# Standardized Catch Rate Indices for Vermilion Snapper (Rhomboplites aurorubens) during 1986-2017 by the U.S. Gulf of Mexico Charterboat and Private Boat Recreational Fishery 

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# Standardized Catch Rate Indices for Vermilion Snapper (Rhomboplites aurorubens) during 1986-2017 by the U.S. Gulf of Mexico Charterboat and Private Boat Recreational Fishery 

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## Keywords

CPUE, catch, effort, recreational fisheries, Vermilion Snapper


#### Abstract

A delta-lognormal index was constructed for the SEDAR67 Standard Vermilion Snapper stock assessment. The index uses data from the Marine Recreational Information Program, which underwent a substantial modification and peer-review in 2018 following a three year transition period (2015-2018). An index for the Eastern U.S. Gulf of Mexico is developed following the same methodology and approach used for SEDAR45 and SEDAR09. The resulting index reveals relatively similar index trends when compared to the SEDAR45 index. The SEDAR67 standardized index indicates catch rates were relatively high from 1990-1995, remained relatively low between 1997 and 2008, and have varied around a mean of 1 since 2009.


## Introduction

The recreational fishery in the Gulf of Mexico is surveyed by the Marine Recreational Information Program (MRIP) conducted by NOAA Fisheries (formerly the Marine Recreational Fisheries Statistics Survey, MRFSS), the Texas Marine Sport-Harvest Monitoring Program conducted by the Texas Parks and Wildlife Department (TPWD), and the Southeast Region Headboat Survey (SRHS) conducted by NOAA Fisheries. MRIP/MRFSS has monitored shore based, charterboat and private/rental boat angler fishing in the Gulf of Mexico since 1981. MRIP data were used to construct an index of Vermilion Snapper catch rates in the Eastern U.S. Gulf of Mexico following the same procedures used in SEDAR45 and SEDAR09. The index was constructed using a delta-lognormal generalized linear model.

## Materials and Methods

## MRIP Transition

The Marine Recreational Information Program completed a three year transition in 2018 (NOAA Fisheries 2018). Estimates of fishing effort for the private and shore modes are now obtained from a Fishing Effort Survey conducted via mail, whereas previously these estimates came from the legacy Coastal Household Telephone Survey. Effort estimates for charter and party boats are still obtained from the For-Hire Telephone Survey and are not affected by the new Fishing Effort Survey. Benchmarking of the Fishing Effort Survey alongside the Coastal Household Telephone Survey for three years allowed for apples-to-apples comparisons between data from the two different surveys and the creation of a peer-reviewed calibration model. The calibration model was peer reviewed by reviewers appointed by the Center for Independent Experts (see Rago et al. (2017)). Additional details can be found at: https://www.fisheries.noaa.gov/event/fishing-effort-survey-calibration-model-peer-review. The MRIP transition also accounted for the 2013 design change in the Access Point Angler Intercept Survey (Foster et al. 2018). The MRIP transition resulted in the release of new recreational catch estimates for all species and all modes, including charter mode estimates. As a result, the SEFSC conducted a calibration analysis using the newly released data to correct for this change from the Coastal Household Telephone Survey to the For-Hire Telephone Survey (Dettloff and Matter 2019).

## MRIP Data

MRIP collects information on participation, effort, and species-specific catch. Data are collected to provide catch and effort estimates in two-month periods ("waves") for each recreational fishing mode (shore fishing, private/rental boat, charterboat, or headboat/charterboat combined prior to 1986) and for each area of fishing (inshore, state Territorial Seas, U.S. Exclusive Economic Zone), in each Gulf of Mexico state (except Texas). Total catch information is collected by MRIP on fish landed whole and observed by interviewers ("Type A"), fish reported as killed by the fishers ("Type B1") and fish reported as released alive by the fishers ("Type B2").

