# Standardized Catch Rate Indices for Vermilion Snapper (Rhomboplites aurorubens) during 1986-2017 by the U.S. Gulf of Mexico Headboat Recreational Fishery 

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## Keywords

CPUE, catch, effort, recreational fisheries, Vermilion Snapper

## Abstract

Two delta-lognormal indices, for the Eastern and Western U.S. Gulf of Mexico, were constructed for the SEDAR67 Standard Vermilion Snapper stock assessment. Each index uses data from the Southeast Region Headboat Survey (HBS). Indices for the Eastern and Western U.S. Gulf of Mexico are developed following the same methodology and approach used for SEDAR45 and SEDAR09. The resulting indices reveal relatively similar trends when compared to the SEDAR45 indices for both regions. For the Eastern U.S. Gulf of Mexico, the SEDAR67 standardized index indicates catch rates were relatively high until 1995 but have remained below the mean since 1996, with the exception of 2017. For the Western U.S. Gulf of Mexico, the SEDAR67 standardized index indicates catch rates have been highly variable throughout the time series, with recent relative abundance near the mean.

## Introduction

The recreational fishery in the Gulf of Mexico is surveyed by the Marine Recreational Information Program (MRIP) conducted by NOAA Fisheries (formerly the Marine Recreational Fisheries Statistics Survey, MRFSS), the Texas Marine Sport-Harvest Monitoring Program conducted by the Texas Parks and Wildlife Department (TPWD), and the Southeast Region Headboat Survey (SRHS) conducted by NOAA Fisheries. The SRHS has monitored catch and effort from party (head) boats in the Gulf of Mexico since 1986. SRHS data were used to construct an index of Vermilion Snapper catch rates in both the Eastern and Western U.S. Gulf
of Mexico following the same procedures used in SEDAR45 and SEDAR09. The index was constructed using a delta-lognormal generalized linear model.

## Materials and Methods

## Headboat Data

The Southeast Region Headboat Survey collects data on the catch and effort for individual headboat trips. Reported information includes landing date and location, vessel identification, the number of anglers, a single fishing location ( $10^{\prime}$ x 10' rectangle of latitude and longitude) for the entire trip, trip duration and/or type (half/three-quarter/full/multi-day, day/night, morning/afternoon), and catch by species in number and weight.

SRHS data were used to characterize abundance trends of Vermilion Snapper in the Eastern and Western U.S. Gulf of Mexico. Catch per unit effort (CPUE) was calculated on an individual trip basis. CPUE for each trip was defined as the number of Vermilion Snapper landed on a trip divided by the effort, where effort was the product of the number of anglers and the total hours fished. To estimate effort for each trip type (i.e., trip duration), the following assumptions were necessary: Half day trip = 5 hours fished; Three-quarter day trip $=7.5$ hours fished, Full day trip $=10$ hours fished; and Multi-day trip $=>10$ hours fished.

## Headboat Data Filtering

Data were filtered following the same steps as SEDAR45 and SEDAR09. Trips were eliminated if they had missing values for any of the key factors, were in anyway incomplete, appeared to be misreported (e.g., reported zero anglers), or represented multiple entries for a single trip. Two indices (Eastern U.S. Gulf of Mexico and Western U.S. Gulf of Mexico) were calculated based on geographic area (east or west of the Mississippi delta) to better represent the variance and abundance trends in each zone, because effort can vary significantly from year-to-year between the two areas.

## Species Association

An indirect method was necessary to infer targeting behavior of fishermen because no direct information was available. Following SEDAR45 and SEDAR09, the Stephens and MacCall (2004) approach was used to restrict the dataset to anglers that likely encountered Vermilion Snapper based on the trip's species composition.

## Standardization

A two-stage delta-lognormal generalized linear model (GLM; Lo et al. 1992) was used to standardize for variability and non-randomness in CPUE data collection methods not caused by the year effect (i.e., to factor out year to year variations in CPUE not due to changes in abundance). This method combines separate generalized linear model (GLM) analyses of the proportion of trips that observed Vermilion Snapper and the catch rates under trips that observed Vermilion Snapper to construct a single standardized index of abundance. In the first step, the proportion positive is modeled using a logit regression assuming a binomial distribution of the response variable. In the second step, the logarithm of CPUE on successful trips (those that caught the target species) was used as the response variable assuming a normal distribution and
an identity link function. The two models were then combined to provide the final standardized index of abundance. Parameterization of each model was accomplished using a GLM procedure. For the lognormal models, the response variable, $\ln (C P U E)$, was calculated:
$\ln ($ CPUE $)=\ln ($ Catch $) /($ anglersxhoursfished $))$
A forward stepwise regression approach was utilized within the GENMOD procedure of SAS 9.2 (SAS Institute, 2008). In this procedure, potential factors were added to the base model one at a time based on the percent reduction in deviance per degree of freedom. With each run of the model, the factor that caused the highest reduction in deviance was added to the base model (assuming the factor was significant based on a Chi-Square test with probability < or $=0.05$ ) until no factor reduced the percent deviance by the pre-specified level (i.e., $1 \%$ ). Since the goal of the standardization process was to model time trends in abundance, it was necessary to force the year effect as a factor even if it was not deemed significant. No interaction terms were examined following SEDAR45. Factors modeled as fixed effects included year, season, Red Snapper season (Table 1), time of day, trip duration, month and area.

