# Preliminary standardized catch rates of Southeast US Atlantic red snapper (Lutjanus campechanus) from headboat logbook data 

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## SEDAR41-DW12

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Final index is found in the addendum.


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Preliminary standardized catch rates of Southeast US Atlantic red snapper (Lutjanus campechanus) from headboat logbook data

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July 222014

## *Addendum at end of document reflecting changes made at Data Workshop


#### Abstract

Standardized catch rates were generated from the Southeast headboat survey trip records (logbooks) from 1976-2009. The analysis included areas from central North Carolina through south Florida. Data filtering and subsetting steps were applied to the data to model trips that were likely to have directed red snapper effort. The preliminary decisions made prior to the data workshop are presented here. The final results of the headboat index will be presented in the SEDAR 41 Data Workshop Report.


## Background

The headboat fishery in the south Atlantic includes for-hire vessels. The fishery uses hook and line gear, generally targets hard bottom reefs as the fishing grounds, and generally targets multiple species in the snapper-grouper complex. One of the key characteristics defining a headboat from other recreational fishing such as charter boats is the number of anglers. Prior to 2000 headboats were defined as vessels carrying 15 or more recreational anglers. This criteria changed to 7 or more passengers in 2000 in the Atlantic (Ken Brennan, pers. comm. Dec. 2011).

Headboats in the south Atlantic are sampled from North Carolina to the Florida Keys.
Data have been collected since 1972, but logbook reporting did not start until 1973. In addition, only North Carolina and South Carolina were included in the earlier years of the data set. In 1976, data were collected from North Carolina, South Carolina, Georgia, and northern Florida, and starting in 1978, data were collected from southern Florida (Areas 1-17, Figure 1).

Variables reported in the data set include year, month, day, area, location, trip type, number of anglers, species, catch, and vessel id. Biological data and discard data were recorded for some trips in some years.

A 20" TL minimum size limit for red snapper has been in place since 1992. A 2 fish bag limit began in 1992. The red snapper fishery closed in 2010.

The headboat logbook index was used for SEDAR 24. Additional headboat records from 2010 to 2013 were examined to determine if sufficient data exists to extend this standardized index of abundance for south Atlantic red snapper. Due to the closure and potential effect on the index, these data were not considered.

## Data treatment

Data from area 1 (Figure 1) were excluded as this area was not recorded during most of the time series. The minimum number of anglers per vessel was set at 6 , which excluded the lower $0.1 \%$ of trips. These trips were excluded because they were possibly misreported and likely don't reflect the behavior of headboats in general.

## Subsetting trips

Trips to be included in the computation of the index need to be determined based on effort directed at red snapper. Effort can be determined directly for trips which had positive red snapper catches, but some trips likely directed effort at red snapper, but were unsuccessful at landing red snapper. Given that information on directed effort for trips without red snapper harvest is not available, another method must be used to compute total effort.
In order to determine effort that was likely directed at red snapper and which trips should be used to compute an index, the method of Stephens and MacCall (2004) was applied. The Stephens and MacCall method uses multiple logistic regression to estimate a probability for each trip that the focal species was caught, given other species caught on that trip. Species compositions differ across the south Atlantic; thus, the method was applied separately for two different regions: north (areas 2-10) and south (areas 11, 12, and 17; Shertzer et al. 2009). To avoid computation errors, the number of species in each analysis was limited to those species that occurred in $1 \%$ or more of trips. The most general model therefore included all species in the snapper-grouper complex which occurred in $1 \%$ or more of trips as main effects, excluding red porgy. Red porgy was removed because of regulation changes, which could erroneously remove trips likely to have caught red snapper in recent years. A backwards stepwise AIC procedure (Venables and Ripley 1997) was then used to perform further selection among possible species as predictor variables. In this procedure, a generalized linear model with Bernoulli response was used to relate presence/absence of red snapper in headboat trips to presence/absence of other species (Figure 2 - Figure 5).

## Model Input

## Response and explanatory variables

CPUE - catch per unit effort (CPUE) has units of fish/angler and was calculated as the number of red snapper caught divided by the number of anglers.