Data from the MRIP dockside interviews were used to characterize abundance trends of Vermilion Snapper in the Eastern U.S. Gulf of Mexico. Information on effort included hours fished and number of anglers as reported to the interviewer. Catch that was not observed by the interviewer (B1 and B2) was adjusted upwards by the ratio of non-interviewed to interviewed anglers in each group of anglers. The catch per unit effort was calculated on an individual group basis (i.e., by leader) and was equal to the number of fish caught ( $\mathrm{A}+\mathrm{B} 1+\mathrm{B} 2$ ) divided by the effort, where effort was the product of the number of anglers and the total hours fished.

## MRIP Data Filtering

Data were filtered following the same steps as SEDAR45 and SEDAR09:

1. Data in the Gulf of Mexico were limited to interviews that took place in Mississippi, Alabama, and Florida (including Monroe County)
2. Only interviews associated with private and charterboat fishing modes fishing hook and line gear were retained.
3. Interviews that reported shore-based fishing or fishing in inshore waters were excluded.
4. Interviews with possible error in effort information or in catch amount were excluded.
5. Data prior to 1986 were excluded.
6. Interviews that reached bag limits for Vermilion Snapper were retained.

## Species Association

An indirect method was necessary to infer targeting behavior of fishermen because no direct information was available. Following SEDAR45 and SEDAR09, the Stephens and MacCall (2004) approach was used to restrict the dataset to anglers that likely encountered Vermilion Snapper based on the trip's species composition.

## Standardization

A two-stage delta-lognormal generalized linear model (GLM; Lo et al. 1992) was used to standardize for variability and non-randomness in CPUE data collection methods not caused by the year effect (i.e., to factor out year to year variations in CPUE not due to changes in abundance). This method combines separate generalized linear model (GLM) analyses of the proportion of leaders that observed Vermilion Snapper and the catch rates under leaders that observed Vermilion Snapper to construct a single standardized index of abundance. In the first step, the proportion positive is modeled using a logit regression assuming a binomial distribution of the response variable. In the second step, the logarithm of CPUE on successful trips (those that caught the target species) was used as the response variable assuming a normal distribution and an identity link function. The two models were then combined to provide the final standardized index of abundance. Parameterization of each model was accomplished using a GLM procedure. For the lognormal models, the response variable, $\ln (C P U E)$, was calculated:
$\ln (C P U E)=\ln ((A+B 1+B 2) /($ anglersxhoursfished $))$
A forward stepwise regression approach was utilized within the GENMOD procedure of SAS 9.2 (SAS Institute, 2008). In this procedure, potential factors were added to the base model one at a time based on the percent reduction in deviance per degree of freedom. With each run of the model, the factor that caused the highest reduction in deviance was added to the base model (assuming the factor was significant based on a Chi-Square test with probability $<$ or $=0.05$ ) until no factor reduced the percent deviance by the pre-specified level (i.e., $1 \%$ ).

The following factors were examined as possible influences on the proportion of positive interviews, and the catch rates on positive interviews:

| Name | DF | Details |
| :--- | ---: | :--- |
| Year | 32 | $1986-2017$ |
| Time of Interview | 5 | $12 \mathrm{am}-1 \mathrm{pm}, 2 \mathrm{pm}, 3 \mathrm{pm}, 4 \mathrm{pm}, 5 \mathrm{pm}-11 \mathrm{pm}$ |
| Season | 4 | Dec-Feb, Mar-May ,Jun-Aug, Sep-Nov |
| Red Snapper Season | 2 | Open, Closed |
| State | 2 | FLW, AL/MS |
| Area | 2 | $<10$ miles offshore, > 10 miles offshore |
| Mode | 2 | Private, Charterboat |


| Name | DF | Details |
| :--- | ---: | :--- |
| Hours Fished* | 4 | $1-2,3-4,5-6,7+$ |
| Anglers* $^{*}$ | 7 | $1,2,3,4,5,6,7,8,9,10+$ |

*Only explored as factors for modeling success because these factors were confounded with effort for the CPUE response variable in the lognormal model.