The variation in catch rates by vessel was examined using a "repeated measures" approach (Littell et al., 1998). The term 'repeated measures' refers to multiple measurements taken over time on the same experimental unit (i.e. vessel). Specifying the repeated measure "VESSEL" and the subject "VESSEL(YEAR)" allows PROC MIXED to model the covariance structure of the data. This is particularly important because catch rates may vary by vessel and because catch rates by a given vessel that are close in time can be more highly correlated than those far apart in time (Littell et al., 1998).

Results of the binomial (proportion positive) and lognormal (mean CPUE on successful trips) were then multiplied to attain a single index of abundance based on the year effect. The final delta-lognormal model was fit using the SAS macro GLIMMIX (glmm800MaOB.sas: Russ Wolfinger, SAS Institute) and the SAS procedure PROC MIXED (SAS Institute Inc. 1997) following the procedures by Lo et al. (1992).

## Results and Discussion

## Western U.S. Gulf of Mexico

## Species Associations - Stephens and MacCall (2004)

The minimum difference between the predicted and the observed number of trips that reported Vermilion Snapper occurred at the probability threshold of 0.34 (Figure 1A). Trips with a predicted probability that was greater than the critical threshold probability were identified as trips that targeted Vermilion Snapper (Figure 1B). This method retained $34.2 \%$ of trips, and $64.3 \%$ of trips that reported Vermilion Snapper. Prior to trip selection, there were 67,723 trips and the proportion positive was 0.34 , and after selection there were 23,145 trips and the proportion positive was 0.64 . Given these diagnostics, sufficient trips were retained to develop a standardized index of abundance.

The Stephens and MacCall (2004) trip subsetting approach identified 23 species which were captured with Vermilion Snapper and reflected either positive or negative associations (Table 2; Figure 2). For example, Scamp, Greater Amberjack, Red Snapper, Lane Snapper, and Gray

Triggerfish are positively correlated to Vermilion Snapper while Sand Seatrout, Crevalle Jack, Spanish Mackerel, Atlantic Spadefish, and Gray Snapper are negatively correlated. Trip selection for SEDAR67 identified many more species compared to SEDAR45, with more pelagics and reef fishes associated with Vermilion Snapper (Figure 2).

## Variable Selection

The following factors were examined as possible influences on the proportion of positive trips, and the catch rates on positive trips:

| Name | DF | Details |
| ---: | ---: | ---: |
| Year | 32 | 1986-2017 |
| Season | 4 | Dec-Feb, Mar-May, Jun-Aug, Sep-Nov |
| Red Snapper Season | 2 | Open, Closed |
| Day/Night* | 3 | Day, Night, Both |
| Trip Duration* | 4 | Half Day, Three Quarter Day, Full Day, Multi Day |
| Month | 12 | Jan, Feb, Mar, Apr, May, Jun, Jul, Aug, Sep, Oct, Nov, Dec |
| Vessel | $51(46)$ | Individual vessels |
| Area | 2 | SW TX, MS LA NE TX |

*Only explored as factors for modeling success because these factors were confounded with effort for the CPUE response variable in the lognormal model. Note the DF in parenthesis for vessels is for the lognormal model.

## Annual Abundance Indices

Table 3 summarizes the standardized index, corresponding lower and upper confidence limits, coefficients of variation, and nominal CPUE. Final deviance tables are included in Table 4. The final models for the binomial and lognormal components were:

ProportionPositive $=Y E A R+$ TRIPDURATION
$\ln (C P U E)=Y E A R+R E D S N A P P E R S E A S O N+V E S S E L$
Variable selection for SEDAR67 identified fewer variables for the binomial model than during SEDAR45 (Table 4, red text). Year was not significant in the binomial model (Table 4) but was included to force the year effect in the standardization process for SEDAR67.

The standardized index, with $95 \%$ confidence intervals, is shown in Figure 3. The majority of nominal values fell within the $95 \%$ confidence intervals, with the exception of 1986, 1993, 2009 and 2014. Relative abundance peaked in 1990 and has varied around the mean throughout much of the time series, with the exception of the lowest value in 2008 (Figure 3). Relative abundance has remained relatively stable since 2010.

Diagnostics for each component of the GLM are provided in Figure 4 and Figure 5. The overdispersion parameter for the binomial component was 3.87 . As in the SEDAR45 index, the binomial model consistently underestimates the proportion positive (Figure 4A). The proportion positive ranged from 0.28 to 0.69 , and has generally remained between 0.49 and 0.56 . The
proportion positive declined substantially in 2008 to approximately $30 \%$. Residual analysis of the binomial model indicated no obvious patterns in the residuals by year (Figure 4B), or trip duration (Figure 4C).