Year - Because year is the explanatory variable of interest, it was necessarily included in the analysis. A summary of the total number of trips with red snapper effort per year and area is provided in Table 1 and 2.

Area - Areas were pooled into regions of North Carolina ( $\mathrm{NC}=2,3,9,10$ ), South Carolina ( $\mathrm{SC}=4,5$ ), Georgia and North Florida (GNFL=6,7,8), and south Florida ( $\mathrm{sFL}=11,12,17$ ).

Season - The seasons were defined as winter (January, February, March), spring (April, May, June), summer (July, August, September) and fall (October, November, December).

Party - Five categories for the number of anglers on a boat were considered in the standardization process. The categories included: $\leq 20$ anglers, 20-40 anglers, 40-60 anglers, 6080 anglers, and $>80$ anglers. The minimum number of anglers per vessel was set at 6 , which excluded the lower $0.5 \%$ of trips. These trips were excluded because they were possibly misreported and likely don't reflect the behavior of headboats in general.

Trip Type - Trip types of half and full day trips were included in the analysis. Three-quarter day trips were pooled with half-day trips ( $<10 \%$ ). Multi-day trips were removed because most were in Florida and likely targeting deepwater species for some portion of the trip. The codes for first and second half-day trips designation for day and night trips were combined.

## Standardization

CPUE was modeled using the delta-glm approach (Lo et al. 1992; Dick 2004; Maunder and Punt 2004). In particular, fits of lognormal and gamma models were compared for positive CPUE. Also, the combination of predictor variables was examined to best explain CPUE patterns (both for positive CPUE and or positive CPUE). All analysis were performed in the R programming language, with much of the code adapted from Dick (2004).

## BERNOULLI SUBMODEL

One component of the delta-GLM is a logistic regression model that attempts to explain the probability of either catching or not catching red snapper on a particular trip. First, a model was fit with all main effects in order to determine which effects should remain in the binomial component of the delta-GLM. Stepwise AIC (Venables and Ripley1997) with a backwards selection algorithm was then used to eliminate those that did not improve model fit.

## POSITIVE CPUE SUBMODEL

Then, to determine predictor variables important for predicting positive CPUE, the positive portion of the model was fitted with all main effects using both the lognormal and gamma distributions. Stepwise AIC (Venables and Ripley1997) with a backwards selection algorithm
was then used to eliminate those that did not improve model fit. All predictor variables were modeled as fixed effects (and as factors rather than continuous variables).

Both components of the model were then fit together (with the code adapted from Dick 2004) using the lognormal and gamma distributions and compared them using AIC. With CPUE as the dependent variable.

Preliminary model diagnostics are presented in Figures 6-7.
It should be noted that the Stephens and MacCall method is most appropriate for species which have strong species associations. In other words, if a species is ubiquitous in the catch, or does not have well-defined effort, Stephens and MacCall may not work well to identify directed effort.

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Table 1. Proportion positive trips of red snapper in the south Atlantic Headboat fishery.