The factor Red Snapper season is defined in Table 1. All factors were modeled as fixed effects and no interaction terms were examined following SEDAR45. Results of the binomial (proportion positive) and lognormal (mean CPUE on successful trips) were then multiplied to attain a single index of abundance based on the year effect. The final delta-lognormal model was fit using the SAS macro GLIMMIX (glmm800MaOB.sas: Russ Wolfinger, SAS Institute) and the SAS procedure PROC MIXED (SAS Institute Inc. 1997) following the procedures by Lo et al. (1992).

## Results and Discussion

## Species Associations - Stephens and MacCall (2004)

The minimum difference between the predicted and the observed number of interviews that reported Vermilion Snapper occurred at the probability threshold of 0.31 (Figure 1A). Interviews with a predicted probability that was greater than the critical threshold probability were identified as interviews that targeted Vermilion Snapper (Figure 1B). This method retained $4.2 \%$ of interviews, and $64.6 \%$ of interviews that reported Vermilion Snapper. Prior to trip selection, there were 214,616 interviews and the proportion positive was 0.04 , and after selection there were 8,929 interviews and the proportion positive was 0.65 . Given these diagnostics, sufficient interviews were retained to develop a standardized index of abundance.

The Stephens and MacCall (2004) trip subsetting approach identified 52 species which were captured with Vermilion Snapper and reflected either positive or negative associations (Table 2; Figure 2). For example, Red Porgy, Red Snapper, Gray Triggerfish, Tomtate, and Lane Snapper are positively correlated to Vermilion Snapper while Common Snook, Spotted Seatrout, Bonnethead, Southern Kingfish, and Sheepshead are negatively correlated. Overall, the trends in species associations were relatively similar to the associations identified during SEDAR45, although a few new species were included during SEDAR67 (Figure 2).

## Annual Abundance Indices

Table 3 summarizes the standardized index, corresponding lower and upper confidence limits, coefficients of variation, and nominal CPUE. Final deviance tables are included in Table 4. The final models for the binomial and lognormal components were:

ProportionPositive $=Y E A R+$ ANGLERS
$\ln (C P U E)=Y E A R$
As noted in Table 4, variable selection for SEDAR67 identified fewer variables for each model component than during SEDAR45 (Table 4, red text). Within the binomial and lognormal GLM
components, neither area nor Red Snapper season explained more than $1 \%$ deviance explained (Table 4).

The standardized index, with 95\% confidence intervals, is shown in Figure 3. Nominal values generally fell within the $95 \%$ confidence intervals, with exceptions noted in 1986 and 1994. Relative abundance peaked in 1986 and generally exceeded 1 (i.e., the mean) until 1995 (Figure 3). The lowest relative abundance occurred in 1997, with relatively low values persisting until 2009 (Figure 3). Since 2009, relatively abundance has varied around the time series mean of 1 (Figure 3).

Diagnostics for each component of the GLM are provided in Figure 4 and Figure 5. The overdispersion parameter for the binomial component was 1.39. The binomial model overestimates the proportion positive at both the beginning (1986-1996) and end (2009+) of the time series (Figure 4A). The proportion positive ranged from 0.49 to 0.94 , and has generally remained between 0.61 and 0.78 . Residual analysis of the binomial model indicated no obvious patterns in the residuals by year (Figure 4B), or number of anglers (Figure 4C).

The lognormal model results suggest a good fit to the data and indicated that the assumption of a lognormal distribution for positive catch was appropriate for the data (Figure 5A-B). Residual analysis of the lognormal model also indicated no obvious patterns in the residuals by year
(Figure 5C).
Figure 6 provides a comparison of the SEDAR67 MRIP index to the MRFSS index derived during SEDAR45. The differences between indices are primarily due to the change in variable selection. Running the index using the SEDAR45 recommended variables results in a much more similar index (results not shown). A fair number of SEDAR67 index values fall outside the confidence intervals of the SEDAR45 index, including 1988, 1992, 1997, 2008, 2009, 2011, 2012 and 2013 (Figure 7). However, overall, the relative trend and magnitude of the SEDAR67 index is generally similar to the SEDAR45 index.

