mächler, M., Bolker, B. & Walker, S. Fitting Linear Mixed-Effects Models Using lme4. *J. Stat. Softw.* **67** (2015).
79. Hothorn, T., Bretz, F., Westfall, P. & Heiberger, R. M. Multcomp: simultaneous inference in general parametric models. *Biometrical J.* **50**, 346–363 (2008).
80. Lo, N. C., Jacobson, L. D. & Squire, J. L. Indices of relative abundance from fish spotter data based on delta-lognormal models. *Can. J. Fish. Aquat. Sci.* **49**, 2515–2526 (1992).
81. Maunder, M. N. & Punt, A. E. Standardizing catch and effort data: a review of recent approaches. *Fish. Res.* **70**, 141–159 (2004).
82. Chagaris, D., Allen, M. & Camp, E. Modeling temporal closures in a multispecies recreational fishery reveals tradeoffs associated with species seasonality and angler effort dynamics. *Fish. Res.* **210**, 106–120 (2019).

83. Valentine, M. M. & Benfield, M. C. Characterization of epibenthic and demersal megafauna at Mississippi Canyon 252 shortly after the Deepwater Horizon oil spill. *Mar. Pollut. Bull.* **77**, 196–209 (2013).
84. Montagna, P. A. *et al.* Deep-sea benthic footprint of the Deepwater Horizon blowout. *Plos One* **8** (2013).
85. Hastings, D. W. *et al.* Changes in sediment redox conditions following the BP DWH blowout event. *Deep. Res. Part II Top. Stud. Oceanogr.* **129**, 167–178 (2016).
86. Felder, D. L. *et al.* Seaweeds and decapod crustaceans on gulf deep banks after the macondo oil spill. *Bioscience* **64**, 808–819 (2014).
87. Jones, K. M. M. Home range areas and activity centres in six species of Caribbean wrasses (Labridae). *J. Fish Biol.* **66**, 150–166 (2005).
88. Ortmann, A. C. *et al.* Dispersed oil disrupts microbial pathways in pelagic food webs. *Plos One* **7**, 1–9 (2012).
89. Paul, J. H. *et al.* Toxicity and mutagenicity of Gulf of Mexico waters during and after the Deepwater Horizon oil spill. *Environ. Sci. Technol.* **47**, 9651–9659 (2013).
90. Milinkovitch, T. *et al.* The effect of hypoxia and hydrocarbons on the anti-predator performance of European sea bass (*Dicentrarchus labrax*). *Environ. Pollut.* **251**, 581–590 (2019).
91. Coleman, F. C., Scanlon, K. M. & Koenig, C. C. Groupers on the edge: shelf edge spawning habitat in and around marine reserves of the northeastern Gulf of Mexico. *Prof. Geogr.* **63**, 456–474 (2011).
92. Addis, D. T., Patterson, W. F. III., Dance, M. A. & Ingram, G. W. Implications of reef fish movement from unreported artificial reef sites in the northern Gulf of Mexico. *Fish. Res.* **147**, 349–358 (2013).
93. Charnov, E. L., Orians, G. H. & Hyatt, K. Ecological implications of resource depression. *Am. Nat.* **110**, 247–259 (1976).
94. Werner, E. E. & Gilliam, J. F. The ontogenetic niche and species interactions in size-structured populations. *Annu. Rev. Ecol. Syst.* **15**, 393–425 (1984).
95. Kirby-Smith, W. W. & Ustach, J. Resistance to hurricane disturbance of an epifaunal community on the continental shelf off North Carolina. *Estuar. Coast. Shelf Sci.* **23**, 433–442 (1986).
96. Bell, M. & Hall, J. W. Effects of hurricane Hugo on South Carolina's marine artificial reefs. *Bull. Mar. Sci.* **55**, 836–847 (1994).
97. Smith, G. B. Relationship of eastern Gulf of Mexico reef-fish communities to the species equilibrium theory of insular biogeography. *J. Biogeogr.* **6**, 49 (1979).
98. Dupont, J. M., Hallock, P. & Jaap, W. C. Ecological impacts of the 2005 red tide on artificial reef epibenthic macroinvertebrate and fish communities in the eastern Gulf of Mexico. *Mar. Ecol. Prog. Ser.* **415**, 189–200 (2010).
99. Driggers, W. B. *et al.* Environmental conditions and catch rates of predatory fishes associated with a mass mortality on the West Florida Shelf. *Estuar. Coast. Shelf Sci.* **168**, 40–49 (2016).
100. Steidinger, K. A. & Ingle, R. M. Observations on the 1971 summer red tide in Tampa Bay, Florida. *Environ. Lett.* **3**, 271–278 (1972).
101. Ogden, J. & Ebersole, J. Scale and Community Structure of Coral Reef Fishes: A Long-Term Study of a Large Artificial Reef. *Mar. Ecol. Prog. Ser.* **4**, 97–103 (1981).
102. Smith, L. Small rotenone stations: a tool for studying coral reef fish communities. *Am. Museum Nat. Hist.* **1** (1973).
103. Dayton, P. K. Experimental evaluation of ecological dominance in a rocky intertidal algal community. *Ecol. Monogr.* **45**, 137–159 (1975).
104. Farrell, T. M. Models and mechanisms of succession: an example from a rocky intertidal community. *Ecol. Monogr.* **61**, 95–113 (1991).
105. Peterson, C. H. *et al.* Long-term ecosystem response to the Exxon Valdez oil spill. *Science*. **302**, 2082–2086 (2003).
106. Schwing, P. T. *et al.* Constraining the spatial extent of marine oil snow sedimentation and flocculent accumulation following the Deepwater Horizon event using an excess ²¹⁰Pb flux approach. *Environ. Sci. Technol.* **51**, 5962–5968 (2017).
107. Romero, I. C. *et al.* Large-scale deposition of weathered oil in the Gulf of Mexico following a deep-water oil spill. *Environ. Pollut.* **228**, 179–189 (2017).
108. Smeltz, M. *et al.* A multi-year study of hepatic biomarkers in coastal fishes from the Gulf of Mexico after the Deepwater Horizon oil spill. *Mar. Environ. Res.* **129**, 57–67 (2017).
109. Graham, W. M. *et al.* Oil carbon entered the coastal planktonic food web during the Deepwater Horizon oil spill. *Environ. Res. Lett.* **5** (2010).
110. Green, S. J., Akins, J. L., Maljković, A. & Côté, I. M. Invasive lionfish drive Atlantic coral reef fish declines. *Plos One* **7**, e32596 (2012).
111. Albins, M. A. Invasive Pacific lionfish *Pterois volitans* reduce abundance and species richness of native Bahamian coral-reef fishes. *Mar. Ecol. Prog. Ser.* **522**, 231–243 (2015).
112. Dahl, K. A. & Patterson, W. F. III. Habitat-specific density and diet of rapidly expanding invasive red lionfish, *Pterois volitans*, populations in the northern Gulf of Mexico. *Plos One* **9**, e105852 (2014).
113. Albins, M. A. & Hixon, M. A. Invasive Indo-Pacific lionfish *Pterois volitans* reduce recruitment of Atlantic coral-reef fishes. *Mar. Ecol. Prog. Ser.* **367**, 233–238 (2008).
114. Raymond, W. W., Albins, M. A. & Pusack, T. J. Competitive interactions for shelter between invasive Pacific red lionfish and native Nassau grouper. *Environ. Biol. Fishes* **98**, 57–65 (2015).
115. Albins, M. A. Effects of invasive Pacific red lionfish *Pterois volitans* versus a native predator on Bahamian coral-reef fish communities. *Biol. Invasions* **15**, 29–43 (2013).
116. Dahl, K. A., Patterson, W. F. III., Robertson, A. & Ortmann, A. C. DNA barcoding significantly improves resolution of invasive lionfish diet in the northern Gulf of Mexico. *Biol. Invasions* **19**, 1917–1933 (2017).
117. Chagaris, D. *et al.* An ecosystem-based approach to evaluating impacts and management of invasive lionfish. *Fisheries* **42**, 421–431 (2017).
118. Rooney, N., McCann, K., Gellner, G. & Moore, J. C. Structural asymmetry and the stability of diverse food webs. *Nature* **442**, 265–269 (2006).
119. National Marine Fisheries Service. *Fisheries Economics of the United States, 2010*. (2011).
120. Holland, S. M. & Ditton, R. B. Fishing trip satisfaction: a typology of anglers. *North Am. J. Fish. Manag.* **12**, 28–33 (1992).

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Author contributions

W.F.P. designed the ROV survey. J.H.T., S.B.G. and J.P.L. conducted field work. J.H.T. and S.B.G. read the ROV footage. J.P.L. and D.D.C. performed the data analysis. J.P.L. and W.F.P. wrote the main manuscript text. J.P.L. and S.B.G. prepared figures and tables. All authors reviewed the manuscript.

Competing interests

The authors declare no competing interests.

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1 *The following supplement accompanies the article*

2

3 **Changes in Reef Fish Community Structure Following the Deepwater**

4 **Horizon Oil Spill**

5

6 **Justin P. Lewis*, Joseph H. Tarnecki, Steven B. Garner, David D. Chagaris, and William F.**

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8

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10 **Table S1.** Species-specific standardized pre-*Deepwater Horizon* densities (Pre-DWH) and percent change (%) following DWH.

11 Standardized densities were estimated from generalized linear mixed effects models (GLMM) for 52 fish species. Pre-DWH densities

12 were back transformed into original measurement units (Fish \times 1000m⁻²). Cell color represents either a decrease (red), no change

13 (white), or increase (green) in fish density. Shading is along a continuum from -100 to 100%. A stroked t (†) denotes a significant

14 ($\alpha=0.05$) pairwise difference in the binomial GLMM (see Table S5). An asterisk is used to identify significant pairwise differences in

15 the log-normal GLMM (see Table S6).

Trophic Guild	Scientific Name	Pre-DWH	2010	2011	2012	2013	2014	2015	2016	2017
Herbivore	<i>Acanthurus chirurgus</i>	0.39	-99.9*	-61	-73.8	-88.2	-66.4	-98.7*	-100.0†	-98.6*
Small Demersal Browser	<i>Canthigaster rostrata</i>	0.76	-100.0†	-76.6*	-67.9*	22.9*	-13.6*	292.1†	-63.2*	-56.9*
	<i>Chaetodon ocellatus</i>	1.31	-82.6	-57.9	-38.5	-63.1	-70.1	-53.4	-86.6	-15.2
	<i>Chaetodon sedentarius</i>	1.85	-75.8*	3.5	-30.3*	36	-24.2*	7.5*	-39.5*	-93.1*
	<i>Parablennius marmoratus</i>	1.95	-100.0†	-29.2	-40.7	-69.9	-68	-100.0†	-44.8	-82.3
	<i>Prognathodes aya</i>	0.52	-97.5*	-41.8*	-55.5*	-40.3*	-41.5	2.5*	-100.0†	-67.8*
	<i>Stegastes leucostictus</i>	1.18	-96.7	-78.2	-75.5	-3.8	21.7	-10.7*	-53.7	-64.6
	<i>Stegastes variabilis</i>	2.15	-83.0*	-93.2*	-94.2*	-92.4*	-97.5	-100.0†	-88.7*	-100.0†
	<i>Aluterus monoceros</i>	0.00	0.0	1	4.2	7.1	9.1	2.1	0	2.2
Large Demersal Browser	<i>Holacanthus bermudensis</i>	3.34	-69.2*	-44.1*	-45.7	-13.4	-44.5	-39.4	-31.1	-62.8
	<i>Holacanthus ciliaris</i>	0.45	-51.5	-50.4*	-90.5*	-72.6*	-83.1*	-92.8*	-82.2	-90.3*
	<i>Bodianus pulchellus</i>	0.86	-37.8*	-60.7	-59.6	-8.2*	-73.3*	-22.1*	-66.8*	-74.2*
Small Demersal Invertivore	<i>Bodianus rufus</i>	0.13	-100.0†	-82.2	-85.8*	-41.5	-63.6*	-68.3*	57.8*	-93.7*
	<i>Equetus lanceolatus</i>	0.57	-89.1*	-82.3	-88.5	-9.6	-52.5*	-37.2	-45.4	31*
	<i>Haemulon aurolineatum</i>	129.53	-64.3	-79.3	-8.4	5	-55.8	-54.9	-36.3	4
	<i>Halichoeres bivittatus</i>	8.15	-98.6†	-83.0*	-80.2†	-86.4†	-92.5†	-56.0	-90.2*	-85.0*
	<i>Holocentrus adscensionis</i>	0.28	-100.0†	-40.8	-6.3	16.3	-88.2	65.4	-83.2	-71.9
	<i>Holocentrus rufus</i>	0.00	0.0	6.9	7.1	0	59.2	0	17.1	10.9
	<i>Pareques acuminatus</i>	0.80	-100.0†	-47.4*	-89.2*	22.2*	-89.2*	-100.0†	-81.0	-100.0†
	<i>Pareques umbrosus</i>	9.50	-85.5*	-68.4*	-69.9	-70.3*	-50.3	0.6	-85.2*	-84.3*
	<i>Archosargus probatocephalus</i>	0.00	-100.0†	-35.0	340.9	-64.4	375.7	-85.5	-100.0†	-86.4
Large Demersal Invertivore	<i>Balistes capriscus</i>	5.11	-40.4	-61.1	-58.8	-37.2	-72.4	0.2	-6.1	35.8
	<i>Chilomycterus schoepfi</i>	0.00	0	9.5	12.9	0	18.2	12.9	24.1	5.9
	<i>Pagrus pagrus</i>	5.76	-86.1*	-14.8	-70.9*	-53.9	-55.2	-88.0	-88.2*	-70.7