| Year | pos.RS.trips | HB.all.trips | $\begin{aligned} & \hline \% \\ & \text { pos } \end{aligned}$ |
| :---: | :---: | :---: | :---: |
| 1973 | 298 | 688 | 43\% |
| 1974 | 366 | 1182 | 31\% |
| 1975 | 421 | 1913 | 22\% |
| 1976 | 1033 | 3002 | 34\% |
| 1977 | 1228 | 3559 | 35\% |
| 1978 | 1803 | 4891 | 37\% |
| 1979 | 1460 | 8173 | 18\% |
| 1980 | 1577 | 11378 | 14\% |
| 1981 | 1416 | 11324 | 13\% |
| 1982 | 1283 | 12256 | 10\% |
| 1983 | 1642 | 12125 | 14\% |
| 1984 | 1493 | 11190 | 13\% |
| 1985 | 1908 | 11157 | 17\% |
| 1986 | 1605 | 13854 | 12\% |
| 1987 | 1758 | 13966 | 13\% |
| 1988 | 1683 | 11996 | 14\% |
| 1989 | 1411 | 10933 | 13\% |
| 1990 | 1335 | 11365 | 12\% |
| 1991 | 1070 | 10740 | 10\% |
| 1992 | 938 | 15007 | 6\% |
| 1993 | 1295 | 13894 | 9\% |
| 1994 | 1411 | 12575 | 11\% |
| 1995 | 1506 | 12275 | 12\% |
| 1996 | 1154 | 9060 | 13\% |
| 1997 | 649 | 6284 | 10\% |
| 1998 | 1250 | 9123 | 14\% |
| 1999 | 1386 | 7618 | 18\% |
| 2000 | 1430 | 7645 | 19\% |
| 2001 | 1602 | 6820 | 23\% |
| 2002 | 1516 | 5590 | 27\% |
| 2003 | 1225 | 5542 | 22\% |
| 2004 | 1558 | 6278 | 25\% |
| 2005 | 1379 | 5695 | 24\% |
| 2006 | 1177 | 5909 | 20\% |
| 2007 | 1326 | 6381 | 21\% |
| 2008 | 1770 | 9215 | 19\% |
| 2009 | 2134 | 10250 | 21\% |
| Total | 49750 | 366756 | 14\% |

Table 2. Number of red snapper headboat trips by area, positive and zero trips following Stephens \& MacCall (SM) method.