## Comments on Adequacy for Assessment

The MRIP index presented in this working paper was deemed adequate for use in the SEDAR45 assessment. This decision during SEDAR45 was based on the long time series and large spatial coverage associated with the MRFSS angler intercept data. Additional work is needed to investigate the apparent shift in relative abundance starting in 1997, which is also evident in the headboat index developed for the Eastern U.S. Gulf of Mexico. While this year corresponds to a change in the recreational size limit from 8 inches total length to 10 inches total length, which would have impacted the discarding of fish, discarded fish (B2) were included when developing the MRIP index. In addition, a 20 reef fish aggregate was implemented in 1997, although no issues with exceeding bag limits were identified during this time.

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## Tables

Table 1. Red Snapper recreational season lengths by mode, open/close dates, and Federal Register references used for specifying the season in federal waters. F,Sa,Su refers to open only during Friday, Saturday, and Sunday.

| Year | Component | Days | Open Date | $\begin{gathered} \text { Close } \\ \text { Date } \end{gathered}$ | Effective Date | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pre-1990 | Private / For-hire | 365 | 1-Jan | $31-\mathrm{Dec}$ |  |  |
| 1990 | " | " | " | " |  |  |
| 1991 | " | " | " | " |  |  |
| 1992 | " | " | " | " |  |  |
| 1993 | " | " | " | " |  |  |
| 1994 | " | " | " | " |  |  |
| 1995 | " | " | " | " |  |  |
| 1996 | " | ${ }^{\prime \prime}$ | " | " |  |  |
| 1997 | " | 330 | " | 27-Nov | 11/27/1997 | 62 FR 61700 |
| 1998 | " | 272 | " | 30-Sep | 8/27/1998 | 63 FR 45760 |
| 1999 | " | 240 | " | 29-Aug | 6/4/1999 | 64 FR 30445 |
| 2000 | " | 194 | 21-Apr | 1-Nov | $\begin{aligned} & 1 / 19 / 2000 \\ & 9 / 18 / 2000 \end{aligned}$ | 64 FR 71056 <br> 65 FR 50158 |
| 2001 | " | " | " | " |  |  |
| 2002 | ${ }^{\prime}$ | " | " | ${ }^{\prime}$ |  |  |
| 2003 | " | " | ${ }^{\prime}$ | " |  |  |
| 2004 | " | " | " | " |  |  |
| 2005 | " | " | " | " |  |  |
| 2006 | " | " | " | " |  |  |
| 2007 | " | " | " | " | 5/2/2007 | 72 FR 15617 |
| 2008 | " | 65 | 1-Jun | 5-Aug | 8/5/2008 | 73 FR 15674 |
| 2009 | " | 75 | " | 15-Aug | 8/15/2009 | 74 FR 21558 |
| 2010 | " | 53 | " | 24-Jul | 6/2/2010 | 75 FR 23186 |
| 2011 | " | 48 | " | 19-Jul | 9/12/2011 | 76 FR 50143 |
| 2012 | " | 46 | " | 17-Jul | 7/11/2012 | 77 FR 39647 |
| 2013 | " | 42 | 1-Jun | $\begin{array}{r} 29-J u n ~ 15- \\ \text { Oct } \end{array}$ | $\begin{aligned} & \text { 6/29/2013 } \\ & 10 / 1 / 2013 \end{aligned}$ | $\begin{aligned} & 78 \text { FR } 34586 \\ & 78 \text { FR } 57313 \end{aligned}$ |
|  |  |  | 1-Oct |  |  |  |
| 2014 | " | 9 | " | 10-Jun | 5/15/2014 | 79 FR 27768 |
| 2015 | Private | 10 | " | 11-Jun | 6/1/2015 | 80 FR 24832 |
|  | For-hire | 44 | " | 15-Jul | 6/1/2015 | 80 FR 24832 |


| Year | Component | Days | Open Date | Close Date | Effective Date | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | Private | 11 | " | 12-Jun | 6/10/2016 | 81 FR 38110 |
|  | For-hire | 46 | " | 17-Jul | 6/10/2016 | 81 FR 25583 |
| 2017 | Private | 42 | 1-Jun | 3-Jun | 6/4/2017 | 82 FR 21140 |
|  |  |  | $\begin{array}{r} \text { 16-Jun } \\ (\mathrm{F}, \mathrm{Sa}, \mathrm{Su} \\ \text { only }) \end{array}$ | 5-Sep | 6/16/2017 | 82 FR 27777 |
|  |  |  | 3-Jul | 4-Jul |  |  |
|  |  |  | 4-Sep | 5-Sep |  |  |
|  | For-hire | 49 | 1-Jun | 19-Jul | 6/4/2017 | 82 FR 21140 |