The lognormal model results suggest a good fit to the data and indicated that the assumption of a lognormal distribution for positive catch was appropriate for the data (Figure 5A-B). Residual analysis of the lognormal model also indicated no obvious patterns in the residuals by year (Figure 5C), red snapper season (Figure 5D) or vessel (Figure 5E).

Figure 6 provides a comparison of the SEDAR67 headboat index to the headboat index derived during SEDAR45 for the Western U.S. Gulf of Mexico. Although some slight differences are evident, likely due to changes in trip selection, the trend and magnitude of the continuity index are similar to the index developed during SEDAR45. The changes observed in 2013 and 2014 are likely the result of correcting coding errors in the red snapper season (see SEDAR52-WP13). Overall, all index values for SEDAR67 with the exception of 2013 remain within the confidence intervals of the SEDAR45 index (Figure 7).

## Eastern U.S. Gulf of Mexico

## Species Associations - Stephens and MacCall (2004)

The minimum difference between the predicted and the observed number of trips that reported Vermilion Snapper occurred at the probability threshold of 0.37 (Figure 8A). Trips with a predicted probability that was greater than the critical threshold probability were identified as trips that targeted Vermilion Snapper (Figure 8B). This method retained $38.9 \%$ of trips, and $89.8 \%$ of trips that reported Vermilion Snapper. Prior to trip selection, there were 180,970 trips and the proportion positive was 0.39 , and after selection there were 70,329 trips and the proportion positive was 0.89 . Given these diagnostics, sufficient trips were retained to develop a standardized index of abundance.

The Stephens and MacCall (2004) trip subsetting approach identified 34 species which were captured with Vermilion Snapper and reflected either positive or negative associations (Table 5;
Figure 9). For example, Red Porgy, Red Snapper, Littlehead Porgy, Gray Triggerfish, and Almaco Jack are positively correlated to Vermilion Snapper while White Grunt, Black Sea Bass, Hogfish, Black Grouper, and Pigfish are negatively correlated. Trip selection for SEDAR67 identified many more species compared to SEDAR45, with more pelagics and reef fishes associated with Vermilion Snapper (Figure 9).

## Variable Selection

The following factors were examined as possible influences on the proportion of positive trips, and the catch rates on positive trips:

| Name | DF | Details |
| ---: | ---: | ---: |
| Year | 32 | 1986-2017 |
| Season | 4 | Dec-Feb, Mar-May, Jun-Aug, Sep-Nov |
| Red Snapper Season | 2 | Open, Closed |
| Day/Night* | 3 | Day, Night, Both |


| Name | DF | Details |
| ---: | ---: | ---: |
| Trip Duration* | 4 | Half Day, Three Quarter Day, Full Day, Multi Day |
| Month | 12 | Jan, Feb, Mar, Apr, May, Jun, Jul, Aug, Sep, Oct, Nov, Dec |
| Vessel | $126(102)$ | Individual vessels |
| Area | 2 | FL MG NW FL AL, SW FL |

*Only explored as factors for modeling success because these factors were confounded with effort for the CPUE response variable in the lognormal model. Note the DF in parenthesis for vessels is for the lognormal model.

## Annual Abundance Indices

Table 6 summarizes the standardized index, corresponding lower and upper confidence limits, coefficients of variation, and nominal CPUE. Final deviance tables are included in Table 7. The final models for the binomial and lognormal components were:

ProportionPositive $=Y E A R+$ AREA + TRIPDURATION
$\ln (C P U E)=Y E A R+M O N T H+V E S S E L$
Variable selection for SEDAR67 identified fewer variables for each model component than during SEDAR45 (Table 7, red text). Note that red snapper season was not significant in the SEDAR67 lognormal model, whereas it was significant in the SEDAR45 lognormal model (Table 7).

The standardized index, with $95 \%$ confidence intervals, is shown in Figure 10. The majority of nominal values fell within the $95 \%$ confidence intervals, with the exception of 1994, 1998, 2000, 2001, 2004 and 2017. Relative abundance was relatively high during the beginning of the time series, peaked in 1992 and declined until the lowest value in 1998 (Figure 10). Relative abundance has gradually increased since 1998, with a few dips noted in 2007 and 2012.

Diagnostics for each component of the GLM are provided in Figure 11 and Figure 12 and are similar to the patterns displayed in the SEDAR45 index. The overdispersion parameter for the binomial component was 7.90 . As in the SEDAR45 index, the binomial model consistently underestimates the proportion positive (Figure 11A). The proportion positive ranged from 0.43 to 0.93 , and has generally remained between 0.61 and 0.78 . Residual analysis of the binomial model indicated no obvious patterns in the residuals by year (Figure 11B), area (Figure 11C) or trip duration (Figure 4D).

The lognormal model results suggest a less than optimal fit to the data and indicated that the assumption of a lognormal distribution for positive catch was appropriate for the data (Figure 12A-B). Residual analysis of the lognormal model also indicated no obvious patterns in the residuals by year (Figure 12C), month (Figure 12D) or vessel (Figure 12E).