Trophic Guild	Scientific Name	Pre-DWH	2010	2011	2012	2013	2014	2015	2016	2017
Generalist Carnivore	<i>Caranx crysos</i>	8.52	-84.0	-99.3	-46.3	-75.4	263.8	-87.2	-100.0†	-94.9
	<i>Centropristis ocyurus</i>	2.52	-97.8	-31.5	-57.9	-47.1	-44.9	-53.7	-88.5*	-96.4*
	<i>Diplectrum formosum</i>	0.00	0.0	9.0	28.8	30.8	7.2	90.2†	0.0	0.0
	<i>Epinephelus cruentatus</i>	0.12	-100.0†	-84.6	-54.7	-52.9	-44.1	64.2	-73.0	68.0
	<i>Epinephelus morio</i>	1.66	-69.9*	-80.1*	-84.2*	-82.6*	-90.7*	-86.5*	-95.8*	-100.0†
	<i>Liopropoma eukrines</i>	0.09	15.1*	141.2	300.2	220.9	420.9	985.3†	-54.2*	127.5
	<i>Lutjanus campechanus</i>	74.59	-68.9*	-65.5*	-72.4*	-78.1*	-83.1*	-90.7*	-89.6*	-81.1*
	<i>Lutjanus griseus</i>	19.49	-84.9	-87.5*	-78.2	-83.5	-95.1*	-70.8	-96.4*	-88.5*
	<i>Lutjanus synagris</i>	4.21	-26.1	-96.6†	-84.2	-49.1	-94.5	-83.6	-92.8	-80.3
	<i>Pterois volitans</i>	0.00	0.0	1.7	76.1†	305.1†	349.6†	332.4†	353.7†	580.2†
	<i>Rypticus maculatus</i>	1.06	-80.0	28.2	-24.3	-14.3	-10.9	19.9	-81.5	-28.9
	<i>Seriola dumerili</i>	13.30	53.8	21.5	-35.6	-65.6	-86.5	-59.6	-90.1	-77.6
	<i>Seriola rivoliana</i>	3.60	-100.0†	-91.6†	-88.7*	-81.2	-94.0	-47.9	-74.1	-92.9*
	<i>Serranus phoebe</i>	1.50	-79.4*	-73.6	2.7	-70.1*	-27.3*	-5.9	-73.4*	-58.9*
	<i>Serranus subligarius</i>	1.44	-100.0†	-54.9	-30.6	-90.5	-91.7	-92.1	-89.5	-89.4
Piscivore	<i>Carcharhinus plumbeus</i>	0.45	-100.0†	-81.1	-75.5	-48.0	-94.2*	-84.9	-69.1*	-92.7*
	<i>Mycteroperca microlepis</i>	2.18	-90.3*	-89.3*	-75.5*	-93.9†	-79.0*	-84.0*	-74.6	-97.9*
	<i>Mycteroperca phenax</i>	5.75	-79.3*	-35.7	-59.3*	-61.2*	-82.8*	-55.3*	-85.2*	-70.4*
Reef Planktivore	<i>Apogon pseudomaculatus</i>	0.74	-79.0	34.9	-84.3*	14.1	-52.9	-16.7	-14.5	-100.0†
	<i>Baldwinella aureorubens</i>	0.00	0.0	0.4	0.0	3.6	2.0	1260.9*	3.4	5.4
	<i>Chromis enchrysur</i>	10.86	-92.3†	-64.7	-70.3	26.3	-30.2	190.9	-51.3	-85.9
	<i>Paranthias furcifer</i>	0.44	-90.5	0.8	-48.4	-89.8	-70.2	351.5	-100.0†	-100.0†
	<i>Priacanthus arenatus</i>	1.74	-77.4*	82.8	133.6	143.7	-18.9	-89.5	-100.0†	-73.6
	<i>Pristigenys alta</i>	0.69	-100.0†	-96.2†	-100.0†	-91.8	-20.6	536.0†	669.3†	504.8†
	<i>Pronotogrammus martinicensis</i>	0.01	15.4	514.3	344.2	1686.2	1622.7	4939.5	134.6	154.5
	<i>Rhomboplites aurorubens</i>	107.84	-35.7	-61.9*	-76.3	12.5	-48.3	66.6	5.0	80.5
	<i>Stegastes partitus</i>	0.20	-98.6*	-100.0†	-100.0†	-95.6	-78.3	-45.1	-97.9	-100.0†
	Damselfish	7.39	-99.8	-97.0*	-93.7†	91.8	170.5	480.7	-81.0	141.4
--	Total Fish	555.54	-61.8	-54.4	-35.4	7.2	-38.3	25.0	-32.8	-29.4

19 **Table S2.** Dunnett's multiple comparisons of generalize linear models assessing changes in species richness,
20 diversity, and evenness and total fish density following the *Deepwater Horizon* oil spill (DHW). Asterisks (*) denote
21 significant differences at $\alpha=0.05$.

Community index	Comparison	Estimate	SE	z	P
Species richness	2010 - Pre-DWH	-6.19	1.48	-4.19	<0.01*
	2011 - Pre-DWH	-2.23	1.13	-1.97	0.25
	2012 - Pre-DWH	-2.07	1.15	-1.81	0.33
	2013 - Pre-DWH	-0.45	1.12	-0.4	1
	2014 - Pre-DWH	0.28	1.27	0.22	1
	2015 - Pre-DWH	4.93	1.44	3.42	<0.01*
	2016 - Pre-DWH	-1.64	1.44	-1.14	0.8
	2017 - Pre-DWH	0.42	1.47	0.29	1
Species diversity (Shannon-Weiner H')	2010 - Pre-DWH	-0.42	0.16	-2.6	0.06
	2011 - Pre-DWH	-0.14	0.12	-1.15	0.8
	2012 - Pre-DWH	-0.42	0.12	-3.36	<0.01*
	2013 - Pre-DWH	-0.5	0.12	-4.09	<0.01*
	2014 - Pre-DWH	-0.37	0.14	-2.65	0.05
	2015 - Pre-DWH	-0.13	0.16	-0.84	0.95
	2016 - Pre-DWH	-0.56	0.16	-3.55	<0.01*
	2017 - Pre-DWH	-0.28	0.16	-1.77	0.36
Species evenness (Pielou's J')	2010 - Pre-DWH	-0.08	0.06	-1.37	0.64
	2011 - Pre-DWH	-0.02	0.04	-0.43	1
	2012 - Pre-DWH	-0.13	0.04	-2.94	0.02*
	2013 - Pre-DWH	-0.18	0.04	-4.11	<0.01*
	2014 - Pre-DWH	-0.13	0.05	-2.64	0.05
	2015 - Pre-DWH	-0.09	0.06	-1.66	0.43
	2016 - Pre-DWH	-0.16	0.06	-2.94	0.02*
	2017 - Pre-DWH	-0.09	0.06	-1.66	0.43
Total fish density	2010 - Pre-DWH	-0.92	0.39	-2.35	0.11
	2011 - Pre-DWH	-0.78	0.3	-2.6	0.06
	2012 - Pre-DWH	-0.65	0.3	-2.14	0.17
	2013 - Pre-DWH	0.04	0.3	0.13	1
	2014 - Pre-DWH	-0.36	0.34	-1.06	0.85
	2015 - Pre-DWH	0.25	0.38	0.66	0.99
	2016 - Pre-DWH	-0.35	0.38	-0.93	0.92
	2017 - Pre-DWH	-0.31	0.39	-0.79	0.97

Table S3. Results from Dunnett's multiple comparisons of linear mixed effects model comparing trophic guild presence/absence between pre- and post-DWH time points. No pair-wise differences were significant ($\alpha = 0.05$).

Trophic guild	Comparison	Estimate	SE	z	P
Herbivore	2010 - Pre-DWH	-1.78	1.28	-1.39	0.64
	2011 - Pre-DWH	-3.1	1.24	-2.51	0.07
	2012 - Pre-DWH	-1.12	0.84	-1.32	0.69
	2013 - Pre-DWH	-2.37	1.01	-2.35	0.11
	2014 - Pre-DWH	-0.37	0.88	-0.42	1
	2015 - Pre-DWH	-1.9	1.28	-1.48	0.57
	2016 - Pre-DWH	--	--	--	--
	2017 - Pre-DWH	-1.81	1.29	-1.41	0.63
Small Demersal Browser	2010 - Pre-DWH	-0.99	0.86	-1.15	0.81
	2011 - Pre-DWH	0.45	0.75	0.6	0.99
	2012 - Pre-DWH	-0.25	0.72	-0.35	1
	2013 - Pre-DWH	1.08	0.77	1.4	0.64
	2014 - Pre-DWH	0.42	0.82	0.51	1
	2015 - Pre-DWH	2.65	1.34	1.97	0.26
	2016 - Pre-DWH	0.86	1.01	0.85	0.96
	2017 - Pre-DWH	0.22	0.95	0.23	1
Large Demersal Browser	2010 - Pre-DWH	0.29	0.86	0.34	1
	2011 - Pre-DWH	-0.15	0.65	-0.23	1
	2012 - Pre-DWH	-0.3	0.66	-0.46	1
	2013 - Pre-DWH	0.63	0.66	0.95	0.92
	2014 - Pre-DWH	-0.7	0.73	-0.96	0.91
	2015 - Pre-DWH	0.11	0.85	0.13	1
	2016 - Pre-DWH	0.11	0.85	0.13	1
	2017 - Pre-DWH	0.04	0.87	0.04	1
Small Demersal Invertivore (with tomtate)	2010 - Pre-DWH	-1.37	1.18	-1.16	0.75
	2011 - Pre-DWH	-0.61	1	-0.61	0.99
	2012 - Pre-DWH	-1.4	0.96	-1.45	0.54
	2013 - Pre-DWH	-1.05	0.97	-1.09	0.8
	2014 - Pre-DWH	-1.46	1.02	-1.44	0.55
	2015 - Pre-DWH	0.84	1.48	0.57	0.99
	2016 - Pre-DWH	-0.26	1.27	-0.21	1
	2017 - Pre-DWH	-0.29	1.28	-0.23	1

Trophic guild	Comparison	Estimate	SE	z	P
Small Demersal Invertivore (without tomtate)	2010 - Pre-DWH	-1.75	0.97	-1.81	0.32
	2011 - Pre-DWH	-0.73	0.83	-0.87	0.93
	2012 - Pre-DWH	-1.3	0.82	-1.59	0.46
	2013 - Pre-DWH	-0.66	0.83	-0.8	0.95
	2014 - Pre-DWH	-1	0.88	-1.13	0.79
	2015 - Pre-DWH	1.49	1.39	1.07	0.82
	2016 - Pre-DWH	-0.27	1.07	-0.25	1
	2017 - Pre-DWH	0.43	1.17	0.37	1
Large Demersal Invertivore	2010 - Pre-DWH	0.52	0.95	0.55	1
	2011 - Pre-DWH	0.87	0.69	1.27	0.74
	2012 - Pre-DWH	-0.58	0.64	-0.91	0.94
	2013 - Pre-DWH	0.38	0.65	0.58	1
	2014 - Pre-DWH	-1.01	0.69	-1.46	0.59
	2015 - Pre-DWH	1.96	1.19	1.65	0.45
	2016 - Pre-DWH	0.08	0.83	0.1	1
	2017 - Pre-DWH	-0.03	0.84	-0.04	1
Generalist Carnivore	2010 - Pre-DWH	-0.73	1.5	-0.49	1
	2011 - Pre-DWH	0.79	1.49	0.53	1
	2012 - Pre-DWH	0.04	1.31	0.03	1
	2013 - Pre-DWH	0.1	1.3	0.08	1
	2014 - Pre-DWH	0.22	1.49	0.15	1
	2015 - Pre-DWH	-0.34	1.51	-0.23	1
	2016 - Pre-DWH	-0.34	1.51	-0.23	1
	2017 - Pre-DWH	-0.38	1.51	-0.25	1
Piscivore	2010 - Pre-DWH	-1	1.01	-0.99	0.85
	2011 - Pre-DWH	-0.94	0.78	-1.21	0.69
	2012 - Pre-DWH	-1.01	0.78	-1.3	0.63
	2013 - Pre-DWH	-0.62	0.78	-0.8	0.94
	2014 - Pre-DWH	-1.23	0.84	-1.47	0.51
	2015 - Pre-DWH	0.79	1.09	0.73	0.96
	2016 - Pre-DWH	-0.76	0.93	-0.82	0.93
	2017 - Pre-DWH	-0.37	0.96	-0.38	1
Reef Planktivore	2010 - Pre-DWH	-2.56	1.62	-1.58	0.44
	2011 - Pre-DWH	-0.37	1.27	-0.29	1
	2012 - Pre-DWH	-1.54	1.27	-1.21	0.71
	2013 - Pre-DWH	-1.26	1.27	-1	0.86
	2014 - Pre-DWH	-0.48	1.33	-0.36	1
	2015 - Pre-DWH	1.91	1.65	1.16	0.75
	2016 - Pre-DWH	1.91	1.65	1.16	0.75
	2017 - Pre-DWH	0.66	1.52	0.43	1

27 **Table S4.** Results from Dunnett's multiple comparisons of the log-normal linear mixed effects model comparing pre-
28 and post-DWH time points. Asterisks (*) denote significant differences at $\alpha = 0.05$.

Trophic guild	Comparison	Estimate	SE	z	P
Herbivore	2010 - Pre-DWH	-1.53	0.44	-3.47	<0.01*
	2011 - Pre-DWH	1.07	0.65	1.65	0.46
	2012 - Pre-DWH	-0.45	0.35	-1.31	0.7
	2013 - Pre-DWH	-0.27	0.35	-0.76	0.97
	2014 - Pre-DWH	-0.56	0.38	-1.48	0.58
	2015 - Pre-DWH	-1.2	0.44	-2.71	0.04*
	2016 - Pre-DWH	--	--	--	--
	2017 - Pre-DWH	-1.2	0.44	-2.71	0.04*
Small Demersal Browser	2010 - Pre-DWH	-1.78	0.29	-6.13	<0.01*
	2011 - Pre-DWH	-0.79	0.21	-3.79	<0.01*
	2012 - Pre-DWH	-0.77	0.22	-3.56	<0.01*
	2013 - Pre-DWH	-0.69	0.2	-3.43	<0.01*
	2014 - Pre-DWH	-0.79	0.24	-3.35	<0.01*
	2015 - Pre-DWH	-0.76	0.25	-3.05	0.02*
	2016 - Pre-DWH	-1.17	0.26	-4.53	<0.01*
	2017 - Pre-DWH	-1.28	0.27	-4.69	<0.01*
Large Demersal Browser	2010 - Pre-DWH	-1.06	0.21	-5.04	<0.01*
	2011 - Pre-DWH	-0.53	0.18	-3.02	0.02*
	2012 - Pre-DWH	-0.54	0.18	-3.02	0.02*
	2013 - Pre-DWH	-0.37	0.17	-2.18	0.15
	2014 - Pre-DWH	-0.32	0.21	-1.52	0.51
	2015 - Pre-DWH	-0.55	0.22	-2.46	0.08
	2016 - Pre-DWH	-0.53	0.22	-2.42	0.09
	2017 - Pre-DWH	-0.77	0.23	-3.42	<0.01*
Small Demersal Invertivore (with tomtate)	2010 - Pre-DWH	-1.38	0.59	-2.33	0.12
	2011 - Pre-DWH	-1.41	0.45	-3.13	0.01*
	2012 - Pre-DWH	-0.29	0.46	-0.62	0.99
	2013 - Pre-DWH	-0.12	0.45	-0.28	1
	2014 - Pre-DWH	-0.37	0.53	-0.7	0.98
	2015 - Pre-DWH	-0.56	0.56	-1.01	0.89
	2016 - Pre-DWH	-0.6	0.57	-1.05	0.86
	2017 - Pre-DWH	-0.03	0.58	-0.05	1

Trophic guild	Comparison	Estimate	SE	z	P
Small Demersal Invertivore (without tomtate)	2010 - Pre-DWH	-1.5	0.37	-4.06	<0.01*
	2011 - Pre-DWH	-0.89	0.27	-3.32	<0.01*
	2012 - Pre-DWH	-0.99	0.28	-3.58	<0.01*
	2013 - Pre-DWH	-0.79	0.27	-2.96	0.02*
	2014 - Pre-DWH	-0.72	0.31	-2.3	0.12
	2015 - Pre-DWH	-0.41	0.33	-1.26	0.74
	2016 - Pre-DWH	-1.45	0.34	-4.22	<0.01*
	2017 - Pre-DWH	-1.3	0.34	-3.79	<0.01*
Large Demersal Invertivore	2010 - Pre-DWH	-1.17	0.4	-2.88	0.03*
	2011 - Pre-DWH	-0.54	0.31	-1.72	0.38
	2012 - Pre-DWH	-0.73	0.33	-2.17	0.16
	2013 - Pre-DWH	-0.72	0.31	-2.29	0.12
	2014 - Pre-DWH	-0.62	0.39	-1.57	0.48
	2015 - Pre-DWH	-1.07	0.38	-2.78	0.04*
	2016 - Pre-DWH	-0.85	0.41	-2.08	0.2
	2017 - Pre-DWH	-0.56	0.42	-1.33	0.67
Generalist Carnivore	2010 - Pre-DWH	-0.98	0.35	-2.77	0.04*
	2011 - Pre-DWH	-1.09	0.27	-4.03	<0.01*
	2012 - Pre-DWH	-0.98	0.28	-3.58	<0.01*
	2013 - Pre-DWH	-1.04	0.27	-3.87	<0.01*
	2014 - Pre-DWH	-1.6	0.3	-5.27	<0.01*
	2015 - Pre-DWH	-1.28	0.34	-3.71	<0.01*
	2016 - Pre-DWH	-1.97	0.34	-5.72	<0.01*
	2017 - Pre-DWH	-1.43	0.35	-4.07	<0.01*
Piscivore	2010 - Pre-DWH	-1.1	0.26	-4.2	<0.01*
	2011 - Pre-DWH	-0.64	0.2	-3.11	0.01*
	2012 - Pre-DWH	-0.8	0.21	-3.81	<0.01*
	2013 - Pre-DWH	-1	0.2	-5.02	<0.01*
	2014 - Pre-DWH	-1.07	0.24	-4.49	<0.01*
	2015 - Pre-DWH	-0.89	0.25	-3.6	<0.01*
	2016 - Pre-DWH	-0.98	0.27	-3.7	<0.01*
	2017 - Pre-DWH	-1.24	0.26	-4.67	<0.01*
Reef Planktivore	2010 - Pre-DWH	-0.25	0.52	-0.48	1
	2011 - Pre-DWH	-0.78	0.38	-2.05	0.22
	2012 - Pre-DWH	-0.91	0.39	-2.32	0.12
	2013 - Pre-DWH	0.59	0.38	1.54	0.52
	2014 - Pre-DWH	0.03	0.44	0.07	1
	2015 - Pre-DWH	1.21	0.47	2.57	0.06
	2016 - Pre-DWH	0.22	0.47	0.46	1
	2017 - Pre-DWH	0.42	0.49	0.85	0.95

32 **Table S5.** Results from the linear mixed effects model comparing species-specific presence/absence between pre- and post-DWH time
33 points. For species not observed during the pre-DWH time bin, comparisons were made between the first and subsequent years with
34 positive density estimates. A stroked t (†) denote significant differences at $\alpha = 0.05$.