Table 2. Association coefficients by species. Positive numbers indicate a positive correlation between a given species and Vermilion Snapper.

| Coefficient | Common Name | Scientific Name |
| ---: | :--- | :--- |
| 2.308 | Red Porgy | Pagrus pagrus |
| 1.913 | Red Snapper | Lutjanus campechanus |
| 1.580 | Gray Triggerfish | Balistes capriscus |
| 1.132 | Tomtate | Haemulon aurolineatum |
| 1.014 | Lane Snapper | Lutjanus synagris |
| 0.736 | Almaco Jack | Seriola rivoliana |
| 0.715 | Greater Amberjack | Seriola dumerili |
| 0.416 | Sand Perch | Diplectrum formosum |
| 0.384 | Red Grouper | Epinephelus morio |
| 0.382 | Scamp | Mycteroperca phenax |
| 0.359 | Little Tunny | Euthynnus alletteratus |
| 0.276 | Bluefish | Pomatomus saltatrix |
| 0.268 | Blackfin Tuna | Thunnus atlanticus |
| 0.238 | Round Scad | Decapterus punctatus |
| 0.162 | Gray Snapper | Lutjanus griseus |
| 0.074 | Requiem Shark Family | Carcharhinidae |
| 0.045 | Inshore Lizardfish | Synodus foetens |
| 0.039 | King Mackerel | Scomberomorus cavalla |
| 0.013 | Dolphin | Coryphaena hippurus |
| -0.004 | Black Grouper | Mycteroperca bonaci |
| -0.045 | Pinfish | Lagodon rhomboides |
| -0.059 | Pigfish | Orthopristis chrysoptera |
| -0.061 | Mutton Snapper | Lutjanus analis |
| -0.096 | Gag | Mycteroperca microlepis |
| -0.103 | Cobia | Rachycentron canadum |
| -0.153 | Blue Runner | Caranx crysos |
| -0.195 | Requiem Shark Genus | Carcharhinus spp |
| -0.241 | Great Barracuda | Sphyraena barracuda |
| -0.319 | Grunt Family | Haemulidae |
| -0.370 | Gulf Flounder | Paralichthys albigutta |
| -0.398 | Cero | Scomberomorus regalis |
| -0.415 | Atlantic Croaker | Micropogonias undulatus |
| -0.419 | White Grunt |  |


| Coefficient | Common Name | Scientific Name |
| ---: | :--- | :--- |
| -0.517 | Sand Seatrout | Cynoscion arenarius |
| -0.545 | Hardhead Catfish | Arius felis |
| -0.548 | Spanish Mackerel | Scomberomorus maculatus |
| -0.553 | Crevalle Jack | Caranx hippos |
| -0.592 | Blacktip Shark | Carcharhinus limbatus |
| -0.647 | Yellowtail Snapper | Ocyurus chrysurus |
| -0.731 | Sailfish | Istiophorus platypterus |
| -0.789 | Stingray Genus | Dasyatis spp. |
| -0.807 | Gafftopsail Catfish | Bagre marinus |
| -0.827 | Red Drum | Sciaenops ocellatus |
| -0.940 | Scaled Sardine | Harengula jaguana |
| -1.069 | Southern Puffer | Sphoeroides nephelus |
| -1.094 | Black Sea Bass | Centropristis striata |
| -1.246 | Ladyfish | Elops saurus |
| -2.038 | Sheepshead | Archosargus probatocephalus |
| -2.510 | Southern Kingfish | Menticirrhus americanus |
| -2.583 | Bonnethead | Sphyrna tiburo |
| -3.272 | Spotted Seatrout | Cynoscion nebulosus |
| -12.310 | Common Snook | Centropomus undecimalis |