Figure 13 provides a comparison of the SEDAR67 headboat index to the headboat index derived during SEDAR45 for the Eastern U.S. Gulf of Mexico. The continuity index is very similar to the index developed during SEDAR45, although some differences are evident starting in 2008 (Figure 13). However, all index values for SEDAR67 remain within the confidence intervals of the SEDAR45 index (Figure 14).

## Comments on Adequacy for Assessment

The headboat indices presented in this working paper reflect the continuity indices of the headboat indices that were deemed adequate for use in the SEDAR45 assessment. Additional work is needed to investigate the apparent shift in relative abundance starting in 1997, which is also evident in the MRIP index developed for the Eastern U.S. Gulf of Mexico. This year corresponds to a change in the recreational size limit from 8 inches total length to 10 inches total length. This regulation change would have impacted discarded fish which are not reported in the SRHS, although they are reported in the MRIP dataset. In addition, a 20 reef fish aggregate was implemented in 1997, although no issues with exceeding bag limits were identified during this time.

## References

Littell, R.C., P.R. Henry and C.B. Ammerman. 1998. Statistical analysis of repeated measures data using SAS procedures. J. Anim. Sci. 76:1216-1231.

Lo, N.C. L.D. Jacobson, and J.L. Squire. 1992. Indices of relative abundance from fish spotter data based on delta-lognormal models. Can. J. Fish. Aquat. Sci. 49: 2515-2526.

SAS Institute Inc. 1997, SAS/STAT? Software: Changes and Enhancements through Release 6.12. Cary, NC:Sas Institute Inc., 1997. 1167 pp.

Stephens, A. and A. MacCall. 2004. A multispecies approach to subsetting logbook data for purposes of estimating CPUE. Fisheries Research 70:299-310.

## Tables

Table 1. Red Snapper recreational season lengths by mode, open/close dates, and references used for specifying the season in federal waters. F,Sa,Su refers to open only during Friday, Saturday, and Sunday.

| Year | Mode | Days | Open Date | Close Date | Effective Date | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pre-1990 | Private / For-hire | 365 | 1-Jan | 31-Dec |  |  |
| 1990 | " | " | " | " |  |  |
| 1991 | " | " | " | " |  |  |
| 1992 | " | " | " | " |  |  |
| 1993 | " | " | " | " |  |  |
| 1994 | " | " | " | " |  |  |
| 1995 | " | " | " | " |  |  |
| 1996 | " | " | " | " |  |  |
| 1997 | " | 330 | " | 27-Nov | 11/27/1997 | 62 FR 61700 |
| 1998 | " | 272 | " | 30-Sep | 8/27/1998 | 63 FR 45760 |
| 1999 | " | 240 | " | 29-Aug | 6/4/1999 | 64 FR 30445 |
| 2000 | " | 194 | 21-Apr | 1-Nov | $\begin{aligned} & 1 / 19 / 2000 \\ & 9 / 18 / 2000 \end{aligned}$ | 64 FR 71056 <br> 65 FR 50158 |
| 2001 | " | " | " | " |  |  |
| 2002 | " | " | " | " |  |  |
| 2003 | " | " | " | " |  |  |
| 2004 | " | " | " | " |  |  |
| 2005 | " | " | " | " |  |  |
| 2006 | " | " | " | " |  |  |
| 2007 | " | " | " | " | 5/2/2007 | 72 FR 15617 |
| 2008 | " | 65 | 1-Jun | 5-Aug | 8/5/2008 | 73 FR 15674 |
| 2009 | " | 75 | " | 15-Aug | 8/15/2009 | 74 FR 21558 |
| 2010 | " | 53 | " | 24-Jul | 6/2/2010 | 75 FR 23186 |
| 2011 | " | 48 | " | 19-Jul | 9/12/2011 | 76 FR 50143 |
| 2012 | " | 46 | " | 17-Jul | 7/11/2012 | 77 FR 39647 |
| 2013 | " | 42 | 1-Jun | $\begin{array}{r} 29-\text { Jun } 15- \\ \text { Oct } \end{array}$ | $\begin{aligned} & \text { 6/29/2013 } \\ & 10 / 1 / 2013 \end{aligned}$ | $\begin{aligned} & 78 \text { FR } 34586 \\ & 78 \text { FR } 57313 \end{aligned}$ |
|  |  |  | 1-Oct |  |  |  |
| 2014 | " | 9 | " | 10-Jun | 5/15/2014 | 79 FR 27768 |
| 2015 | Private | 10 | " | 11-Jun | 6/1/2015 | 80 FR 24832 |
|  | For-hire | 44 | " | 15-Jul | 6/1/2015 | 80 FR 24832 |


| Year | Mode | Days | Open Date | Close Date | Effective Date | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | Private | 11 | " | 12-Jun | 6/10/2016 | 81 FR 38110 |
|  | For-hire | 46 | " | 17-Jul | 6/10/2016 | 81 FR 25583 |
| 2017 | Private | 42 | 1-Jun | 3-Jun | 6/4/2017 | 82 FR 21140 |
|  |  |  | $\begin{array}{r} 16-\mathrm{Jun} \\ (\mathrm{~F}, \mathrm{Sa}, \mathrm{Su} \\ \text { only }) \end{array}$ | 5-Sep | 6/16/2017 | 82 FR 27777 |
|  |  |  | 3-Jul | 4-Jul |  |  |
|  |  |  | 4-Sep | 5-Sep |  |  |
|  | For-hire | 49 | 1-Jun | 19-Jul | 6/4/2017 | 82 FR 21140 |