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Herbivore	<i>Acanthurus chirurgus</i>	2010 - Pre-DWH	-1.57	1.34	-1.17	0.8
		2011 - Pre-DWH	-2.9	1.28	-2.27	0.13
		2012 - Pre-DWH	-1.17	0.94	-1.25	0.74
		2013 - Pre-DWH	-2.92	1.28	-2.29	0.13
		2014 - Pre-DWH	-0.53	0.99	-0.54	1
		2015 - Pre-DWH	-1.67	1.33	-1.26	0.74
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-1.57	1.33	-1.18	0.79
Small Demersal Browser	<i>Canthigaster rostrata</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.42	0.73	-0.57	0.99
		2012 - Pre-DWH	-0.17	0.74	-0.22	1
		2013 - Pre-DWH	1.51	0.67	2.26	0.12
		2014 - Pre-DWH	1.17	0.75	1.56	0.46
		2015 - Pre-DWH	3.27	0.94	3.49	<0.01†
		2016 - Pre-DWH	0.86	0.85	1.01	0.85
		2017 - Pre-DWH	0.57	0.89	0.65	0.98
	<i>Chaetodon ocellatus</i>	2010 - Pre-DWH	0.11	1	0.11	1
		2011 - Pre-DWH	-0.27	0.83	-0.32	1
		2012 - Pre-DWH	-0.08	0.63	-0.13	1
		2013 - Pre-DWH	0.04	0.63	0.06	1
		2014 - Pre-DWH	-0.15	0.62	-0.25	1
		2015 - Pre-DWH	0.16	0.7	0.23	1
		2016 - Pre-DWH	0.77	0.75	1.03	0.87
		2017 - Pre-DWH	-0.81	0.92	-0.88	0.94

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Browser	<i>Chaetodon sedentarius</i>	2010 - Pre-DWH	-0.02	0.84	-0.03	1
		2011 - Pre-DWH	0.66	0.66	1.01	0.89
		2012 - Pre-DWH	0.57	0.68	0.84	0.96
		2013 - Pre-DWH	1.04	0.65	1.59	0.5
		2014 - Pre-DWH	0.83	0.76	1.09	0.85
		2015 - Pre-DWH	1.54	0.89	1.72	0.41
		2016 - Pre-DWH	0.57	0.85	0.66	0.99
		2017 - Pre-DWH	-0.61	0.93	-0.66	0.99
	<i>Parablennius marmoreus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.06	0.57	-0.1	1
		2012 - Pre-DWH	-0.28	0.58	-0.47	0.99
		2013 - Pre-DWH	-0.91	0.62	-1.47	0.51
		2014 - Pre-DWH	-1.13	0.78	-1.45	0.52
		2015 - Pre-DWH	--	--	--	--
		2016 - Pre-DWH	-1.09	0.89	-1.23	0.68
		2017 - Pre-DWH	-1.02	0.9	-1.14	0.74
	<i>Prognathodes aya</i>	2010 - Pre-DWH	-1.71	1.22	-1.4	0.6
		2011 - Pre-DWH	0.21	0.73	0.29	1
		2012 - Pre-DWH	0.15	0.76	0.2	1
		2013 - Pre-DWH	0.49	0.73	0.68	0.98
		2014 - Pre-DWH	0.17	0.84	0.2	1
		2015 - Pre-DWH	1.04	0.95	1.1	0.82
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	0.11	1	0.11	1
	<i>Stegastes leucostictus</i>	2010 - Pre-DWH	-1.44	1.29	-1.12	0.79
		2011 - Pre-DWH	-0.65	0.82	-0.79	0.96
		2012 - Pre-DWH	-0.5	0.81	-0.62	0.99
		2013 - Pre-DWH	1.2	0.76	1.58	0.46
		2014 - Pre-DWH	1.79	0.85	2.11	0.17
		2015 - Pre-DWH	2.02	0.94	2.16	0.16
		2016 - Pre-DWH	0.78	0.93	0.84	0.94
		2017 - Pre-DWH	-1.42	0.56	-2.55	0.06

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Browser	<i>Stegastes variabilis</i>	2010 - Pre-DWH	-0.47	0.94	-0.49	0.99
		2011 - Pre-DWH	-1.53	0.81	-1.88	0.26
		2012 - Pre-DWH	-1.24	0.76	-1.62	0.42
		2013 - Pre-DWH	-1.25	0.76	-1.65	0.4
		2014 - Pre-DWH	-2.01	1.16	-1.74	0.35
		2015 - Pre-DWH	--	--	--	--
		2016 - Pre-DWH	-0.28	0.85	-0.33	1
		2017 - Pre-DWH	--	--	--	--
Large Demersal Browser	<i>Aluterus monoceros</i>	2012 - 2011	1.35	0.98	1.38	0.5
		2013 - 2011	0.91	0.99	0.92	0.83
		2014 - 2011	1.81	1.12	1.62	0.35
		2015 - 2011	0.51	1.4	0.36	1
		2016 - 2011	--	--	--	--
		2017 - 2011	0.53	1.4	0.38	1
	<i>Holacanthus bermudensis</i>	2010 - Pre-DWH	0.11	0.79	0.14	1
		2011 - Pre-DWH	-0.04	0.61	-0.06	1
		2012 - Pre-DWH	-0.18	0.62	-0.29	1
		2013 - Pre-DWH	0.66	0.62	1.05	0.87
		2014 - Pre-DWH	-0.42	0.69	-0.61	0.99
		2015 - Pre-DWH	0.35	0.81	0.43	1
		2016 - Pre-DWH	0.35	0.81	0.43	1
		2017 - Pre-DWH	-0.59	0.8	-0.73	0.98
	<i>Holacanthus ciliaris</i>	2010 - Pre-DWH	-0.91	1.31	-0.7	0.99
		2011 - Pre-DWH	0.36	0.83	0.43	1
		2012 - Pre-DWH	-1.24	1.04	-1.2	0.8
		2013 - Pre-DWH	-0.6	0.9	-0.67	0.99
		2014 - Pre-DWH	-0.55	1.05	-0.53	1
		2015 - Pre-DWH	-0.94	1.29	-0.73	0.98
		2016 - Pre-DWH	-0.94	1.29	-0.73	0.98
		2017 - Pre-DWH	-0.88	1.3	-0.68	0.99

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Invertivore	<i>Bodianus pulchellus</i>	2010 - Pre-DWH	0.53	0.85	0.62	0.99
		2011 - Pre-DWH	-0.12	0.68	-0.18	1
		2012 - Pre-DWH	-0.6	0.72	-0.83	0.96
		2013 - Pre-DWH	0.84	0.68	1.24	0.75
		2014 - Pre-DWH	-0.31	0.82	-0.38	1
		2015 - Pre-DWH	1.17	0.9	1.31	0.71
		2016 - Pre-DWH	0.68	0.89	0.77	0.97
		2017 - Pre-DWH	-0.01	0.93	-0.01	1
	<i>Bodianus rufus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-1.24	1.55	-0.8	0.91
		2012 - Pre-DWH	0.16	1.31	0.12	1
		2013 - Pre-DWH	0.03	1.31	0.02	1
		2014 - Pre-DWH	1.47	1.35	1.09	0.73
		2015 - Pre-DWH	1.01	1.44	0.7	0.95
		2016 - Pre-DWH	1.72	1.39	1.24	0.62
		2017 - Pre-DWH	0.06	1.61	0.04	1
	<i>Equetus lanceolatus</i>	2010 - Pre-DWH	-0.27	1.05	-0.26	1
		2011 - Pre-DWH	-0.7	0.81	-0.87	0.95
		2012 - Pre-DWH	-1.44	0.88	-1.63	0.47
		2013 - Pre-DWH	0.55	0.78	0.71	0.98
		2014 - Pre-DWH	0.67	0.89	0.75	0.98
		2015 - Pre-DWH	0.76	1.02	0.75	0.98
		2016 - Pre-DWH	0.12	1.02	0.12	1
		2017 - Pre-DWH	2.14	1.08	1.98	0.25
	<i>Haemulon aurolineatum</i>	2010 - Pre-DWH	-0.78	0.75	-1.05	0.87
		2011 - Pre-DWH	-0.52	0.58	-0.9	0.94
		2012 - Pre-DWH	0.02	0.59	0.03	1
		2013 - Pre-DWH	0.16	0.58	0.28	1
		2014 - Pre-DWH	0.29	0.67	0.44	1
		2015 - Pre-DWH	0.67	0.79	0.85	0.95
		2016 - Pre-DWH	0.67	0.79	0.85	0.95
		2017 - Pre-DWH	1.55	0.9	1.71	0.41

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Invertivore	<i>Halichoeres bivittatus</i>	2010 - Pre-DWH	-3.06	1.13	-2.72	0.04†
		2011 - Pre-DWH	-1.19	0.54	-2.22	0.15
		2012 - Pre-DWH	-1.51	0.55	-2.73	0.04†
		2013 - Pre-DWH	-1.94	0.57	-3.42	<0.01†
		2014 - Pre-DWH	-1.93	0.67	-2.9	0.03†
		2015 - Pre-DWH	0.03	0.69	0.04	1
		2016 - Pre-DWH	-0.79	0.68	-1.17	0.81
		2017 - Pre-DWH	-0.67	0.69	-0.98	0.91
	<i>Holocentrus adscensionis</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.39	0.82	-0.47	1
		2012 - Pre-DWH	0.16	0.79	0.2	1
		2013 - Pre-DWH	0.34	0.76	0.45	1
		2014 - Pre-DWH	-1.23	1.22	-1	0.86
		2015 - Pre-DWH	1.14	0.88	1.3	0.66
		2016 - Pre-DWH	-1.01	1.24	-0.81	0.94
		2017 - Pre-DWH	-0.9	1.25	-0.72	0.97
	<i>Holocentrus rufus</i>	2012 - 2011	-0.74	0.94	-0.79	0.87
		2013 - 2011	--	--	--	--
		2014 - 2011	0.8	0.9	0.89	0.82
		2015 - 2011	--	--	--	--
		2016 - 2011	0.51	1	0.51	0.97
		2017 - 2011	0.76	1.02	0.75	0.89
	<i>Pareques acuminatus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-1.52	0.93	-1.62	0.38
		2012 - Pre-DWH	-1.54	0.94	-1.64	0.37
		2013 - Pre-DWH	-2.28	1.17	-1.94	0.21
		2014 - Pre-DWH	-1	0.95	-1.05	0.79
		2015 - Pre-DWH	--	--	--	--
		2016 - Pre-DWH	-0.3	0.97	-0.31	1
		2017 - Pre-DWH	--	--	--	--

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Invertivore	<i>Pareques umbrosus</i>	2010 - Pre-DWH	-0.3	0.71	-0.42	1
		2011 - Pre-DWH	-0.07	0.55	-0.13	1
		2012 - Pre-DWH	-0.51	0.56	-0.91	0.93
		2013 - Pre-DWH	-0.16	0.54	-0.29	1
		2014 - Pre-DWH	0.48	0.62	0.77	0.97
		2015 - Pre-DWH	-0.06	0.7	-0.08	1
		2016 - Pre-DWH	-0.06	0.7	-0.08	1
		2017 - Pre-DWH	-0.52	0.73	-0.7	0.98
Large Demersal Invertivore	<i>Archosargus probatocephalus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.3	1.46	-0.2	1
		2012 - Pre-DWH	1.36	1.32	1.03	0.73
		2013 - Pre-DWH	0.13	1.4	0.1	1
		2014 - Pre-DWH	0.82	1.44	0.57	0.97
		2015 - Pre-DWH	0.13	1.72	0.08	1
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	0.13	1.72	0.08	1
	<i>Balistes caprisus</i>	2010 - Pre-DWH	-0.32	0.84	-0.38	1
		2011 - Pre-DWH	-0.6	0.63	-0.94	0.91
		2012 - Pre-DWH	-0.53	0.63	-0.83	0.95
		2013 - Pre-DWH	0.14	0.63	0.22	1
		2014 - Pre-DWH	-0.47	0.71	-0.67	0.99
		2015 - Pre-DWH	0.67	0.82	0.81	0.96
		2016 - Pre-DWH	0.67	0.82	0.81	0.96
		2017 - Pre-DWH	1.49	0.9	1.65	0.43
	<i>Chilomycterus schoepfi</i>	2012 - 2011	-0.29	0.82	-0.36	1
		2013 - 2011	--	--	--	--
		2014 - 2011	0.54	0.85	0.64	0.97
		2015 - 2011	0.45	0.96	0.47	0.99
		2016 - 2011	0.98	0.87	1.12	0.73
		2017 - 2011	-0.29	1.2	-0.25	1