Table 3. Numbers of total and positive interviews, proportion of positive interviews (PPT), relative nominal CPUE, and standardized abundance index statistics for Vermilion Snapper in the Eastern U.S. Gulf of Mexico.

|  |  |  |  | Relative |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | N | Positive N | PPT | Rominal <br> CPUE | Relative <br> Index | Lower <br> $95 \% \mathrm{CI}$ | Upper <br> $95 \% \mathrm{CI}$ | CV |
| 1986 | 176 | 127 | 0.722 | 1.711 | 2.800 | 2.296 | 3.416 | 0.100 |
| 1987 | 96 | 60 | 0.625 | 1.151 | 1.179 | 0.826 | 1.682 | 0.179 |
| 1988 | 30 | 27 | 0.900 | 1.716 | 1.911 | 1.281 | 2.851 | 0.202 |
| 1989 | 79 | 37 | 0.468 | 0.898 | 0.885 | 0.543 | 1.443 | 0.248 |
| 1990 | 59 | 45 | 0.763 | 1.783 | 2.229 | 1.548 | 3.208 | 0.184 |
| 1991 | 103 | 83 | 0.806 | 1.525 | 1.470 | 1.125 | 1.919 | 0.134 |
| 1992 | 182 | 152 | 0.835 | 1.238 | 1.382 | 1.129 | 1.691 | 0.101 |
| 1993 | 127 | 97 | 0.764 | 1.444 | 1.536 | 1.195 | 1.975 | 0.126 |
| 1994 | 103 | 64 | 0.621 | 2.078 | 1.434 | 1.018 | 2.020 | 0.173 |
| 1995 | 60 | 49 | 0.817 | 2.197 | 1.983 | 1.406 | 2.796 | 0.173 |
| 1996 | 67 | 41 | 0.612 | 1.123 | 1.007 | 0.644 | 1.574 | 0.226 |
| 1997 | 132 | 88 | 0.667 | 0.305 | 0.274 | 0.198 | 0.379 | 0.164 |
| 1998 | 190 | 116 | 0.611 | 0.454 | 0.361 | 0.269 | 0.484 | 0.147 |
| 1999 | 365 | 232 | 0.636 | 0.464 | 0.387 | 0.314 | 0.477 | 0.104 |
| 2000 | 429 | 248 | 0.578 | 0.360 | 0.347 | 0.253 | 0.475 | 0.159 |
| 2001 | 395 | 241 | 0.610 | 0.509 | 0.488 | 0.360 | 0.660 | 0.153 |
| 2002 | 465 | 242 | 0.520 | 0.352 | 0.363 | 0.269 | 0.489 | 0.151 |
| 2003 | 482 | 284 | 0.589 | 0.440 | 0.422 | 0.324 | 0.550 | 0.133 |
| 2004 | 726 | 469 | 0.646 | 0.624 | 0.543 | 0.439 | 0.672 | 0.107 |
| 2005 | 531 | 368 | 0.693 | 0.661 | 0.581 | 0.455 | 0.743 | 0.123 |
| 2006 | 413 | 287 | 0.695 | 0.583 | 0.537 | 0.410 | 0.703 | 0.136 |
| 2007 | 369 | 210 | 0.569 | 0.441 | 0.425 | 0.311 | 0.581 | 0.157 |
| 2008 | 305 | 187 | 0.613 | 0.772 | 0.662 | 0.475 | 0.922 | 0.167 |
| 2009 | 263 | 180 | 0.684 | 1.167 | 1.023 | 0.734 | 1.428 | 0.168 |
| 2010 | 281 | 167 | 0.594 | 0.749 | 0.561 | 0.393 | 0.801 | 0.180 |
| 2011 | 478 | 363 | 0.759 | 1.297 | 1.311 | 1.041 | 1.650 | 0.116 |
| 2012 | 416 | 264 | 0.635 | 0.773 | 0.881 | 0.670 | 1.159 | 0.138 |
| 2013 | 204 | 132 | 0.647 | 0.809 | 1.022 | 0.746 | 1.401 | 0.159 |
| 2014 | 333 | 251 | 0.754 | 1.287 | 1.186 | 0.950 | 1.481 | 0.111 |
| 2015 | 346 | 244 | 0.705 | 1.024 | 0.958 | 0.761 | 1.207 | 0.116 |
|  |  |  |  |  |  |  |  |  |