Table 2. Association coefficients by species for Western U.S. Gulf of Mexico. Positive numbers indicate a positive correlation between a given species and Vermilion Snapper.

| Coefficient | Common Name | Scientific Name |
| ---: | :--- | :--- |
| 1.828 | Scamp | Mycteroperca phenax |
| 1.807 | Greater Amberjack | Seriola dumerili |
| 1.699 | Red Snapper | Lutjanus campechanus |
| 1.596 | Lane Snapper | Lutjanus synagris |
| 1.270 | Gray Triggerfish | Balistes capriscus |
| 1.145 | Almaco Jack | Seriola rivoliana |
| 0.949 | Warsaw Grouper | Epinephelus nigritus |
| 0.846 | Atlantic Sharpnose Shark | Rhizoprionodon terraenovae |
| 0.750 | Blackfin Tuna | Thunnus atlanticus |
| 0.739 | King Mackerel | Scomberomorus cavalla |
| 0.671 | Rock Hind | Epinephelus adscensionis |
| 0.417 | Dolphin | Coryphaena hippurus |
| 0.250 | Blue Runner | Caranx crysos |
| 0.199 | Cobia | Rachycentron canadum |
| 0.025 | Blacktip Shark | Carcharhinus limbatus |
| -0.108 | Gag | Mycteroperca microlepis |
| -0.255 | Bluefish | Pomatomus saltatrix |
| -0.310 | Little Tunny | Euthynnus alletteratus |
| -0.337 | Gray Snapper | Lutjanus griseus |
| -0.616 | Atlantic Spadefish | Chaetodipterus faber |
| -0.829 | Spanish Mackerel | Scomberomorus maculatus |
| -0.845 | Crevalle Jack | Caranx hippos |
| -1.186 | Sand Seatrout | Cynoscion arenarius |
|  |  |  |

Table 3. Numbers of total and positive trips, proportion of positive trips (PPT), relative nominal CPUE, and standardized abundance index statistics for Vermilion Snapper in the Western U.S. Gulf of Mexico.

|  |  |  |  | Relative |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | N | Positive N | PPT | Nominal <br> CPUE | Relative <br> Index | Lower <br> $95 \% \mathrm{CI}$ | Upper <br> $95 \% \mathrm{CI}$ | CV |
| 1986 | 666 | 425 | 0.638 | 0.996 | 1.752 | 1.121 | 2.738 | 0.226 |
| 1987 | 833 | 576 | 0.691 | 0.872 | 1.223 | 0.799 | 1.873 | 0.215 |
| 1988 | 736 | 450 | 0.611 | 0.732 | 0.928 | 0.586 | 1.470 | 0.233 |
| 1989 | 703 | 439 | 0.624 | 1.135 | 1.291 | 0.832 | 2.001 | 0.222 |
| 1990 | 706 | 500 | 0.708 | 1.622 | 1.767 | 1.175 | 2.657 | 0.206 |
| 1991 | 761 | 471 | 0.619 | 1.011 | 0.983 | 0.648 | 1.493 | 0.211 |
| 1992 | 1084 | 683 | 0.630 | 0.832 | 0.945 | 0.638 | 1.398 | 0.198 |
| 1993 | 1184 | 798 | 0.674 | 0.725 | 1.150 | 0.797 | 1.659 | 0.185 |
| 1994 | 1292 | 891 | 0.690 | 0.830 | 1.137 | 0.795 | 1.627 | 0.180 |
| 1995 | 1331 | 926 | 0.696 | 0.874 | 1.214 | 0.851 | 1.732 | 0.179 |
| 1996 | 1064 | 736 | 0.692 | 0.749 | 0.886 | 0.612 | 1.281 | 0.186 |
| 1997 | 898 | 579 | 0.645 | 0.794 | 0.837 | 0.564 | 1.242 | 0.199 |
| 1998 | 915 | 582 | 0.636 | 0.699 | 0.796 | 0.545 | 1.163 | 0.191 |
| 1999 | 513 | 343 | 0.669 | 0.524 | 0.687 | 0.444 | 1.063 | 0.221 |
| 2000 | 678 | 457 | 0.674 | 0.709 | 0.519 | 0.340 | 0.793 | 0.214 |
| 2001 | 815 | 503 | 0.617 | 0.757 | 0.836 | 0.556 | 1.256 | 0.206 |
| 2002 | 874 | 556 | 0.636 | 0.742 | 0.974 | 0.664 | 1.429 | 0.193 |
| 2003 | 974 | 551 | 0.566 | 0.696 | 0.636 | 0.435 | 0.929 | 0.191 |
| 2004 | 986 | 603 | 0.612 | 1.007 | 1.091 | 0.751 | 1.585 | 0.188 |
| 2005 | 1034 | 609 | 0.589 | 0.957 | 1.218 | 0.843 | 1.761 | 0.186 |
| 2006 | 1067 | 550 | 0.515 | 0.581 | 0.652 | 0.437 | 0.972 | 0.202 |
| 2007 | 838 | 496 | 0.592 | 1.223 | 1.438 | 0.976 | 2.117 | 0.195 |
| 2008 | 329 | 124 | 0.377 | 0.282 | 0.261 | 0.142 | 0.481 | 0.313 |
| 2009 | 524 | 242 | 0.462 | 0.574 | 0.344 | 0.215 | 0.551 | 0.238 |
| 2010 | 381 | 264 | 0.693 | 1.215 | 1.140 | 0.728 | 1.784 | 0.227 |
| 2011 | 320 | 227 | 0.709 | 1.706 | 1.165 | 0.744 | 1.824 | 0.227 |
| 2012 | 353 | 211 | 0.598 | 1.290 | 0.913 | 0.571 | 1.460 | 0.238 |
| 2013 | 296 | 234 | 0.791 | 1.642 | 1.103 | 0.686 | 1.771 | 0.240 |
| 2014 | 240 | 175 | 0.729 | 1.651 | 0.896 | 0.526 | 1.527 | 0.271 |
| 2015 | 277 | 214 | 0.773 | 1.627 | 1.053 | 0.660 | 1.680 | 0.237 |
|  |  |  |  |  |  |  |  |  |