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Large Demersal Invertivore	<i>Pagrus pagrus</i>	2010 - Pre-DWH	0.12	0.77	0.15	1
		2011 - Pre-DWH	1.44	0.62	2.3	0.12
		2012 - Pre-DWH	0.05	0.62	0.08	1
		2013 - Pre-DWH	0.25	0.6	0.42	1
		2014 - Pre-DWH	0.39	0.68	0.57	0.99
		2015 - Pre-DWH	-0.61	0.81	-0.76	0.97
		2016 - Pre-DWH	-0.22	0.78	-0.28	1
		2017 - Pre-DWH	0.26	0.78	0.33	1
Generalist Carnivore	<i>Caranx crysos</i>	2010 - Pre-DWH	-0.7	1.24	-0.56	0.99
		2011 - Pre-DWH	-1.17	0.99	-1.18	0.78
		2012 - Pre-DWH	0.06	0.81	0.08	1
		2013 - Pre-DWH	-1.22	0.99	-1.23	0.74
		2014 - Pre-DWH	-1.24	1.22	-1.01	0.88
		2015 - Pre-DWH	0.55	0.95	0.58	0.99
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-0.73	1.25	-0.59	0.99
	<i>Centropristis ocyurus</i>	2010 - Pre-DWH	-1.75	1.14	-1.55	0.53
		2011 - Pre-DWH	0.13	0.59	0.22	1
		2012 - Pre-DWH	-0.44	0.61	-0.73	0.98
		2013 - Pre-DWH	-0.17	0.59	-0.29	1
		2014 - Pre-DWH	0.4	0.66	0.6	0.99
		2015 - Pre-DWH	0.52	0.74	0.7	0.99
		2016 - Pre-DWH	-0.56	0.79	-0.7	0.98
		2017 - Pre-DWH	-1.45	0.94	-1.53	0.54
	<i>Diplectrum formosum</i>	2012 - 2011	1.47	1.16	1.27	0.44
		2013 - 2011	1.74	1.14	1.53	0.29
		2014 - 2011	0.59	1.46	0.41	0.97
		2015 - 2011	2.93	1.22	2.41	0.05†

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Generalist Carnivore	<i>Epinephelus cruentatus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.57	0.98	-0.58	0.99
		2012 - Pre-DWH	0.3	0.93	0.32	1
		2013 - Pre-DWH	0.57	0.88	0.65	0.98
		2014 - Pre-DWH	0.83	1	0.83	0.93
		2015 - Pre-DWH	1.67	1.08	1.54	0.46
		2016 - Pre-DWH	0.32	1.16	0.27	1
		2017 - Pre-DWH	1.71	1.09	1.57	0.44
	<i>Epinephelus morio</i>	2010 - Pre-DWH	-0.23	0.74	-0.31	1
		2011 - Pre-DWH	-1.16	0.6	-1.92	0.28
		2012 - Pre-DWH	-1.36	0.64	-2.13	0.18
		2013 - Pre-DWH	-1.18	0.6	-1.97	0.25
		2014 - Pre-DWH	-1.61	0.78	-2.07	0.2
		2015 - Pre-DWH	-1.18	0.81	-1.46	0.58
		2016 - Pre-DWH	-2.52	1.14	-2.2	0.15
		2017 - Pre-DWH	--	--	--	--
	<i>Liopropoma eukrines</i>	2010 - Pre-DWH	1.2	0.95	1.26	0.71
		2011 - Pre-DWH	1.19	0.79	1.5	0.53
		2012 - Pre-DWH	2.09	0.82	2.54	0.07
		2013 - Pre-DWH	1.73	0.79	2.2	0.15
		2014 - Pre-DWH	2.32	0.91	2.56	0.06
		2015 - Pre-DWH	3.56	1.05	3.39	<0.01†
		2016 - Pre-DWH	0.74	1.03	0.71	0.98
		2017 - Pre-DWH	1.03	1.06	0.97	0.9
	<i>Lutjanus campechanus</i>	2010 - Pre-DWH	-0.98	1.13	-0.87	0.91
		2011 - Pre-DWH	-0.52	0.94	-0.55	0.99
		2012 - Pre-DWH	-1.53	0.9	-1.7	0.36
		2013 - Pre-DWH	-0.88	0.9	-0.97	0.86
		2014 - Pre-DWH	-1.09	0.99	-1.11	0.78
		2015 - Pre-DWH	-1.71	1.03	-1.66	0.38
		2016 - Pre-DWH	-1.24	1.06	-1.17	0.73
		2017 - Pre-DWH	0.12	1.34	0.09	1

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Generalist Carnivore	<i>Lutjanus griseus</i>	2010 - Pre-DWH	-1.03	0.93	-1.1	0.81
		2011 - Pre-DWH	-0.79	0.69	-1.13	0.79
		2012 - Pre-DWH	-0.94	0.7	-1.34	0.64
		2013 - Pre-DWH	-1.39	0.7	-1.98	0.23
		2014 - Pre-DWH	-1.33	0.79	-1.69	0.39
		2015 - Pre-DWH	-0.81	0.86	-0.94	0.9
		2016 - Pre-DWH	-1.24	0.88	-1.41	0.58
		2017 - Pre-DWH	-0.31	0.87	-0.35	1
	<i>Lutjanus synagris</i>	2010 - Pre-DWH	-1.38	0.93	-1.48	0.59
		2011 - Pre-DWH	-2.72	0.89	-3.07	0.02†
		2012 - Pre-DWH	-0.73	0.62	-1.18	0.81
		2013 - Pre-DWH	-1.15	0.64	-1.8	0.36
		2014 - Pre-DWH	-1.22	0.75	-1.62	0.49
		2015 - Pre-DWH	-0.58	0.78	-0.75	0.98
		2016 - Pre-DWH	-1.52	0.92	-1.65	0.47
		2017 - Pre-DWH	-0.15	0.76	-0.2	1
	<i>Pterois volitans</i>	2012 - 2011	3.64	1.12	3.25	<0.01†
		2013 - 2011	5.2	1.14	4.55	<0.01†
		2014 - 2011	4.93	1.2	4.12	<0.01†
		2015 - 2011	6.44	1.34	4.79	<0.01†
		2016 - 2011	5.94	1.3	4.55	<0.01†
		2017 - 2011	6.37	1.35	4.71	<0.01†
	<i>Rypticus maculatus</i>	2010 - Pre-DWH	-1.05	1	-1.06	0.85
		2011 - Pre-DWH	0.54	0.64	0.83	0.95
		2012 - Pre-DWH	-0.05	0.66	-0.07	1
		2013 - Pre-DWH	0.03	0.65	0.05	1
		2014 - Pre-DWH	0.31	0.72	0.43	1
		2015 - Pre-DWH	0.98	0.79	1.24	0.73
		2016 - Pre-DWH	-1.61	1.02	-1.58	0.47
		2017 - Pre-DWH	0.11	0.83	0.13	1

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Generalist Carnivore	<i>Seriola dumerili</i>	2010 - Pre-DWH	-0.08	0.83	-0.1	1
		2011 - Pre-DWH	-1.35	0.62	-2.18	0.16
		2012 - Pre-DWH	-0.81	0.63	-1.3	0.7
		2013 - Pre-DWH	-1.09	0.61	-1.78	0.35
		2014 - Pre-DWH	-1.48	0.7	-2.12	0.18
		2015 - Pre-DWH	0.63	0.85	0.74	0.98
		2016 - Pre-DWH	-0.97	0.78	-1.24	0.74
		2017 - Pre-DWH	-0.26	0.8	-0.33	1
	<i>Seriola rivoliana</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-1.86	0.68	-2.75	0.04†
		2012 - Pre-DWH	-1.36	0.65	-2.1	0.18
		2013 - Pre-DWH	-1.36	0.63	-2.18	0.15
		2014 - Pre-DWH	-1.73	0.8	-2.16	0.16
		2015 - Pre-DWH	-0.23	0.74	-0.32	1
		2016 - Pre-DWH	-1.36	0.84	-1.63	0.44
		2017 - Pre-DWH	-1.24	0.85	-1.47	0.55
	<i>Serranus phoebe</i>	2010 - Pre-DWH	-0.56	0.96	-0.58	0.99
		2011 - Pre-DWH	0.13	0.66	0.2	1
		2012 - Pre-DWH	0.38	0.67	0.56	1
		2013 - Pre-DWH	0.05	0.66	0.08	1
		2014 - Pre-DWH	1.33	0.73	1.81	0.33
		2015 - Pre-DWH	1.65	0.82	2	0.23
		2016 - Pre-DWH	0.48	0.83	0.57	1
		2017 - Pre-DWH	1.17	0.83	1.4	0.62
	<i>Serranus subligarius</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.56	0.74	-0.76	0.97
		2012 - Pre-DWH	-0.83	0.75	-1.11	0.83
		2013 - Pre-DWH	-2.07	0.89	-2.33	0.11
		2014 - Pre-DWH	-2.81	1.24	-2.26	0.13
		2015 - Pre-DWH	-1.14	1.04	-1.1	0.84
		2016 - Pre-DWH	-2.07	1.27	-1.63	0.46
		2017 - Pre-DWH	-2.04	1.28	-1.6	0.48

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Piscivore	<i>Carcharhinus plumbeus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-1.89	1.19	-1.59	0.5
		2012 - Pre-DWH	-1.15	0.97	-1.19	0.79
		2013 - Pre-DWH	-0.48	0.82	-0.58	0.99
		2014 - Pre-DWH	-1.2	1.2	-1	0.89
		2015 - Pre-DWH	-0.79	1.21	-0.65	0.99
		2016 - Pre-DWH	-0.02	0.99	-0.02	1
		2017 - Pre-DWH	-0.72	1.22	-0.59	0.99
	<i>Mycteroperca microlepis</i>	2010 - Pre-DWH	-1.09	0.9	-1.22	0.8
		2011 - Pre-DWH	-1.49	0.68	-2.19	0.17
		2012 - Pre-DWH	-0.52	0.6	-0.87	0.96
		2013 - Pre-DWH	-2.11	0.77	-2.73	0.04†
		2014 - Pre-DWH	-0.33	0.66	-0.49	1
		2015 - Pre-DWH	-0.41	0.76	-0.54	1
		2016 - Pre-DWH	-0.81	0.81	-1	0.92
		2017 - Pre-DWH	-1.99	1.15	-1.73	0.42
	<i>Mycteroperca phenax</i>	2010 - Pre-DWH	-0.88	0.84	-1.04	0.84
		2011 - Pre-DWH	0.28	0.69	0.41	1
		2012 - Pre-DWH	-0.37	0.68	-0.54	0.99
		2013 - Pre-DWH	-0.45	0.67	-0.68	0.98
		2014 - Pre-DWH	-1.05	0.74	-1.42	0.57
		2015 - Pre-DWH	0.05	0.84	0.06	1
		2016 - Pre-DWH	-1.05	0.82	-1.29	0.67
		2017 - Pre-DWH	0.01	0.85	0.01	1
Reef Planktivore	<i>Apogon pseudomaculatus</i>	2010 - Pre-DWH	-0.05	1.27	-0.04	1
		2011 - Pre-DWH	0.83	0.85	0.97	0.84
		2012 - Pre-DWH	0.52	0.89	0.58	0.99
		2013 - Pre-DWH	0.14	0.91	0.15	1
		2014 - Pre-DWH	0.9	0.92	0.97	0.84
		2015 - Pre-DWH	1.08	0.98	1.09	0.76
		2016 - Pre-DWH	-0.17	1.27	-0.13	1
		2017 - Pre-DWH	--	--	--	--

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Reef Planktivore	<i>Baldwinella aureorubens</i>	2012 - 2011	--	--	--	--
		2013 - 2011	0.41	0.97	0.42	0.99
		2014 - 2011	-0.32	1.37	-0.24	1
		2015 - 2011	2.85	1.23	2.32	0.09
		2016 - 2011	-0.05	1.43	-0.04	1
		2017 - 2011	2.26	1.26	1.79	0.28
	<i>Chromis enchrysur</i>	2010 - Pre-DWH	-2.94	0.99	-2.95	0.02*
		2011 - Pre-DWH	-1.01	0.67	-1.5	0.57
		2012 - Pre-DWH	-0.93	0.69	-1.34	0.69
		2013 - Pre-DWH	-0.13	0.67	-0.2	1
		2014 - Pre-DWH	0.28	0.79	0.36	1
		2015 - Pre-DWH	2.56	1.13	2.26	0.14
		2016 - Pre-DWH	0.03	0.89	0.04	1
		2017 - Pre-DWH	-1.05	0.88	-1.18	0.8
	<i>Paranthias furcifer</i>	2010 - Pre-DWH	-0.26	0.85	-0.31	1
		2011 - Pre-DWH	1.55	0.76	2.04	0.21
		2012 - Pre-DWH	1.48	0.77	1.91	0.27
		2013 - Pre-DWH	2.03	0.78	2.6	0.05
		2014 - Pre-DWH	0.27	0.79	0.34	1
		2015 - Pre-DWH	-3.19	1.27	-2.51	0.07
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-1.43	0.96	-1.49	0.54
	<i>Priacanthus arenatus</i>	2010 - Pre-DWH	-0.26	0.85	-0.31	1
		2011 - Pre-DWH	1.55	0.76	2.04	0.21
		2012 - Pre-DWH	1.48	0.77	1.91	0.27
		2013 - Pre-DWH	2.03	0.78	2.6	0.05
		2014 - Pre-DWH	0.27	0.79	0.34	1
		2015 - Pre-DWH	-3.19	1.27	-2.51	0.07
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-1.43	0.96	-1.49	0.54

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Reef Planktivore	<i>Pristigenys alta</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-3.12	1.15	-2.72	0.04†
		2012 - Pre-DWH	--	--	--	--
		2013 - Pre-DWH	-3.14	1.14	-2.74	0.03*
		2014 - Pre-DWH	-0.26	0.85	-0.3	1
		2015 - Pre-DWH	4.25	1.34	3.17	<0.01†
		2016 - Pre-DWH	2.79	1.04	2.68	0.04†
		2017 - Pre-DWH	2.78	1.04	2.66	0.04†
	<i>Pronotoqrammus martinicensis</i>	2010 - Pre-DWH	0.02	1.19	0.02	1
		2011 - Pre-DWH	1.16	1.09	1.07	0.85
		2012 - Pre-DWH	1.03	1.11	0.92	0.93
		2013 - Pre-DWH	2.26	1.16	1.95	0.26
		2014 - Pre-DWH	2.38	1.36	1.75	0.38
		2015 - Pre-DWH	3.2	1.6	2	0.24
		2016 - Pre-DWH	0.71	1.34	0.53	1
		2017 - Pre-DWH	1.54	1.51	1.02	0.88
	<i>Rhomboplites aurorubens</i>	2010 - Pre-DWH	0.31	0.95	0.32	1
		2011 - Pre-DWH	-0.1	0.68	-0.15	1
		2012 - Pre-DWH	0.29	0.69	0.41	1
		2013 - Pre-DWH	0.38	0.68	0.56	1
		2014 - Pre-DWH	0.17	0.76	0.22	1
		2015 - Pre-DWH	2.35	1.08	2.17	0.17
		2016 - Pre-DWH	2.35	1.08	2.17	0.17
		2017 - Pre-DWH	1.59	0.99	1.6	0.48
	<i>Stegastes partitus</i>	2010 - Pre-DWH	-1.32	1.43	-0.93	0.84
		2011 - Pre-DWH	--	--	--	--
		2012 - Pre-DWH	--	--	--	--
		2013 - Pre-DWH	-1.81	1.16	-1.56	0.41
		2014 - Pre-DWH	-0.67	1.17	-0.57	0.98
		2015 - Pre-DWH	0.52	1.17	0.44	0.99
		2016 - Pre-DWH	-1.28	1.44	-0.89	0.86
		2017 - Pre-DWH	--	--	--	--

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Reef Planktivore	Damselfish	2010 - Pre-DWH	-2.1	1.08	-1.95	0.26
		2011 - Pre-DWH	-1.33	0.78	-1.72	0.39
		2012 - Pre-DWH	-2.44	0.87	-2.8	0.03t
		2013 - Pre-DWH	0.71	0.74	0.95	0.91
		2014 - Pre-DWH	0.06	0.85	0.07	1
		2015 - Pre-DWH	2.42	0.98	2.47	0.08
		2016 - Pre-DWH	-1.41	0.99	-1.42	0.6
		2017 - Pre-DWH	0.72	0.95	0.76	0.97

62 **Table S6.** Results from the linear mixed effects model comparing species-specific abundance estimates between pre- and post-DWH
63 time points. For species not observed during the pre-DWH time bin, comparisons were made between the first and subsequent years
64 with positive density estimates. Asterisks (*) denote significant differences at $\alpha = 0.05$.