| Year | N | Positive N | PPT | Relative <br> Nominal <br> CPUE | Relative <br> Index | Lower <br> $95 \% \mathrm{CI}$ | Upper <br> $95 \% \mathrm{CI}$ | CV |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2016 | 401 | 258 | 0.643 | 0.837 | 0.679 | 0.538 | 0.855 | 0.116 |
| 2017 | 307 | 209 | 0.681 | 1.226 | 1.176 | 0.929 | 1.489 | 0.118 |

Table 4. Final deviance tables for the regressions for Vermilion Snapper in the Eastern U.S. Gulf of Mexico. The table shows the order of the factors as they were sequentially added to each model. Fit diagnostics listed for each factor were the diagnostics from a model that included that factor and all of the factors listed above it in the tables below. Note that variables in red were included during SEDAR45 but not included for SEDAR67 due to a percent deviance reduction < $1 \%$.

| Factor | DF | Deviance | Residual DF | Residual Deviance | AIC | Deviance Reduced | Log <br> likelihood | Likelihood Ratio Test |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Binomial |  |  |  |  |  |  |  |  |
| Null | 1 | 3825 | 2823 | 3825 | 3825 | - | -1912 | - |
| Year | 32 | 3524 | 2792 | 301 | 3524 | 6.86\% | -1762 | 301.6 |
| Anglers | 10 | 3424 | 2783 | 99 | 3424 | 2.51\% | -1712 | 99.6 |
| Area | 2 | 3408 | 2782 | 16 | 3408 | 0.44\% | -1704 | 16.4 |
| Lognormal |  |  |  |  |  |  |  |  |
| Null | 1 | 3047 | 1661 | 3047 | 5724 | - | -2862 | - |
| Year | 32 | 2531 | 1630 | 515 | 5415 | 15.35\% | -2707 | 308.2 |
| Red Snapper Season | 2 | 2513 | 1629 | 17 | 5404 | 0.65\% | -2702 | 11.8 |

## Figures



Figure 1. The difference between the number of records in which Vermilion Snapper are observed and the number in which they are predicted to occur for each probability threshold (A). Histogram of probabilities generated by the species-based regression (B). The dashed vertical line indicates the critical value where false prediction is minimized.


Figure 2. Comparison of coefficients obtained from the Stephens and MacCall (2004) trip selection approach for SEDAR67 and the previous SEDAR45 assessment.


Figure 3. Standardized indices with $95 \%$ confidence intervals and nominal CPUE for Vermilion Snapper in the Eastern U.S. Gulf of Mexico.


Figure 4. Diagnostic plots for the binomial model for Vermilion Snapper in the Eastern U.S. Gulf of Mexico. Shown here are the predicted (solid line) and observed proportion of positive interviews by year (A), and the residuals from the binomial model by year (B) and number of anglers (C).


Figure 5. Diagnostic plots for the lognormal model of catch rates on positive trips for Vermilion Snapper in the Eastern U.S. Gulf of Mexico. Shown here are the frequency distribution of catch rates (A), the cumulative normalized residuals (B), and the distribution of residuals by year (C). The red lines represent the expected normal distribution.


Figure 6. Standardized index for Vermilion Snapper in the Eastern U.S. Gulf of Mexico for SEDAR67 compared to the index provided during SEDAR45. For comparison, both indices have been normalized by their respective means.


Figure 7. Comparison of index for Vermilion Snapper in the Eastern U.S. Gulf of Mexico for SEDAR67 compared to the index provided during SEDAR45 with confidence intervals.