| Year | N | Positive N | PPT | Relative <br> Nominal <br> CPUE | Relative <br> Index | Lower <br> $95 \% \mathrm{CI}$ | Upper <br> $95 \% \mathrm{CI}$ | CV |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2016 | 235 | 194 | 0.826 | 1.731 | 1.151 | 0.707 | 1.874 | 0.247 |
| 2017 | 238 | 160 | 0.672 | 1.216 | 1.015 | 0.591 | 1.742 | 0.275 |

Table 4. Final deviance tables for the regressions for Vermilion Snapper in the Western U.S. Gulf of Mexico. The table shows the order of the factors as they were sequentially added to each model. Fit diagnostics listed for each factor were the diagnostics from a model that included that factor and all of the factors listed above it in the tables below. Note that variable in red was included in SEDAR67 to force the year effect in the standardization process.

| Factor | DF | Deviance | Residual <br> DF | Residual <br> Deviance | AIC | Deviance <br> Reduced | LogLikelihood <br> Rakelihood |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :--- |
| Binomial Test |  |  |  |  |  |  |  |

Table 5. Association coefficients by species for Eastern U.S. Gulf of Mexico. Positive numbers indicate a positive correlation between a given species and Vermilion Snapper.

| Coefficient | Common Name | Scientific Name |
| ---: | :--- | :--- |
| 3.006 | Red Porgy | Pagrus pagrus |
| 1.856 | Red Snapper | Lutjanus campechanus |
| 1.535 | Littlehead Porgy | Calamus proridens |
| 1.484 | Gray Triggerfish | Balistes capriscus |
| 1.321 | Almaco Jack | Seriola rivoliana |
| 1.157 | Whitebone Porgy | Calamus leucosteus |
| 1.081 | Banded Rudderfish | Seriola zonata |
| 0.820 | Lane Snapper | Lutjanus synagris |
| 0.743 | King Mackerel | Scomberomorus cavalla |
| 0.688 | Tomtate | Haemulon aurolineatum |
| 0.650 | Dolphin | Coryphaena hippurus |
| 0.641 | Scamp | Mycteroperca phenax |
| 0.562 | Knobbed Porgy | Calamus nodosus |
| 0.537 | Greater Amberjack | Seriola dumerili |
| 0.435 | Bank Sea Bass | Centropristis ocyurus |
| 0.362 | Little Tunny | Euthynnus alletteratus |
| 0.359 | Yellowtail Snapper | Ocyurus chrysurus |
| 0.159 | Red Grouper | Epinephelus morio |
| 0.112 | Blue Runner | Caranx crysos |
| 0.085 | Cobia | Rachycentron canadum |
| 0.035 | Pinfish | Lagodon rhomboides |
| 0.021 | Gray Snapper | Lutjanus griseus |
| -0.060 | Saucereye Porgy | Calamus calamus |
| -0.194 | Jolthead Porgy | Calamus bajonado |
| -0.255 | Gag | Mycteroperca microlepis |
| -0.327 | Gulf Flounder | Paralichthys albigutta |
| -0.390 | Sand Perch | Diplectrum formosum |
| -0.448 | Grass Porgy | Calamus arctifrons |
| -0.612 | Spanish Mackerel | Scomberomorus maculatus |
| -0.862 | Pigfish | Orthopristis chrysoptera |
| -0.972 | Black Grouper | Mycteroperca bonaci |
| -0.988 | Hogfish | Lachnolaimus maximus |
| -1.257 | Black Sea Bass | Centropristis striata |
| Cla |  |  |


| Coefficient | Common Name | Scientific Name |
| ---: | :--- | :--- |
| -1.879 | White Grunt | Haemulon plumieri |