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Herbivore	<i>Acanthurus chirurgus</i>	2010 - Pre-DWH	-1.71	0.46	-3.72	<0.01*
		2011 - Pre-DWH	1.67	0.81	2.07	0.2
		2012 - Pre-DWH	-0.19	0.43	-0.44	1
		2013 - Pre-DWH	0.58	0.53	1.11	0.82
		2014 - Pre-DWH	-0.47	0.41	-1.16	0.79
		2015 - Pre-DWH	-1.38	0.46	-3	0.02*
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-1.38	0.46	-3	0.02*
Small Demersal Browser	<i>Canthigaster rostrata</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.88	0.28	-3.16	<0.01*
		2012 - Pre-DWH	-0.82	0.28	-2.87	0.02*
		2013 - Pre-DWH	-0.82	0.25	-3.33	<0.01*
		2014 - Pre-DWH	-0.91	0.27	-3.41	<0.01*
		2015 - Pre-DWH	-0.53	0.25	-2.12	0.14
		2016 - Pre-DWH	-1.29	0.3	-4.3	<0.01*
		2017 - Pre-DWH	-1.04	0.32	-3.25	<0.01*
	<i>Chaetodon ocellatus</i>	2010 - Pre-DWH	-1.14	0.28	-4.09	<0.01*
		2011 - Pre-DWH	-0.66	0.21	-3.1	0.01*
		2012 - Pre-DWH	-0.44	0.22	-2.02	0.22
		2013 - Pre-DWH	-0.72	0.21	-3.38	<0.01*
		2014 - Pre-DWH	-1.03	0.23	-4.41	<0.01*
		2015 - Pre-DWH	-1.02	0.23	-4.48	<0.01*
		2016 - Pre-DWH	-0.99	0.33	-2.95	0.02*
		2017 - Pre-DWH	-0.51	0.24	-2.09	0.19

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Browser	<i>Chaetodon sedentarius</i>	2010 - Pre-DWH	-1.13	0.32	-3.55	<0.01*
		2011 - Pre-DWH	-0.41	0.23	-1.74	0.36
		2012 - Pre-DWH	-0.68	0.24	-2.83	0.03*
		2013 - Pre-DWH	-0.38	0.23	-1.66	0.41
		2014 - Pre-DWH	-0.75	0.26	-2.82	0.03*
		2015 - Pre-DWH	-0.76	0.27	-2.81	0.03*
		2016 - Pre-DWH	-0.77	0.29	-2.66	0.05*
		2017 - Pre-DWH	-1.58	0.35	-4.5	<0.01*
	<i>Parablennius marmoreus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.27	0.35	-0.79	0.94
		2012 - Pre-DWH	-0.28	0.36	-0.78	0.94
		2013 - Pre-DWH	-0.41	0.41	-1	0.84
		2014 - Pre-DWH	-0.2	0.51	-0.4	1
		2015 - Pre-DWH	--	--	--	--
		2016 - Pre-DWH	0.25	0.59	0.42	1
		2017 - Pre-DWH	-0.71	0.6	-1.2	0.71
	<i>Prognathodes aya</i>	2010 - Pre-DWH	-1.43	0.44	-3.25	<0.01*
		2011 - Pre-DWH	-0.64	0.21	-3.04	0.02*
		2012 - Pre-DWH	-0.8	0.22	-3.62	<0.01*
		2013 - Pre-DWH	-0.82	0.2	-4.04	<0.01*
		2014 - Pre-DWH	-0.61	0.24	-2.47	0.08
		2015 - Pre-DWH	-0.76	0.24	-3.1	0.01*
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-0.99	0.29	-3.37	<0.01*
	<i>Stegastes leucostictus</i>	2010 - Pre-DWH	-1.55	0.92	-1.69	0.35
		2011 - Pre-DWH	-0.83	0.55	-1.5	0.47
		2012 - Pre-DWH	-0.84	0.53	-1.6	0.41
		2013 - Pre-DWH	-0.91	0.47	-1.91	0.23
		2014 - Pre-DWH	-1.02	0.48	-2.1	0.16
		2015 - Pre-DWH	-1.33	0.51	-2.6	0.05*
		2016 - Pre-DWH	-1.19	0.57	-2.1	0.16
		2017 - Pre-DWH	-1.42	0.56	-2.55	0.06

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Browser	<i>Stegastes variabilis</i>	2010 - Pre-DWH	-1.14	0.36	-3.16	<0.01*
		2011 - Pre-DWH	-1.12	0.35	-3.17	<0.01*
		2012 - Pre-DWH	-1.41	0.37	-3.77	<0.01*
		2013 - Pre-DWH	-1.23	0.31	-3.93	<0.01*
		2014 - Pre-DWH	-1.44	0.6	-2.39	0.09
		2015 - Pre-DWH	--	--	--	--
		2016 - Pre-DWH	-1.52	0.31	-4.96	<0.01*
		2017 - Pre-DWH	--	--	--	--
Large Demersal Browser	<i>Aluterus monoceros</i>	2012 - 2011	-0.02	0.5	-0.04	1
		2013 - 2011	0.67	0.48	1.4	0.48
		2014 - 2011	0.22	0.47	0.47	0.99
		2015 - 2011	0.16	0.64	0.26	1
		2016 - 2011	--	--	--	--
		2017 - 2011	0.16	0.64	0.26	1
	<i>Holacanthus bermudensis</i>	2010 - Pre-DWH	-0.93	0.21	-4.42	<0.01*
		2011 - Pre-DWH	-0.47	0.17	-2.79	0.03*
		2012 - Pre-DWH	-0.44	0.17	-2.53	0.07
		2013 - Pre-DWH	-0.33	0.16	-2.01	0.22
		2014 - Pre-DWH	-0.33	0.2	-1.61	0.45
		2015 - Pre-DWH	-0.52	0.21	-2.44	0.08
		2016 - Pre-DWH	-0.41	0.21	-1.97	0.24
		2017 - Pre-DWH	-0.56	0.23	-2.38	0.1
	<i>Holacanthus ciliaris</i>	2010 - Pre-DWH	0.11	0.28	0.38	1
		2011 - Pre-DWH	-0.8	0.19	-4.16	<0.01*
		2012 - Pre-DWH	-0.87	0.25	-3.47	<0.01*
		2013 - Pre-DWH	-0.59	0.21	-2.76	0.04*
		2014 - Pre-DWH	-0.93	0.23	-3.99	<0.01*
		2015 - Pre-DWH	-1.16	0.3	-3.94	<0.01*
		2016 - Pre-DWH	-0.63	0.29	-2.13	0.19
		2017 - Pre-DWH	-1.04	0.3	-3.45	<0.01*

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Invertivore	<i>Bodianus pulchellus</i>	2010 - Pre-DWH	-0.77	0.29	-2.64	0.05*
		2011 - Pre-DWH	-0.7	0.27	-2.59	0.06
		2012 - Pre-DWH	-0.32	0.29	-1.11	0.79
		2013 - Pre-DWH	-0.67	0.24	-2.75	0.04*
		2014 - Pre-DWH	-0.86	0.31	-2.72	0.04*
		2015 - Pre-DWH	-0.95	0.29	-3.29	<0.01*
		2016 - Pre-DWH	-1.24	0.3	-4.16	<0.01*
		2017 - Pre-DWH	-1.05	0.33	-3.23	<0.01*
	<i>Bodianus rufus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.48	0.41	-1.17	0.64
		2012 - Pre-DWH	-1.69	0.33	-5.2	<0.01*
		2013 - Pre-DWH	-0.54	0.32	-1.73	0.29
		2014 - Pre-DWH	-1.87	0.37	-5	<0.01*
		2015 - Pre-DWH	-1.69	0.42	-4.07	<0.01*
		2016 - Pre-DWH	-1.07	0.34	-3.14	<0.01*
		2017 - Pre-DWH	-1.96	0.41	-4.76	<0.01*
	<i>Equetus lanceolatus</i>	2010 - Pre-DWH	-1.52	0.47	-3.2	<0.01*
		2011 - Pre-DWH	-0.92	0.36	-2.57	0.06
		2012 - Pre-DWH	-0.69	0.4	-1.7	0.4
		2013 - Pre-DWH	-0.56	0.33	-1.7	0.4
		2014 - Pre-DWH	-1.15	0.37	-3.09	0.01*
		2015 - Pre-DWH	-0.99	0.41	-2.42	0.09
		2016 - Pre-DWH	-0.62	0.43	-1.43	0.59
		2017 - Pre-DWH	-1.24	0.37	-3.33	<0.01*
	<i>Haemulon aurolineatum</i>	2010 - Pre-DWH	-0.62	0.7	-0.88	0.94
		2011 - Pre-DWH	-1.32	0.53	-2.5	0.08
		2012 - Pre-DWH	-0.1	0.51	-0.2	1
		2013 - Pre-DWH	-0.03	0.51	-0.06	1
		2014 - Pre-DWH	-0.93	0.57	-1.64	0.44
		2015 - Pre-DWH	-1.01	0.62	-1.63	0.45
		2016 - Pre-DWH	-0.67	0.62	-1.08	0.84
		2017 - Pre-DWH	-0.37	0.6	-0.61	0.99

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Invertivore	<i>Halichoeres bivittatus</i>	2010 - Pre-DWH	-1.63	0.91	-1.79	0.39
		2011 - Pre-DWH	-1.02	0.33	-3.12	0.01*
		2012 - Pre-DWH	-0.7	0.34	-2.07	0.23
		2013 - Pre-DWH	-0.74	0.37	-2.01	0.26
		2014 - Pre-DWH	-1.22	0.46	-2.65	0.06
		2015 - Pre-DWH	-0.74	0.36	-2.02	0.25
		2016 - Pre-DWH	-1.58	0.41	-3.83	<0.01*
		2017 - Pre-DWH	-1.34	0.41	-3.31	<0.01*
	<i>Holocentrus adscensionis</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.14	0.33	-0.43	1
		2012 - Pre-DWH	-0.17	0.29	-0.59	0.99
		2013 - Pre-DWH	-0.13	0.27	-0.47	1
		2014 - Pre-DWH	-0.58	0.48	-1.22	0.71
		2015 - Pre-DWH	-0.32	0.3	-1.08	0.81
		2016 - Pre-DWH	-0.5	0.48	-1.05	0.83
		2017 - Pre-DWH	-0.25	0.48	-0.52	0.99
	<i>Holocentrus rufus</i>	2012 - 2011	0.53	0.48	1.11	0.65
		2013 - 2011	--	--	--	--
		2014 - 2011	1.05	0.44	2.39	0.06
		2015 - 2011	--	--	--	--
		2016 - 2011	0.3	0.48	0.63	0.93
		2017 - 2011	-0.14	0.49	-0.3	1
	<i>Pareques acuminatus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	0.66	0.21	3.14	<0.01*
		2012 - Pre-DWH	-0.64	0.22	-2.89	0.02*
		2013 - Pre-DWH	2.19	0.28	7.82	<0.01*
		2014 - Pre-DWH	-0.96	0.23	-4.12	<0.01*
		2015 - Pre-DWH	--	--	--	--
		2016 - Pre-DWH	-1.02	0.22	-4.64	<0.01*
		2017 - Pre-DWH	--	--	--	--

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Small Demersal Invertivore	<i>Pareques umbrosus</i>	2010 - Pre-DWH	-1.56	0.48	-3.27	<0.01*
		2011 - Pre-DWH	-1.04	0.36	-2.85	0.03*
		2012 - Pre-DWH	-0.83	0.38	-2.19	0.16
		2013 - Pre-DWH	-1.05	0.37	-2.87	0.03*
		2014 - Pre-DWH	-0.89	0.39	-2.26	0.14
		2015 - Pre-DWH	0.04	0.47	0.09	1
		2016 - Pre-DWH	-1.65	0.47	-3.53	<0.01*
		2017 - Pre-DWH	-1.38	0.52	-2.67	0.05*
Large Demersal Invertivore	<i>Archosargus probatocephalus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.19	1.67	-0.11	1
		2012 - Pre-DWH	-0.12	1.16	-0.1	1
		2013 - Pre-DWH	-0.93	1.23	-0.76	0.9
		2014 - Pre-DWH	0.45	1.27	0.36	1
		2015 - Pre-DWH	-1.21	1.55	-0.79	0.88
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-1.21	1.55	-0.79	0.88
	<i>Balistes caprisus</i>	2010 - Pre-DWH	-0.3	0.42	-0.72	0.97
		2011 - Pre-DWH	-0.54	0.32	-1.68	0.39
		2012 - Pre-DWH	-0.53	0.32	-1.68	0.39
		2013 - Pre-DWH	-0.49	0.3	-1.61	0.44
		2014 - Pre-DWH	-0.88	0.36	-2.46	0.08
		2015 - Pre-DWH	-0.25	0.37	-0.67	0.98
		2016 - Pre-DWH	-0.3	0.37	-0.83	0.94
		2017 - Pre-DWH	-0.17	0.36	-0.48	1
	<i>Chilomycterus schoepfi</i>	2012 - 2011	0.38	0.16	2.32	0.09
		2013 - 2011	--	--	--	--
		2014 - 2011	0.09	0.16	0.58	0.98
		2015 - 2011	-0.06	0.18	-0.31	1
		2016 - 2011	0.03	0.15	0.18	1
		2017 - 2011	-0.08	0.22	-0.38	1

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Large Demersal Invertivore	<i>Pagrus pagrus</i>	2010 - Pre-DWH	-1.76	0.48	-3.7	<0.01*
		2011 - Pre-DWH	-0.86	0.36	-2.41	0.09
		2012 - Pre-DWH	-1.16	0.4	-2.86	0.03*
		2013 - Pre-DWH	-0.88	0.38	-2.35	0.11
		2014 - Pre-DWH	-0.98	0.44	-2.23	0.15
		2015 - Pre-DWH	-1.45	0.57	-2.54	0.07
		2016 - Pre-DWH	-1.69	0.54	-3.15	0.01*
		2017 - Pre-DWH	-1.26	0.5	-2.5	0.08
Generalist Carnivore	<i>Caranx crysos</i>	2010 - Pre-DWH	-1.18	1.42	-0.83	0.92
		2011 - Pre-DWH	-3.37	1.6	-2.1	0.16
		2012 - Pre-DWH	-0.67	1.3	-0.52	0.99
		2013 - Pre-DWH	-0.28	1.54	-0.18	1
		2014 - Pre-DWH	2.43	1.42	1.71	0.34
		2015 - Pre-DWH	-2.39	1.61	-1.49	0.49
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-2.13	1.91	-1.11	0.76
	<i>Centropristis ocyurus</i>	2010 - Pre-DWH	-1.66	0.79	-2.1	0.2
		2011 - Pre-DWH	-0.42	0.31	-1.35	0.68
		2012 - Pre-DWH	-0.48	0.33	-1.44	0.62
		2013 - Pre-DWH	-0.46	0.32	-1.44	0.61
		2014 - Pre-DWH	-0.74	0.33	-2.23	0.15
		2015 - Pre-DWH	-0.92	0.38	-2.46	0.09
		2016 - Pre-DWH	-1.33	0.45	-2.95	0.02*
		2017 - Pre-DWH	-1.56	0.59	-2.65	0.05*
	<i>Diplectrum formosum</i>	2012 - 2011	-0.25	0.64	-0.4	0.97
		2013 - 2011	-0.4	0.63	-0.64	0.87
		2014 - 2011	-0.61	0.84	-0.73	0.82
		2015 - 2011	-0.35	0.62	-0.56	0.91