|  | Total Trips |  | SC | SF | Total | Positive Trips |  | SC | SF | Total | Proportion Positive |  | SC | SF | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | GF | NC |  |  |  | GF | NC |  |  |  | GF | NC |  |  |  |
| 1976 | 464 | 142 | 229 |  | 835 | 441 | 37 | 118 |  | 596 | 0.95 | 0.26 | 0.52 |  | 0.71 |
| 1977 | 608 | 57 | 208 |  | 873 | 542 | 30 | 69 |  | 641 | 0.89 | 0.53 | 0.33 |  | 0.73 |
| 1978 | 1132 | 144 | 249 | 3 | 1528 | 953 | 67 | 99 | 1 | 1120 | 0.84 | 0.47 | 0.40 | 0.33 | 0.73 |
| 1979 | 1028 | 163 | 78 | 28 | 1297 | 821 | 78 | 30 | 3 | 932 | 0.80 | 0.48 | 0.38 | 0.11 | 0.72 |
| 1980 | 1032 | 118 | 176 | 48 | 1374 | 787 | 50 | 104 | 10 | 951 | 0.76 | 0.42 | 0.59 | 0.21 | 0.69 |
| 1981 | 871 | 107 | 52 | 63 | 1093 | 772 | 69 | 27 | 28 | 896 | 0.89 | 0.64 | 0.52 | 0.44 | 0.82 |
| 1982 | 911 | 189 | 211 | 49 | 1360 | 733 | 108 | 110 | 4 | 955 | 0.80 | 0.57 | 0.52 | 0.08 | 0.70 |
| 1983 | 1212 | 173 | 208 | 54 | 1647 | 1005 | 91 | 109 | 7 | 1212 | 0.83 | 0.53 | 0.52 | 0.13 | 0.74 |
| 1984 | 1160 | 84 | 194 | 86 | 1524 | 915 | 37 | 130 | 21 | 1103 | 0.79 | 0.44 | 0.67 | 0.24 | 0.72 |
| 1985 | 1258 | 72 | 255 | 147 | 1732 | 1105 | 40 | 169 | 46 | 1360 | 0.88 | 0.56 | 0.66 | 0.31 | 0.79 |
| 1986 | 1591 | 98 | 264 | 184 | 2137 | 995 | 64 | 118 | 26 | 1203 | 0.63 | 0.65 | 0.45 | 0.14 | 0.56 |
| 1987 | 1564 | 106 | 306 | 171 | 2147 | 1048 | 44 | 149 | 23 | 1264 | 0.67 | 0.42 | 0.49 | 0.13 | 0.59 |
| 1988 | 1529 | 112 | 346 | 87 | 2074 | 902 | 64 | 196 | 15 | 1177 | 0.59 | 0.57 | 0.57 | 0.17 | 0.57 |
| 1989 | 1142 | 46 | 196 | 43 | 1427 | 855 | 19 | 128 | 6 | 1008 | 0.75 | 0.41 | 0.65 | 0.14 | 0.71 |
| 1990 | 1135 | 65 | 242 | 16 | 1458 | 828 | 19 | 161 | 1 | 1009 | 0.73 | 0.29 | 0.67 | 0.06 | 0.69 |
| 1991 | 1043 | 135 | 284 | 11 | 1473 | 695 | 45 | 138 | 1 | 879 | 0.67 | 0.33 | 0.49 | 0.09 | 0.60 |
| 1992 | 1612 | 245 | 231 | 62 | 2150 | 406 | 72 | 110 | 16 | 604 | 0.25 | 0.29 | 0.48 | 0.26 | 0.28 |
| 1993 | 1451 | 175 | 274 | 66 | 1966 | 420 | 81 | 217 | 17 | 735 | 0.29 | 0.46 | 0.79 | 0.26 | 0.37 |
| 1994 | 1167 | 181 | 233 | 44 | 1625 | 605 | 57 | 138 | 16 | 816 | 0.52 | 0.31 | 0.59 | 0.36 | 0.50 |
| 1995 | 1108 | 186 | 209 | 19 | 1522 | 620 | 57 | 103 | 5 | 785 | 0.56 | 0.31 | 0.49 | 0.26 | 0.52 |
| 1996 | 746 | 177 | 207 | 14 | 1144 | 445 | 42 | 66 | 6 | 559 | 0.60 | 0.24 | 0.32 | 0.43 | 0.49 |
| 1997 | 560 | 115 | 116 | 8 | 799 | 331 | 24 | 32 | 2 | 389 | 0.59 | 0.21 | 0.28 | 0.25 | 0.49 |
| 1998 | 1209 | 207 | 213 | 4 | 1633 | 692 | 30 | 80 | 1 | 803 | 0.57 | 0.14 | 0.38 | 0.25 | 0.49 |
| 1999 | 1301 | 177 | 208 | 1 | 1687 | 729 | 61 | 137 |  | 927 | 0.56 | 0.34 | 0.66 | 0.00 | 0.55 |
| 2000 | 1026 | 192 | 206 | 13 | 1437 | 672 | 59 | 86 | 7 | 824 | 0.65 | 0.31 | 0.42 | 0.54 | 0.57 |
| 2001 | 1079 | 162 | 285 | 11 | 1537 | 732 | 106 | 175 | 2 | 1015 | 0.68 | 0.65 | 0.61 | 0.18 | 0.66 |
| 2002 | 991 | 179 | 276 | 7 | 1453 | 687 | 100 | 205 | 1 | 993 | 0.69 | 0.56 | 0.74 | 0.14 | 0.68 |
| 2003 | 825 | 134 | 155 | 15 | 1129 | 558 | 49 | 111 |  | 718 | 0.68 | 0.37 | 0.72 | 0.00 | 0.64 |
| 2004 | 1059 | 219 | 288 | 30 | 1596 | 818 | 43 | 173 | 4 | 1038 | 0.77 | 0.20 | 0.60 | 0.13 | 0.65 |
| 2005 | 949 | 103 | 184 | 35 | 1271 | 776 | 7 | 87 | 8 | 878 | 0.82 | 0.07 | 0.47 | 0.23 | 0.69 |
| 2006 | 993 | 133 | 222 | 43 | 1391 | 687 | 14 | 70 | 13 | 784 | 0.69 | 0.11 | 0.32 | 0.30 | 0.56 |
| 2007 | 1085 | 94 | 280 | 40 | 1499 | 767 | 3 | 89 | 31 | 890 | 0.71 | 0.03 | 0.32 | 0.78 | 0.59 |
| 2008 | 1116 | 130 | 174 | 113 | 1533 | 985 | 23 | 68 | 31 | 1107 | 0.88 | 0.18 | 0.39 | 0.27 | 0.72 |
| 2009 | 1389 | 123 | 149 | 255 | 1916 | 1256 | 33 | 43 | 78 | 1410 | 0.90 | 0.27 | 0.29 | 0.31 | 0.74 |
| Grand Tot | 37346 | 4743 | 7408 | 1770 | 51267 | 25583 | 1723 | 3845 | 430 | 31581 | 0.69 | 0.36 | 0.52 | 0.24 | 0.62 |