Table 6. Numbers of total and positive trips, proportion of positive trips (PPT), relative nominal CPUE, and standardized abundance index statistics for Vermilion Snapper in the Eastern U.S. Gulf of Mexico.

|  |  |  |  |  | Relative |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | N | Positive N | PPT | Rominal <br> CPUE | Relative <br> Index | Lower <br> $95 \%$ CI | Upper <br> $95 \%$ CI | CV |
| 1986 | 641 | 452 | 0.705 | 0.851 | 0.900 | 0.546 | 1.485 | 0.254 |
| 1987 | 829 | 650 | 0.784 | 1.244 | 1.009 | 0.624 | 1.629 | 0.243 |
| 1988 | 1782 | 1664 | 0.934 | 2.167 | 2.163 | 1.546 | 3.027 | 0.169 |
| 1989 | 1779 | 1580 | 0.888 | 1.413 | 1.343 | 0.958 | 1.882 | 0.170 |
| 1990 | 2322 | 1976 | 0.851 | 1.475 | 1.689 | 1.234 | 2.312 | 0.158 |
| 1991 | 2229 | 1953 | 0.876 | 1.485 | 1.803 | 1.321 | 2.461 | 0.157 |
| 1992 | 2417 | 2233 | 0.924 | 1.879 | 2.499 | 1.855 | 3.366 | 0.150 |
| 1993 | 2783 | 2502 | 0.899 | 1.359 | 1.599 | 1.175 | 2.175 | 0.155 |
| 1994 | 2554 | 2334 | 0.914 | 1.263 | 1.766 | 1.303 | 2.393 | 0.153 |
| 1995 | 2586 | 2382 | 0.921 | 1.242 | 1.489 | 1.076 | 2.061 | 0.164 |
| 1996 | 2530 | 2257 | 0.892 | 0.854 | 0.822 | 0.581 | 1.163 | 0.175 |
| 1997 | 2599 | 2295 | 0.883 | 0.793 | 0.736 | 0.522 | 1.036 | 0.173 |
| 1998 | 2466 | 1926 | 0.781 | 0.301 | 0.190 | 0.130 | 0.279 | 0.193 |
| 1999 | 1525 | 1313 | 0.861 | 0.579 | 0.421 | 0.280 | 0.632 | 0.205 |
| 2000 | 2186 | 1788 | 0.818 | 0.548 | 0.354 | 0.240 | 0.521 | 0.196 |
| 2001 | 2227 | 1935 | 0.869 | 0.644 | 0.442 | 0.304 | 0.642 | 0.188 |
| 2002 | 2145 | 1892 | 0.882 | 0.611 | 0.482 | 0.333 | 0.698 | 0.186 |
| 2003 | 2085 | 1891 | 0.907 | 0.821 | 0.587 | 0.408 | 0.846 | 0.184 |
| 2004 | 2015 | 1842 | 0.914 | 0.922 | 0.629 | 0.440 | 0.897 | 0.179 |
| 2005 | 1692 | 1579 | 0.933 | 0.895 | 0.812 | 0.567 | 1.162 | 0.181 |
| 2006 | 1597 | 1433 | 0.897 | 0.699 | 0.561 | 0.381 | 0.824 | 0.195 |
| 2007 | 1629 | 1377 | 0.845 | 0.515 | 0.372 | 0.248 | 0.557 | 0.204 |
| 2008 | 2341 | 2013 | 0.860 | 0.748 | 0.667 | 0.470 | 0.948 | 0.177 |
| 2009 | 2618 | 2207 | 0.843 | 0.990 | 0.790 | 0.560 | 1.114 | 0.173 |
| 2010 | 1572 | 1418 | 0.902 | 1.031 | 0.860 | 0.591 | 1.252 | 0.189 |
| 2011 | 2587 | 2313 | 0.894 | 1.184 | 1.058 | 0.755 | 1.484 | 0.170 |
| 2012 | 2643 | 2352 | 0.890 | 0.788 | 0.656 | 0.467 | 0.921 | 0.171 |
| 2013 | 2700 | 2473 | 0.916 | 0.876 | 0.892 | 0.653 | 1.219 | 0.157 |
| 2014 | 2742 | 2569 | 0.937 | 0.859 | 0.948 | 0.707 | 1.270 | 0.147 |
| 2015 | 2714 | 2532 | 0.933 | 0.870 | 0.898 | 0.671 | 1.202 | 0.146 |
|  |  |  |  |  |  |  |  |  |


| Year | N | Positive N | PPT | Relative <br> Nominal <br> CPUE | Relative <br> Index | Lower <br> $95 \% \mathrm{CI}$ | Upper <br> $95 \% \mathrm{CI}$ | CV |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2016 | 2936 | 2789 | 0.950 | 0.911 | 0.957 | 0.726 | 1.262 | 0.139 |
| 2017 | 2858 | 2778 | 0.972 | 1.183 | 1.603 | 1.237 | 2.079 | 0.130 |