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Generalist Carnivore	<i>Epinephelus cruentatus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-1.06	0.33	-3.17	<0.01*
		2012 - Pre-DWH	-0.91	0.29	-3.14	<0.01*
		2013 - Pre-DWH	-1.07	0.27	-4.03	<0.01*
		2014 - Pre-DWH	-1.12	0.28	-3.98	<0.01*
		2015 - Pre-DWH	-0.93	0.3	-3.15	<0.01*
		2016 - Pre-DWH	-1.25	0.34	-3.63	<0.01*
		2017 - Pre-DWH	-0.92	0.3	-3.11	0.01*
	<i>Epinephelus morio</i>	2010 - Pre-DWH	-0.77	0.13	-5.92	<0.01*
		2011 - Pre-DWH	-0.57	0.12	-4.89	<0.01*
		2012 - Pre-DWH	-0.6	0.12	-4.85	<0.01*
		2013 - Pre-DWH	-0.64	0.12	-5.39	<0.01*
		2014 - Pre-DWH	-0.8	0.17	-4.8	<0.01*
		2015 - Pre-DWH	-0.8	0.16	-4.95	<0.01*
		2016 - Pre-DWH	-0.75	0.25	-2.98	0.02*
		2017 - Pre-DWH	--	--	--	--
	<i>Liopropoma eukrines</i>	2010 - Pre-DWH	-0.73	0.24	-3.1	0.01*
		2011 - Pre-DWH	-0.21	0.21	-1.04	0.8
		2012 - Pre-DWH	-0.44	0.2	-2.23	0.12
		2013 - Pre-DWH	-0.38	0.19	-1.94	0.22
		2014 - Pre-DWH	-0.39	0.21	-1.88	0.25
		2015 - Pre-DWH	-0.47	0.21	-2.22	0.12
		2016 - Pre-DWH	-0.98	0.26	-3.75	<0.01*
		2017 - Pre-DWH	-0.14	0.26	-0.54	0.99
	<i>Lutjanus campechanus</i>	2010 - Pre-DWH	-1.07	0.38	-2.86	0.03*
		2011 - Pre-DWH	-1.02	0.28	-3.65	<0.01*
		2012 - Pre-DWH	-1.13	0.29	-3.86	<0.01*
		2013 - Pre-DWH	-1.43	0.28	-5.18	<0.01*
		2014 - Pre-DWH	-1.64	0.32	-5.15	<0.01*
		2015 - Pre-DWH	-2.1	0.38	-5.52	<0.01*
		2016 - Pre-DWH	-2.06	0.37	-5.62	<0.01*
		2017 - Pre-DWH	-1.61	0.36	-4.49	<0.01*

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Generalist Carnivore	<i>Lutjanus griseus</i>	2010 - Pre-DWH	-1.28	0.56	-2.27	0.14
		2011 - Pre-DWH	-1.59	0.44	-3.61	<0.01*
		2012 - Pre-DWH	-1.02	0.44	-2.34	0.12
		2013 - Pre-DWH	-1	0.44	-2.25	0.14
		2014 - Pre-DWH	-2.07	0.52	-3.97	<0.01*
		2015 - Pre-DWH	-0.8	0.57	-1.4	0.64
		2016 - Pre-DWH	-2.31	0.6	-3.84	<0.01*
		2017 - Pre-DWH	-1.85	0.56	-3.32	<0.01*
	<i>Lutjanus synagris</i>	2010 - Pre-DWH	0.74	0.83	0.89	0.96
		2011 - Pre-DWH	-0.91	0.84	-1.09	0.88
		2012 - Pre-DWH	-1.13	0.52	-2.19	0.18
		2013 - Pre-DWH	0.19	0.55	0.34	1
		2014 - Pre-DWH	-1.58	0.64	-2.48	0.09
		2015 - Pre-DWH	-1.17	0.65	-1.81	0.38
		2016 - Pre-DWH	-1.17	0.92	-1.28	0.77
		2017 - Pre-DWH	-1.29	0.59	-2.17	0.19
	<i>Pterois volitans</i>	2012 - 2011	0.25	0.64	0.39	0.93
		2013 - 2011	0.62	0.63	0.98	0.51
		2014 - 2011	0.84	0.65	1.29	0.33
		2015 - 2011	0.41	0.64	0.63	0.76
		2016 - 2011	0.54	0.65	0.83	0.61
		2017 - 2011	0.89	0.65	1.37	0.29
	<i>Rypticus maculatus</i>	2010 - Pre-DWH	-0.53	0.44	-1.22	0.71
		2011 - Pre-DWH	-0.11	0.25	-0.42	1
		2012 - Pre-DWH	-0.2	0.27	-0.76	0.96
		2013 - Pre-DWH	-0.16	0.26	-0.6	0.99
		2014 - Pre-DWH	-0.27	0.29	-0.94	0.89
		2015 - Pre-DWH	-0.33	0.29	-1.13	0.78
		2016 - Pre-DWH	-0.22	0.44	-0.51	1
		2017 - Pre-DWH	-0.32	0.33	-0.98	0.87

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Generalist Carnivore	<i>Seriola dumerili</i>	2010 - Pre-DWH	0.45	0.54	0.83	0.96
		2011 - Pre-DWH	0.76	0.45	1.69	0.43
		2012 - Pre-DWH	-0.13	0.44	-0.29	1
		2013 - Pre-DWH	-0.58	0.44	-1.33	0.69
		2014 - Pre-DWH	-1.19	0.56	-2.11	0.19
		2015 - Pre-DWH	-0.96	0.51	-1.88	0.31
		2016 - Pre-DWH	-1.65	0.59	-2.8	0.04*
		2017 - Pre-DWH	-1.26	0.56	-2.27	0.14
	<i>Seriola rivoliana</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.87	0.37	-2.36	0.11
		2012 - Pre-DWH	-0.98	0.34	-2.84	0.03*
		2013 - Pre-DWH	-0.57	0.32	-1.8	0.35
		2014 - Pre-DWH	-1.18	0.45	-2.6	0.06
		2015 - Pre-DWH	-0.45	0.36	-1.22	0.77
		2016 - Pre-DWH	-0.29	0.45	-0.64	0.99
		2017 - Pre-DWH	-1.33	0.45	-2.94	0.02*
	<i>Serranus phoebe</i>	2010 - Pre-DWH	-0.91	0.56	-1.65	0.39
		2011 - Pre-DWH	-1.2	0.38	-3.15	0.01*
		2012 - Pre-DWH	-0.26	0.39	-0.68	0.97
		2013 - Pre-DWH	-1.07	0.39	-2.74	0.03*
		2014 - Pre-DWH	-1.1	0.4	-2.78	0.03*
		2015 - Pre-DWH	-1.03	0.41	-2.48	0.07
		2016 - Pre-DWH	-1.36	0.47	-2.91	0.02*
		2017 - Pre-DWH	-1.41	0.44	-3.23	<0.01*
	<i>Serranus subligarius</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.3	0.35	-0.88	0.95
		2012 - Pre-DWH	0.29	0.35	0.83	0.96
		2013 - Pre-DWH	-0.4	0.46	-0.88	0.95
		2014 - Pre-DWH	0.09	0.69	0.13	1
		2015 - Pre-DWH	-1.16	0.52	-2.21	0.16
		2016 - Pre-DWH	-0.29	0.69	-0.42	1
		2017 - Pre-DWH	-0.29	0.69	-0.42	1

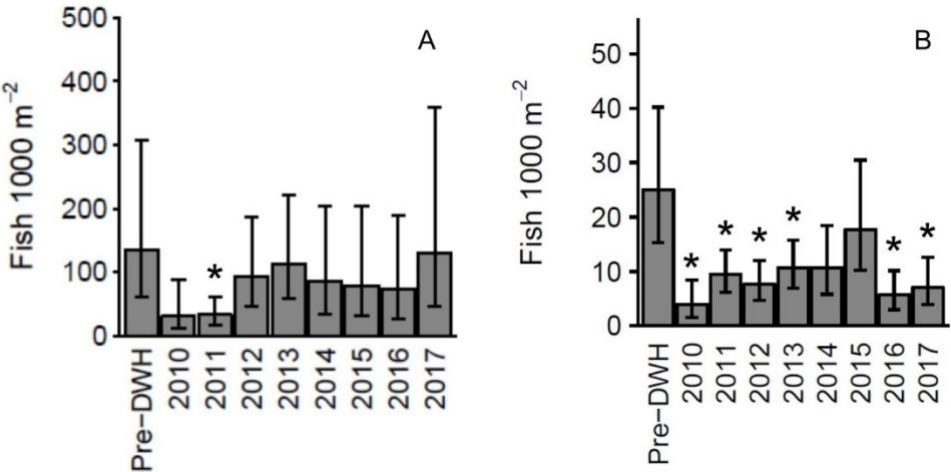
Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Piscivore	<i>Carcharhinus plumbeus</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	0.09	0.36	0.25	1
		2012 - Pre-DWH	-0.27	0.25	-1.09	0.85
		2013 - Pre-DWH	-0.18	0.28	-0.65	0.99
		2014 - Pre-DWH	-0.99	0.36	-2.73	0.04*
		2015 - Pre-DWH	-0.75	0.32	-2.31	0.12
		2016 - Pre-DWH	-0.75	0.24	-3.07	0.01*
		2017 - Pre-DWH	-1.06	0.36	-2.92	0.02*
	<i>Mycteroperca microlepis</i>	2010 - Pre-DWH	-1.11	0.39	-2.82	0.03*
		2011 - Pre-DWH	-0.83	0.28	-2.94	0.02*
		2012 - Pre-DWH	-0.83	0.25	-3.34	<0.01*
		2013 - Pre-DWH	-0.81	0.35	-2.35	0.12
		2014 - Pre-DWH	-1.03	0.26	-4.02	<0.01*
		2015 - Pre-DWH	-1.16	0.31	-3.74	<0.01*
		2016 - Pre-DWH	-0.61	0.34	-1.79	0.39
		2017 - Pre-DWH	-1.45	0.51	-2.86	0.03*
	<i>Mycteroperca phenax</i>	2010 - Pre-DWH	-0.97	0.26	-3.72	<0.01*
		2011 - Pre-DWH	-0.47	0.19	-2.44	0.09
		2012 - Pre-DWH	-0.65	0.2	-3.22	<0.01*
		2013 - Pre-DWH	-0.67	0.2	-3.38	<0.01*
		2014 - Pre-DWH	-1.04	0.24	-4.31	<0.01*
		2015 - Pre-DWH	-0.7	0.25	-2.74	0.04*
		2016 - Pre-DWH	-1.13	0.29	-3.95	<0.01*
		2017 - Pre-DWH	-1	0.26	-3.83	<0.01*
Reef Planktivore	<i>Apogon pseudomaculatus</i>	2010 - Pre-DWH	-1.2	0.78	-1.54	0.46
		2011 - Pre-DWH	-0.41	0.52	-0.8	0.94
		2012 - Pre-DWH	-1.72	0.55	-3.11	0.01*
		2013 - Pre-DWH	-0.02	0.56	-0.04	1
		2014 - Pre-DWH	-1.27	0.56	-2.28	0.11
		2015 - Pre-DWH	-0.96	0.58	-1.66	0.38
		2016 - Pre-DWH	0.03	0.88	0.03	1
		2017 - Pre-DWH	--	--	--	--

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Reef Planktivore	<i>Baldwinella aureorubens</i>	2012 - 2011	--	--	--	--
		2013 - 2011	1.36	0.87	1.57	0.39
		2014 - 2011	1.58	2.04	0.78	0.91
		2015 - 2011	4.77	0.93	5.14	<0.01*
		2016 - 2011	1.78	1.46	1.22	0.63
		2017 - 2011	0.22	0.96	0.23	1
	<i>Chromis enchrysur</i>	2010 - Pre-DWH	-0.32	0.79	-0.4	1
		2011 - Pre-DWH	-0.44	0.36	-1.2	0.79
		2012 - Pre-DWH	-0.65	0.39	-1.67	0.44
		2013 - Pre-DWH	0.27	0.35	0.77	0.97
		2014 - Pre-DWH	-0.46	0.4	-1.16	0.81
		2015 - Pre-DWH	0.48	0.41	1.16	0.81
		2016 - Pre-DWH	-0.67	0.45	-1.49	0.57
		2017 - Pre-DWH	-1.22	0.51	-2.37	0.11
	<i>Paranthias furcifer</i>	2010 - Pre-DWH	-1.68	0.76	-2.23	0.13
		2011 - Pre-DWH	0.45	0.68	0.67	0.97
		2012 - Pre-DWH	-0.27	0.64	-0.42	1
		2013 - Pre-DWH	-0.82	1.04	-0.79	0.94
		2014 - Pre-DWH	-1.2	0.68	-1.75	0.34
		2015 - Pre-DWH	1.58	1.02	1.54	0.47
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	--	--	--	--
	<i>Priacanthus arenatus</i>	2010 - Pre-DWH	-1.04	0.35	-2.96	0.02*
		2011 - Pre-DWH	-0.37	0.26	-1.42	0.6
		2012 - Pre-DWH	-0.12	0.27	-0.45	1
		2013 - Pre-DWH	-0.29	0.25	-1.14	0.8
		2014 - Pre-DWH	-0.38	0.31	-1.2	0.76
		2015 - Pre-DWH	0.67	0.8	0.83	0.95
		2016 - Pre-DWH	--	--	--	--
		2017 - Pre-DWH	-0.04	0.49	-0.08	1

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Reef Planktivore	<i>Pristigenys alta</i>	2010 - Pre-DWH	--	--	--	--
		2011 - Pre-DWH	-0.21	0.59	-0.36	1
		2012 - Pre-DWH	--	--	--	--
		2013 - Pre-DWH	0.47	0.59	0.79	0.92
		2014 - Pre-DWH	-0.01	0.46	-0.02	1
		2015 - Pre-DWH	-0.26	0.37	-0.7	0.95
		2016 - Pre-DWH	0.18	0.38	0.47	0.99
		2017 - Pre-DWH	-0.02	0.39	-0.06	1
	<i>Pronotogrammus martinicensis</i>	2010 - Pre-DWH	0.07	0.74	0.1	1
		2011 - Pre-DWH	0.58	0.57	1.01	0.87
		2012 - Pre-DWH	0.41	0.61	0.66	0.99
		2013 - Pre-DWH	0.56	0.55	1.01	0.87
		2014 - Pre-DWH	0.38	0.62	0.61	0.99
		2015 - Pre-DWH	0.67	0.65	1.04	0.86
		2016 - Pre-DWH	0.13	0.76	0.17	1
		2017 - Pre-DWH	-0.58	0.77	-0.76	0.97
	<i>Rhomboplites aurorubens</i>	2010 - Pre-DWH	-0.53	0.6	-0.88	0.94
		2011 - Pre-DWH	-0.93	0.48	-1.94	0.26
		2012 - Pre-DWH	-1.52	0.48	-3.15	0.01*
		2013 - Pre-DWH	-0.02	0.47	-0.04	1
		2014 - Pre-DWH	-0.72	0.55	-1.3	0.69
		2015 - Pre-DWH	0.11	0.56	0.2	1
		2016 - Pre-DWH	-0.34	0.56	-0.61	0.99
		2017 - Pre-DWH	0.25	0.59	0.43	1
	<i>Stegastes partitus</i>	2010 - Pre-DWH	-1.73	0.63	-2.75	0.02*
		2011 - Pre-DWH	--	--	--	--
		2012 - Pre-DWH	--	--	--	--
		2013 - Pre-DWH	-1.01	0.86	-1.18	0.58
		2014 - Pre-DWH	-0.72	0.86	-0.84	0.82
		2015 - Pre-DWH	-0.9	0.93	-0.97	0.72
		2016 - Pre-DWH	-1.62	1.04	-1.56	0.34
		2017 - Pre-DWH	--	--	--	--

Trophic Guild	Scientific Name	Comparison	Estimate	SE	z	P
Reef Planktivore	Damselfish	2010 - Pre-DWH	-3.08	1.18	-2.6	0.06
		2011 - Pre-DWH	-2.12	0.75	-2.85	0.03*
		2012 - Pre-DWH	-0.58	0.92	-0.63	0.99
		2013 - Pre-DWH	0.14	0.63	0.22	1
		2014 - Pre-DWH	0.9	0.76	1.18	0.79
		2015 - Pre-DWH	0.56	0.7	0.8	0.96
		2016 - Pre-DWH	-0.43	1.01	-0.42	1
		2017 - Pre-DWH	0.36	0.78	0.46	1

92 **Figure S1.** Standardized trophic guild density ($\pm 95\%$ CIs) estimates for small demersal
93 invertivores with (A) and without (B) tomtate. An asterisk (*) denotes a significant difference for
94 the log-normal model (Table S4).