Figure 1. Map of headboat sampling area definition. These areas were pooled into regions of North Carolina ( $\mathrm{NC}=2,3,9,10$ ), South Carolina ( $\mathrm{SC}=4,5$ ), Georgia and North Florida ( $\mathrm{GNFL}=6,7,8$ ), and south Florida ( $\mathrm{sFL}=11,12,17$ ).

Figure 2. Estimates of species-specific regression coefficients from Stephens and MacCall method applied to headboat data from areas in the northern region (excludes areas 11, 12, and 17), as used to estimate each trip's probability of catching the focal species.


Figure 3. Estimates of species-specific regression coefficients from Stephens and MacCall method applied to headboat data from areas in the southern region (includes areas 11, 12, and 17), as used to estimate each trip's probability of catching the focal species.


Figure 4. Absolute difference between observed and predicted number of positive trips from Stephens and MacCall method applied to headboat data from the northern region (excludes areas 11,12 , and 17). Left and right panels differ only in the range of probabilities shown.


Figure 5. Absolute difference between observed and predicted number of positive trips from Stephens and MacCall method applied to headboat data from the southern region (includes areas 11,12 , and 17). Left and right panels differ only in the range of probabilities shown.


Figure 6. CPUE binomial residuals for year, area, season, trip type and party size.


Standarized (quantile) residuals: (proportion positive)


Figure 6. Continued.


Standarized (quantile) residuals: (proportion positive)


Figure 7. The lognormal distribution of catch for the south Atlantic red snapper headboat logbook during 1976-2009.

Red Snapper pos headboat CPUE


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## ADDENDUM

Standardized catch rates of Southeast US Atlantic red snapper (Lutjanus campechanus) from headboat logbook data

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August 2015


#### Abstract

Standardized catch rates were generated from the Southeast headboat survey trip records (logbooks) from 1976-2009. The analysis included areas from central North Carolina through south Florida. Data filtering and subsetting steps were applied to the data to model trips that were likely to have directed red snapper effort.


## SEDAR 41 Index Working Group Review

The SEDAR 41 index working group (IWG) reviewed the methods used to develop an index of abundance for red snapper from headboat logbook data. The following topics were discussed at the data workshop and include the final decisions and justification.

## Headboat data evaluation

The SEDAR41 DW (August 2014) recommended the headboat index for use in the assessment. As part of that recommendation, the index was split in 1992. The justification for that split was a possible shift in angler behavior following implementation of the 20 " minimum size limit.

SEDAR 41 DW2 Scoping Calls recommended that the headboat index be reconsidered following the headboat data evaluation (SEDAR 41 DW46). As a result, several new data filters were applied ( $95 \%$ of these trips/vessels were filtered previously), and the data were examined for evidence to support splitting the index in 1992.

The headboat data evaluation showed that the number of red snapper landed versus the number sampled by year and region were highly correlated, particularly in the area off of Georgia-north Florida (Figure 8). Figure 9 illustrates the dockside sampling catch rate (red= red snapper $>20$ " TL, green= red snapper between $12 " \& 20 " \mathrm{TL}$, blue $=$ red snapper $<12 "$ ) plotted with the headboat logbook index (black line). These two figures illustrate that headboats landed legal size red snapper following the regulations in 1984 and 1992, but did not avoid red snapper. Before and after 1992, the species associated with red snapper were very similar (Figure 10 \& 11). In addition, general species associations in landings off Georgia-north Florida did not demonstrate any clear break in 1992 (Figure 12 \& 13). Figure 14 illustrate a hypothetical example of distinct clusters. The dockside sampling and additional analysis suggest a shift in selectivity in 1992, but not necessarily an abrupt change in catchability.