Table 7. Final deviance tables for the regressions for Vermilion Snapper in the Eastern U.S. Gulf of Mexico. The table shows the order of the factors as they were sequentially added to each model. Fit diagnostics listed for each factor were the diagnostics from a model that included that factor and all of the factors listed above it in the tables below. Note that variable in red was included during SEDAR45 but not included for SEDAR67 due to a percent deviance reduction < $1 \%$.

| Factor | DF | Deviance | Residual <br> DF | Residual <br> Deviance | AIC | Deviance <br> Reduced | Log <br> likelihood | Likelihood <br> Ratio Test |
| :--- | ---: | ---: | ---: | ---: | ---: | :--- | ---: | :--- |
| Binomial |  |  |  |  |  |  |  |  |
| Null | 1 | 48302 | 70328 | 48302 | 48303 | - | -24151 | - |
| Area | 2 | 43047 | 70327 | 5255 | 43047 | $10.88 \%$ | -21523 | 5255.8 |
| Year | 32 | 41274 | 70296 | 1772 | 41274 | $4.08 \%$ | -20637 | 1772.8 |
| Trip | 4 | 40763 | 70293 | 511 | 40763 | $1.23 \%$ | -20381 | 511.2 |
| Duration |  |  |  |  |  |  |  |  |
| Lognormal |  |  |  |  |  |  |  |  |
| Null | 1 | 98083 | 62696 | 98083 | 205983 | - | -102991 | - |
| Vessel | 102 | 77153 | 62595 | 20929 | 190935 | $21.21 \%$ | -95467 | 15048.2 |
| Year | 32 | 66452 | 62564 | 10700 | 181574 | $13.83 \%$ | -90787 | 9361.2 |
| Month | 12 | 65230 | 62553 | 1222 | 180409 | $1.82 \%$ | -90204 | 1164.2 |
| Red Snapper | 2 | 64610 | 62552 | 620 | 179810 | $0.95 \%$ | -89905 | 599 |
| Season | 2 |  |  |  |  |  |  |  |

## Figures



Figure 1. The difference between the number of records in the Western U.S. Gulf of Mexico in which Vermilion Snapper are observed and the number in which they are predicted to occur for each probability threshold (A). Histogram of probabilities generated by the species-based regression (B). The dashed vertical line indicates the critical value where false prediction is minimized.


Figure 2. Comparison of coefficients obtained from the Stephens and MacCall (2004) trip selection approach for SEDAR67 and the previous SEDAR45 assessment in the Western U.S. Gulf of Mexico.


Figure 3. Standardized indices with 95\% confidence intervals and nominal CPUE for Vermilion Snapper in the Western U.S. Gulf of Mexico.


Figure 4. Diagnostic plots for the binomial model for Vermilion Snapper in the Western U.S. Gulf of Mexico. Shown here are the predicted (solid line) and observed proportion of positive trips by year (A), and the residuals from the binomial model by year (B) and trip duration(C).


Figure 5. Diagnostic plots for the lognormal model of catch rates on positive trips for Vermilion Snapper in the Western U.S. Gulf of Mexico. Shown here are the frequency distribution of catch rates (A), the cumulative normalized residuals (B), and the distribution of residuals by year (C), red snapper season (D) and vessel (E). The red lines represent the expected normal distribution. Note vessel numbers have been excluded due to confidentiality.


Figure 6. Standardized index for Vermilion Snapper in the Western U.S. Gulf of Mexico for SEDAR67 compared to the index provided during SEDAR45. For comparison, both indices have been normalized by their respective means.


Figure 7. Comparison of index for Vermilion Snapper in the Western U.S. Gulf of Mexico for SEDAR67 compared to the index provided during SEDAR45 with confidence intervals.


Figure 8. The difference between the number of records in the Eastern U.S. Gulf of Mexico in which Vermilion Snapper are observed and the number in which they are predicted to occur for each probability threshold (A). Histogram of probabilities generated by the species-based regression (B). The dashed vertical line indicates the critical value where false prediction is minimized.


Figure 9. Comparison of coefficients obtained from the Stephens and MacCall (2004) trip selection approach for SEDAR67 and the previous SEDAR45 assessment in the Eastern U.S. Gulf of Mexico.


Figure 10. Standardized indices with $95 \%$ confidence intervals and nominal CPUE for Vermilion Snapper in the Eastern U.S. Gulf of Mexico.


Figure 11. Diagnostic plots for the binomial model for Vermilion Snapper in the Eastern U.S. Gulf of Mexico. Shown here are the predicted (solid line) and observed proportion of positive trips by year (A), and the residuals from the binomial model by year (B), area (C) and trip duration (D).



Figure 13. Standardized index for Vermilion Snapper in the Eastern U.S. Gulf of Mexico for SEDAR67 compared to the index provided during SEDAR45. For comparison, both indices have been normalized by their respective means.


Figure 14. Comparison of index for Vermilion Snapper in the Eastern U.S. Gulf of Mexico for SEDAR67 compared to the index provided during SEDAR45 with confidence intervals.