The following appendix accompanies the article

Changes in Reef Fish Community Structure Following the Deepwater Horizon Oil Spill

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Appendix A.

Trophic Guild	Family	Scientific Name	Common Name	Reference
Herbivore	Acanthuridae	<i>Acanthurus chirurgus</i>	Doctorfish	1,2
	Pomacanthidae	<i>Centropyge argi</i>	Cherubfish	2
	Pomacentridae	<i>Microspathodon chrysurus</i>	Yellowtail Damselfish	3
Small Demersal Browser	Blenniidae	<i>Parablennius marmoreus</i>	Seaweed Blenny	2
	Chaetodontidae	<i>Chaetodon ocellatus</i>	Spotfin Butterflyfish	4,5
	Chaetodontidae	<i>Chaetodon sedentarius</i>	Reef Butterflyfish	2,6
	Chaetodontidae	<i>Prognathodes aya</i>	Bank Butterflyfish	6
	Monocanthidae	<i>Stephanolepis hispidus</i>	Planehead Filefish	
	Monocanthidae	<i>Stephanolepis setifer</i>	Pygmy Filefish	7
	Pomacentridae	<i>Stegastes leucostictus</i>	Beaugregory	2,3
	Pomacentridae	<i>Stegastes variabilis</i>	Cocoa Damselfish	3,8
	Tetraodontidae	<i>Canthigaster rostrata</i>	Sharpnose Puffer	2
	Tetraodontidae	<i>Sphoeroides parvus</i>	Least Puffer	
Large Demersal Browser	Ephippidae	<i>Chaetodipterus faber</i>	Spadefish	2,9
	Monocanthidae	<i>Aluterus monoceros</i>	Unicorn Filefish	
	Monocanthidae	<i>Aluterus schoepfii</i>	Orange Filefish	2,10
	Monocanthidae	<i>Aluterus scriptus</i>	Scrawled Filefish	2
	Pomacanthidae	<i>Holacanthus bermudensis</i>	Blue Angelfish	2
	Pomacanthidae	<i>Holacanthus ciliaris</i>	Queen Angelfish	2
	Pomacanthidae	<i>Holacanthus tricolor</i>	Rock Beauty Angelfish	2,11
	Pomacanthidae	<i>Pomacanthus arcuatus</i>	Gray Angelfish	2,11
	Pomacanthidae	<i>Pomacanthus paru</i>	French Angelfish	2,11
Small Demersal Invertivore	Haemulidae	<i>Anisotremus virginicus</i>	Porkfish	2
	Haemulidae	<i>Haemulon aurolineatum</i>	Tomtate	2,12
	Holocentridae	<i>Corniger spinosus</i>	Spinycheek Soldierfish	6
	Holocentridae	<i>Holocentrus adscensionis</i>	Squirrelfish	2
	Holocentridae	<i>Holocentrus rufus</i>	Longspine Squirrelfish	2
	Holocentridae	<i>Plectrypops retrospinis</i>	Cardinal Soldierfish	13
	Labridae	<i>Bodianus pulchellus</i>	Spotfin Hogfish	
	Labridae	<i>Bodianus rufus</i>	Spanish Hogfish	2
	Labridae	<i>Halichoeres bathyphilus</i>	Greenband Wrasse	
	Labridae	<i>Halichoeres bivittatus</i>	Slippery Dick	2,14
	Labridae	<i>Thalassoma bifasciatum</i>	Bluehead Wrasse	14
	Labridae	<i>Xyrichtys novacula</i>	Pearly Razorfish	
	Microdesmidae	<i>Ptereleotris calliura</i>	Blue Goby	
	Sciaenidae	<i>Equetus lanceolatus</i>	Jackknife fish	2,15,16
	Sciaenidae	<i>Equetus punctatus</i>	Spotted Drum	2
	Sciaenidae	<i>Micropogonias undulatus</i>	Atlantic Croaker	17
	Sciaenidae	<i>Pareques acuminatus</i>	Highhat	2
	Sciaenidae	<i>Pareques iwamotoi</i>	Blackbar Drum	6
	Sciaenidae	<i>Pareques umbrosus</i>	Cubby	6,16
	Triglidae	<i>Bellator militaris</i>	Horned Searobin	

Appendix A continued.

Trophic Guild	Family	Scientific Name	Common Name	Reference
Large Demersal Invertivore	Balistidae	<i>Balistes capriscus</i>	Gray Triggerfish	18
	Carangidae	<i>Alectis ciliaris</i>	African Pompano	19
	Dasyatidae	<i>Dasyatis americana</i>	Southern Stingray	20
	Dasyatidae	<i>Dasyatis centroura</i>	Roughtail Stingray	21
	Dasyatidae	<i>Dasyatis sabina</i>	Atlantic Stingray	22
	Dasyatidae	<i>Urobatis jamaicensis</i>	Yellow Stingray	23
	Diodontidae	<i>Chilomycterus schoepfi</i>	Striped Burrfish	24
	Ostraciidae	<i>Acanthostracion quadricornis</i>	Scrawled Cowfish	2
	Rajidae	<i>Raja ackleyi</i>	Ocellate Skate	
	Sciaenidae	<i>Pogonias cromis</i>	Black Drum	25
	Serranidae	<i>Alphestes afer</i>	Mutton Hamlet	2
	Sparidae	<i>Archosargus probatocephalus</i>	Sheepshead	2,26
	Sparidae	<i>Calamus bajonado</i>	Jolthead Porgy	2
	Sparidae	<i>Calamus calamus</i>	Saucereye Porgy	2
	Sparidae	<i>Calamus leucosteus</i>	Whitebone Porgy	27
	Sparidae	<i>Calamus nodosus</i>	Knobbed Porgy	28
	Sparidae	<i>Calamus proridens</i>	Littlehead Porgy	29
	Sparidae	<i>Pagrus pagrus</i>	Red Porgy	6
	Sparidae	<i>Stenotomus caprinus</i>	Longspine Porgy	27
	Squalidae	<i>Mustelus canis</i>	Smooth Dogfish	30
	Squalidae	<i>Mustelus norrisi</i>	Florida Smoothhound	
Generalist Carnivore	Batrachoididae	<i>Opsanus beta</i>	Gulf toadfish	31
	Batrachoididae	<i>Opsanus pardus</i>	Leopard Toadfish	6
	Carangidae	<i>Caranx crysos</i>	Blue Runner	2,32,33
	Carangidae	<i>Elagatis bipinnulata</i>	Rainbow runner	34
	Carangidae	<i>Seriola dumerili</i>	Greater Amberjack	2,19,27,35
	Carangidae	<i>Seriola rivoliana</i>	Almaco Jack	35
	Echeneidae	<i>Echeneis naucrates</i>	Sharksucker	2
	Echeneidae	<i>Remora remora</i>	Remora	2
	Haemulidae	<i>Orthopristis chrysoptera</i>	Pigfish	36
	Lutjanidae	<i>Lutjanus campechanus</i>	Red Snapper	37,38
	Lutjanidae	<i>Lutjanus griseus</i>	Gray Snapper	39
	Lutjanidae	<i>Lutjanus synagris</i>	Lane Snapper	2,39
	Malacanthidae	<i>Malacanthus plumieri</i>	Sand Tilefish	2,6
	Muraenidae	<i>Echidna catenata</i>	Chain Moray Eel	2,40
	Muraenidae	<i>Gymnothorax funebris</i>	Green Moray Eel	19
	Muraenidae	<i>Gymnothorax moringa</i>	Spotted Moray Eel	2,41,42
	Ogcocephalidae	<i>Ogcocephalus corniger</i>	Longnose Batfish	
	Paralichthyidae	<i>Paralichthys albigutta</i>	Gulf Flounder	43
	Rachycentridae	<i>Rachycentron canadum</i>	Cobia	44
	Sciaenidae	<i>Sciaenops ocellatus</i>	Red Drum	45,46
	Scorpaenidae	<i>Pterois volitans</i>	Red Lionfish	47,48
	Scorpaenidae	<i>Scorpaena brasiliensis</i>	Barbfish	2
	Scorpaenidae	<i>Scorpaena plumieri</i>	Spotted Scorpionfish	2
	Serranidae	<i>Centropristis ocyurus</i>	Bank Sea Bass	6,49
	Serranidae	<i>Diplectrum formosum</i>	Sand Perch	50,51

Appendix A continued.

Trophic Guild	Family	Scientific Name	Common Name	Reference
Generalist Carnivore	Serranidae	<i>Epinephelus adscensionis</i>	Rock Hind	2,49
	Serranidae	<i>Epinephelus cruentatus</i>	Graysby	2
	Serranidae	<i>Epinephelus guttatus</i>	Red Hind	2,52,53
	Serranidae	<i>Epinephelus morio</i>	Red Grouper	2,49
	Serranidae	<i>Gonioplectrus hispanus</i>	Spanish Flag	
	Serranidae	<i>Hypoplectrus indigo</i>	Indigo Hamlet	49
	Serranidae	<i>Hyporthodus niveatus</i>	Snowy Grouper	54
	Serranidae	<i>Liopropoma eukrines</i>	Wrasse Bass	6
	Serranidae	<i>Rypticus maculatus</i>	Whitespotted Soapfish	49
	Serranidae	<i>Serranus phoebe</i>	Tattler	49
	Serranidae	<i>Serranus subligarius</i>	Belted Sandfish	49,55
Piscivore	Antennariidae	<i>Fowlerichthys ocellatus</i>	Ocellated Frogfish	
	Belonidae	<i>Tylosurus crocodilus</i>	Houndfish	2
	Carangidae	<i>Seriola fasciata</i>	Lesser Amberjack	
	Carangidae	<i>Seriola zonata</i>	Banded Rudderfish	
	Carcharhinidae	<i>Carcharhinus falciformis</i>	Silky Shark	56
	Carcharhinidae	<i>Carcharhinus leucas</i>	Bull Shark	57
	Carcharhinidae	<i>Carcharhinus limbatus</i>	Blacktip	58
	Carcharhinidae	<i>Carcharhinus obscurus</i>	Dusky Shark	30
	Carcharhinidae	<i>Carcharhinus plumbeus</i>	Sandbar shark	59
	Carcharhinidae	<i>Rhizoprionodon terraenovae</i>	Atlantic Sharpnose Shark	60–62
	Fistulariidae	<i>Fistularia petimba</i>	Red Cornetfish	
	Muraenidae	<i>Muraena retifera</i>	Reticulate Moray	63
	Paralichthyidae	<i>Paralichthys lethostigma</i>	Southern Flounder	64
	Pomatomidae	<i>Pomatomus saltatrix</i>	Bluefish	65
	Scombridae	<i>Scomberomorus cavalla</i>	King Mackerel	2
	Scombridae	<i>Scomberomorus maculatus</i>	Spanish Mackerel	
	Serranidae	<i>Mycteroperca microlepis</i>	Gag	49,66
	Serranidae	<i>Mycteroperca phenax</i>	Scamp	67
	Sphyrnaeidae	<i>Sphyrna barracuda</i>	Great Barracuda	2,68
	Synodontidae	<i>Synodus intermedius</i>	Sand Diver	2
Reef Planktivore	Apogonidae	<i>Apogon pseudomaculatus</i>	Twospot Cardinalfish	6
	Apogonidae	<i>Paroncheilus affinis</i>	Bigtooth Cardinalfish	
	Holocentridae	<i>Myripristis jacobus</i>	Blackbar Soldierfish	2,19
	Lutjanidae	<i>Rhomboplites aurorubens</i>	Vermilion Snapper	69
	Pomacentridae	<i>Chromis cyanea</i>	Blue Chromis	2,3
	Pomacentridae	<i>Chromis enchrysur</i>	Yellowtail Reeffish	3
	Pomacentridae	<i>Stegastes partitus</i>	Bicolor Damselfish	3
	Pomacentridae	Damselfish	Purple reeffish/dusky damselfish	2,3,70

Appendix A continued.

Trophic Guild	Family	Scientific Name	Common Name	Reference
Reef Planktivore	Priacanthidae	<i>Priacanthus arenatus</i>	Atlantic Bigeye	2,6,27,71
	Priacanthidae	<i>Pristigenys alta</i>	Short Bigeye	6
	Serranidae	<i>Baldwinella aureorubens</i>	Streamer Bass	
	Serranidae	<i>Baldwinella vivanus</i>	Red Barbier	49,72
	Serranidae	<i>Hemanthias leptus</i>	Longtail Bass	49
	Serranidae	<i>Paranthias furcifer</i>	Creolefish	2,49,73
	Serranidae	<i>Pronotogrammus martinicensis</i>	Roughtongue Bass	49
Pelagic Planktivore	Carangidae	<i>Decapterus macarellus</i>	Mackerel Scad	2
	Carangidae	<i>Selar crumenophthalmus</i>	Bigeye Scad	2
	Carangidae	<i>Selene setapinnis</i>	Atlantic Moonfish	
	Myliobatidae	<i>Manta birostris</i>	Manta Ray	

References

1. Dias, T. L. P., Rosa, I. L. & Feitoza, B. M. Food resource and habitat sharing by the three western South Atlantic surgeonfishes (Teleostei: Acanthuridae: Acanthurus) off Paraíba coast, north-eastern Brazil. *aqua, Int. J. Ichthyol.* **5**, 1–10 (2001).
2. Randall, J. E. Food habits of reef fishes of the West Indies. *Hawaii Inst. Mar. Biol.* 665–847 (1967).
3. Emery, A. R. Comparative ecology and functional osteology of fourteen species of damselfish (Pisces: Pomacentridae) at Alligator Reef, Florida Keys. *Bull. Mar. Sci.* **23**, 649–770 (1973).
4. Motta, P. J. Dentition patterns among Pacific and Western Atlantic butterflyfishes (Perciformes, Chaetodontidae): relationship to feeding ecology and evolutionary history. *Environ. Biol. Fishes* **25**, 159–170 (1989).
5. Aiken, K. *The Biology, Ecology, and Bionomics of Butterfly and Angelfishes, Chaetodontidae. Caribbean Coral Reef Fishery Resources* (International Center for Living Aquatic Resources Management, 1983).
6. Weaver, D. C., Dennis, G. D. & Sulak, K. J. *Community Structure and Trophic Ecology of Fishes on the Pinnacles Reef Tract. Final Synthesis Report* (2002).
7. Clements, W. H. & Livingston, R. J. Overlap and pollution-induced variability in the feeding habits of filefish (Pisces: Monacanthidae) from Apalachee Bay, Florida. *Copeia* **1983**, 331 (1983).
8. Nelson, B. D. & Bortone, S. A. Feeding guilds among artificial-reef fishes Northern Gulf of Mexico. *Gulf Mex. Sci.* **2**, 66–80 (1996).
9. Hayse, J. W. Feeding habits, age, growth, and reproduction of Atlantic spadefish *Chaetodipterus faber* (Pisces: Ephippidae) in South Carolina. *Fish. Bull.* **88**, 67–83 (1990).
10. Cargo, D. G. & Schultz, L. P. Notes on the biology of the sea nettle, *Chrysaora quinquecirrha*, in Chesapeake Bay. *Chesap. Sci.* **7**, 95 (1966).
11. Hourigan, T. F., Stanton, F. G., Motta, P. J., Kelley, C. D. & Carlson, B. The feeding ecology of three species of Caribbean angelfishes (family Pomacanthidae). *Environ. Biol. Fishes* **24**, 105–116 (1989).
12. Norberg, M. J. The ecology of tomtate, *Haemulon aurolineatum*, in the northern Gulf of Mexico and effects of the Deepwater Horizon oil spill. (University of South Alabama, 2015).
13. Gladfelter, W. B. & Johnson, W. S. Feeding niche separation in a guild of tropical reef fishes (Holocentridae). *Ecology* **64**, 552–563 (1983).
14. Clifton, K. B. & Motta, P. J. Feeding morphology, diet, and ecomorphological relationships among five Caribbean labrids (Teleostei, Labridae). *Copeia* **1998**, 953 (1998).

15. Lowe, R. H. The sciaenid fishes of Bristh Guiana. *Bull. Mar. Sci.* **16**, 26–57 (1966).
16. Darovev, J. E. *Sciaenid fishes (Osteichthyes: Perciformes) of western peninsular Florida. Memoirs of the Hourglass Cruises* **6**, (1983).
17. Overstreet, R. M. & Heard, R. W. Food of the Atlantic croaker, *Micropogonias undulatus*, from Mississippi Sound and the Gulf of Mexico. *Gulf Res. Reports* **6**, (1978).
18. Vose, F. E. & Nelson, W. G. Gray triggerfish (*Balistes capriscus* Gmelin) feeding from artifical and natural substrate in shallow Atlantic waters of Florida. *Bull. Mar. Sci.* **55**, 1316–1323 (1994).
19. Bohnsack, J. A., Harper, D. E., McClellan, D. B., Sutherland, D. L. & White, M. W. Resource survey of fishes within Looe Key National Marine Sanctuary. *NOAA Tech. Memo.* 1–108 (1987).
20. Gilliam, D. & Sullivan, K. M. Diet and feeding habits of the southern stingray *Dasyatis americana* in the central Bahamas. *Bull. Mar. Sci.* **52**, 1007–1013 (1993).
21. Struhsaker, P. Observations on the biology and distribution of the thorny stingray, *Dasyatis centroura* (Pisces: Dasyatidae). *Bull. Mar. Sci.* **19**, 456–481 (1969).
22. Snelson, F. F. & Williams, S. E. Notes on the occurrence, distribution, and biology of elasmobranch fishes in the Indian River Lagoon system, Florida. *Estuaries* **4**, 110 (1981).
23. O'Shea, O. R., Wueringer, B. E., Winchester, M. M. & Brooks, E. J. Comparative feeding ecology of the yellow ray *Urolophus hannah* (Urolophidae) from the bahamas. *J. Fish Biol.* **92**, 73–84 (2018).
24. Motta, P. J. *et al.* Feeding relationships among nine species of seagrass fishes of Tampa Bay, Florida. *Bull. Mar. Sci.* **56**, 185–200 (1995).
25. Overstreet, R. M. & Heard, R. W. Food contents of six commercial fishes from Mississippi Sound. *Gulf Res. Reports* **7**, (1982).
26. Sedberry, G. R. Feeding habits of sheepshead, *Archosargus probatocephalus*, in offshore reef habitats of the southeastern continental shelf. *Northeast Gulf Sci.* **9**, (1987).
27. Bowman, R. E., Stillwell, C. E., Michaels, W. L. & Grosslein, M. D. National Oceanic and Atmospheric Administration National Marine Fisheries Service Food of Northwest Atlantic Fishes and Two Common Species of Squid. *NOAA Tech. Memo. NMFS-NE-155* (2000).
28. Horvath, M. L., Grimes, C. B., Huntsman, G. R. & Carolina, S. Growth, mortality, reproduction and feeding of knobbed porgy, *Calamus nodosus*, along the southeastern United States coast. *Bull. Mar. Sci.* **46**, 677–687 (1990).
29. Darcy, G. H. *Synopsis of biological data on the porgies, Calamus arctifrons and C. proridens (Pisces: Sparidae).* (1986).
30. Gelsleichter, J., Musick, J. A. & Nichols, S. Food habits of the smooth dogfish, *Mustelus canis*, dusky shark, *Carcharhinus obscurus*, Atlantic sharpnose shark, *Rhizoprionodon terraenovae*, and the sand tiger, *Carcharias taurus*, from the northwest Atlantic Ocean.

Environ. Biol. Fishes **54**, 205–217 (1999).

31. Springer, V. G. & Woodburn, K. D. *An ecological study of the fishes of the Tampa Bay area*. Florida State Board of Conservation (1960).
32. Keenan, S. F. The importance of zooplankton in the diets of the blue runner (*Caranx crysos*) near offshore petroleum platforms in the northern Gulf of Mexico. 166 (1996).
33. Sley, A., Jarboui, O., Ghorbel, M. & Bouain, A. Food and feeding habits of *Caranx crysos* from the Gulf of Gabs (Tunisia). *J. Mar. Biol. Assoc. United Kingdom* **89**, 1375–1380 (2009).
34. Ménard, F. *et al.* Pelagic cephalopods in the western Indian Ocean: New information from diets of top predators. *Deep. Res. Part II Top. Stud. Oceanogr.* **95**, 83–92 (2013).
35. Manooch, C. S. I. & Haimovici, M. Foods of greater amberjack, *Seriola dumerili*, and Almaco Jack, *Seriola rivoliana* (Pisces: Carangidae), from the South Atlantic Bight. *The Journal of the Elisha Mitchell Scientific Society* **99**, 1–9 (1983).
36. Howe, J. C. Diet Composition of Juvenile Pigfish, *Orthopristis chrysoptera* (Perciformes: Haemulidae), from the Northern Gulf of Mexico. *Gulf Mex. Sci.* **19**, 55–60 (2001).
37. Wells, R. J. D., Cowan, J. H. & Fry, B. Feeding ecology of red snapper *Lutjanus campechanus* in the northern Gulf of Mexico. *Mar. Ecol. Prog. Ser.* **361**, 213–225 (2008).
38. Tarnecki, J. H. & Patterson, W. F. Changes in red snapper diet and trophic ecology following the Deepwater Horizon oil spill. *Mar. Coast. Fish.* **7**, 135–147 (2015).
39. Franks, J. S. & VanderKooy, K. E. Feeding habits of juvenile lane snapper *Lutjanus synagris* from Mississippi coastal waters, with comments on the diet of gray snapper *Lutjanus griseus*. *Gulf Caribb. Res.* **12**, 11–17 (2000).
40. Mehta, R. S. Ecomorphology of the moray bite: relationship between dietary extremes and morphological diversity. *Physiol. Biochem. Zool.* **82**, 90–103 (2009).
41. Zokan, M. A. The life history of morays (Anguilliformes: Muraenidae) off the southeastern Atlantic coast of the United States. *Masters Abstr. Int. Vol. 46, no. 05, 118 p.* 2008. 1–118 (2008).
42. Young, R. F. & Winn, H. E. Activity patterns , diet , and shelter site use for two species of moray eels , *Gymnothorax moringa* and *Gymnothorax vicinus*, in Belize. *Am. Soc. Ichthyol. Herpetol.* **2003**, 44–55 (2003).
43. Peebles, E. B. & Hopkins, T. L. *Feeding habits of eight fish species from Tampa Bay, with observations on opportunistic predation.* (1993).
44. Meyer, G. H. & Franks, J. S. Food of Cobia, *Rachycentron canadum*, from the northcentral Gulf of Mexico. *Gulf Res. Reports* **9**, 161–167 (1996).
45. Scharf, F. S. & Schlight, K. K. Feeding habits of red drum (*Sciaenops ocellatus*) in Galveston Bay, Texas: seasonal diet variation and predator-prey size relationships. *Estuaries* **23**, 128 (2000).

46. Overstreet, R. M. & Heard, R. W. Food of the red drum, *Sciaenops ocellata*, from Mississippi Sound. *Gulf Res. Reports* **6**, 131–135 (1978).
47. Dahl, K. A., Patterson III, W. F., Robertson, A. & Ortmann, A. C. DNA barcoding significantly improves resolution of invasive lionfish diet in the northern Gulf of Mexico. *Biol. Invasions* **19**, 1917–1933 (2017).
48. Dahl, K. A. & Patterson III, W. F. Habitat-specific density and diet of rapidly expanding invasive red lionfish, *Pterois volitans*, populations in the northern Gulf of Mexico. *PLoS One* **9**, e105852 (2014).
49. Bullock, L. H. & Smith, G. B. *Seabasses (Pisces: Serranidae)*. (1991).
50. Bortone. *Studies on the biology of the sand perch Diplectrum formosum, Serranidae*. (1971).
51. Darcy, G. H. *Synopsis of Biological Data on the Sand Perch, Diplectrum formosum (Pisces: Serranidae)*. (1985).
52. Menzel, D. W. Utilization of food by a Bermuda reef fish, *Epinephelus guttatus*. *ICES J. Mar. Sci.* **25**, 216–222 (1960).
53. Thompson, R. & Munro, J. L. Aspects of the biology and ecology of Caribbean reef fishes: Serranidae (hinds and groupers). *J. Fish Biol.* **12**, 115–146 (1978).
54. Bielsa, L. M. & Labisky, R. F. Food habits of blueline tilefish, *Caulolatilus microps*, and snowy grouper, *Epinephelus niveatus*, from the lower Florida Keys. *Northeast Gulf Sci.* **9**, (1987).
55. Hastings, P. A. & Bortone, S. A. Observations on the life history of the belted sandfish, *Serranus subligarius* (Serranidae). *Environ. Biol. Fishes* **5**, 365–374 (1980).
56. Bonfil, R. S. The biology and ecology of the silky shark, *Carcharhinus falciformis*. in *Sharks of the open ocean: biology, fisheries and conservation* (eds. Camhi, M. D., Pikitch, E. K. & Babcock, E. A.) 114–127 (Blackwell Publishing, 2008).
57. Snelson, F. F., Mulligan, T. J. & Williams, S. E. Food habits, occurrence, and population structure of the bull shark, *Carcharhinus leucas*, in Florida coastal lagoons. *Bull. Mar. Sci.* **34**, 71–80 (1984).
58. Castro, J. Biology of the blacktip shark, *Carcharhinus limbatus*, off the southeastern United States. *Bull. Mar. Sci.* **59**, 508–522 (1996).
59. Stillwell, C. E. & Kohler, N. E. Food habits of the sandbar shark *Carcharhinus plumbeus* off the U.S. northeast coast, with estimates of daily ration. *Fish. Bulletin* **91**, 138–150 (1992).
60. Bethea, D. M., Carlson, J. K., Buckel, J. A. & Satterwhite, M. Ontogenetic and site-related trends in the diet of the Atlantic sharpnose shark *Rhizoprionodon terraenovae* from the northeast Gulf of Mexico. *Bull. Mar. Sci.* **78**, 287–307 (2006).
61. Drymon, J. M., Powers, S. P. & Carmichael, R. H. Trophic plasticity in the Atlantic sharpnose shark (*Rhizoprionodon terraenovae*) from the north central Gulf of Mexico.

- Environ. Biol. Fishes* **95**, 21–35 (2012).
62. Delorenzo, D. M., Bethea, D. M. & Carlson, J. K. An assessment of the diet and trophic level of Atlantic sharpnose shark *Rhizoprionodon terraenovae*. *J. Fish Biol.* **86**, 385–391 (2015).
 63. Randall, J. E. *Guide to Hawaiian Reef Fishes*. (1985).
 64. Powell, A. B. & Schwartz, F. J. Food of *Paralichthys dentatus* and *P. lethostigma* (Pisces: Bothidae) in North Carolina Estuaries. *Estuaries* **2**, 276 (1979).
 65. Harding, J. M. & Mann, R. Diet and habitat use by bluefish, *Pomatomus saltatrix*, in a Chesapeake Bay estuary. *Environ. Biol. Fishes* **60**, 401–409 (2001).
 66. Naughton, S. P. & Saloman, C. H. Food of gag (*Mycteroperca microlepis*) from North Carolina and three areas of Florida. (1985). doi:10.1360/zd-2013-43-6-1064
 67. Matheson, R. H., Huntsman, G. R. & Manooch, C. S. Age, growth, mortality, food and reproduction of the scamp, *Mycteroperca phenax*, collected off North Carolina and South Carolina. *Bull. Mar. Sci.* **38**, 300–312 (1986).
 68. Schmidt, T. W. Food habits, length-weight relationship and condition factor of young great barracuda, *Syphraena barracuda* (Walbaum), from Florida Bay, Everglades National Park, Florida. *Bull. Mar. Sci.* **44**, 163–170 (1989).
 69. Grimes, C. B. Diet and feeding ecology of the vermilion snapper, *Rhomboplites aurorubens* (Cuvier) from North Carolina and South Carolina waters. *Bull. Mar. Sci.* **29**, 53–61 (1979).
 70. Feitosa, J. L. L., Concentino, A. M., Teixeira, S. F. & Ferreira, B. P. Food resource use by two territorial damselfish (Pomacentridae: Stegastes) on South-Western Atlantic algal-dominated reefs. *J. Sea Res.* **70**, 42–49 (2012).
 71. Cardozo, A. L. P. *et al.* Feeding ecology and ingestion of plastic fragments by *Priacanthus arenatus*: What's the fisheries contribution to the problem? *Mar. Pollut. Bull.* **130**, 19–27 (2018).
 72. Lindquist, D. & Clavijo, I. Quantifying deep reef fishes from a submersible and notes on a live collection and diet of the red barbier, *Hemanthias vivanus*. *J. Elisha Mitchell Sci. Soc.* **3**, 135–140 (1993).
 73. Nelson, R. S. The life history, ecology, and population dynamics of four sympatric reef predators (*Rhomboplites aurorubens*, *Lutjanus campechanus*, Lutjanidae; *Haemulon melanurum*, Haemulidae; and *Pagrus pagrus*, Sparidae) on the east and west F. (North Carolina State University, 1988).