- Do not split the index. Rerun headboat logbook index 1976-2009 with recommended filters (SEDAR41-DW46)

Considerations for Assessment workshop

- Allow selectivity to change with time blocks of size limit regulations
- Consider modeling any perceived changes in catchability within the assessment model


## Start year

The index includes data from 1976 to 2009.

## End year

SEDAR 41 IWG participants along with fisherman present at the meeting discussed the red snapper closure in 2010 and its potential impact on the red snapper headboat logbook index in 2010-2014. Because of this shift in behavior (avoidance), the IWG recommended to end the red snapper headboat logbook index in 2009.

Subsetting technique- Stephens \& MacCall
A run using a 5\% cutoff was explored. Red snapper in the southern region did not meet this upper cutoff so the $1 \%$ was used in the final model run.

The following information represents the final dGLM results for the red snapper headboat logbook index (Table 3 \& Figures 15-17).

## Model Input

## Response and explanatory variables

CPUE - catch per unit effort (CPUE) has units of fish/angler and was calculated as the number of red snapper caught divided by the number of anglers.

Year-1976-2009
Area - Areas were pooled into regions of North Carolina ( $\mathrm{NC}=2,3,9,10$ ), South Carolina ( $\mathrm{SC}=4,5$ ), Georgia and North Florida (GNFL=6,7,8), and south Florida (sFL=11,12,17).

Season - The seasons were defined as winter (January, February, March), spring (April, May, June), summer (July, August, September) and fall (October, November, December).

Party - Five categories for the number of anglers on a boat were considered in the standardization process. The categories included: $\leq 20$ anglers, 20-40 anglers, 40-60 anglers, 6080 anglers, and $>80$ anglers. The minimum number of anglers per vessel was set at 6 , which excluded the lower $0.5 \%$ of trips. These trips were excluded because they were possibly misreported and likely don't reflect the behavior of headboats in general.

Trip Type - Trip types of half and full day trips were included in the analysis. Three-quarter day trips were pooled with half-day trips $(<10 \%)$. Multi-day trips were removed because most were in Florida and likely targeting deepwater species for some portion of the trip. The codes for first and second half-day trips designation for day and night trips were combined.

## Standardization

CPUE was modeled using the delta-glm approach (Lo et al. 1992; Dick 2004; Maunder and Punt 2004). In particular, fits of lognormal and gamma models were compared for positive CPUE. Also, the combination of predictor variables was examined to best explain CPUE patterns (both for positive CPUE and or positive CPUE). All analysis were performed in the R programming language, with much of the code adapted from Dick (2004).

## BERNOULLI SUBMODEL

One component of the delta-GLM is a logistic regression model that attempts to explain the probability of either catching or not catching red snapper on a particular trip. First, a model was fit with all main effects in order to determine which effects should remain in the binomial component of the delta-GLM. Stepwise AIC (Venables and Ripley1997) with a backwards selection algorithm was then used to eliminate those that did not improve model fit.

## POSITIVE CPUE SUBMODEL

Then, to determine predictor variables important for predicting positive CPUE, the positive portion of the model was fitted with all main effects using both the lognormal and gamma distributions. Stepwise AIC (Venables and Ripley1997) with a backwards selection algorithm
was then used to eliminate those that did not improve model fit. All predictor variables were modeled as fixed effects (and as factors rather than continuous variables).

Both components of the model were then fit together (with the code adapted from Dick 2004) using the lognormal and gamma distributions and compared them using AIC. With CPUE as the dependent variable.

The lognormal was the preferred model.

## LITERATURE CITED

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Figure 8. The number of Red Snapper landed versus the number sampled by year and region. The landings and fish sampled are scaled to their means to make them comparable, and the Spearman rank correlation coefficient $\rho$ values are provided in each region panel.


Figure 9. Biological profile (dockside sampling ) catch rate (red= red snapper $>20$ " TL max, green $=$ red snapper between $12 " \& 20 "$ TL max, blue $=$ red snapper $<12 "$ ) plotted with the headboat logbook index (black line) scaled to the mean catch rate from the sampled fish.


Figure 10. Estimates of species-specific regression coefficients from Stephens and MacCall method applied to headboat data from areas in the northern region. Left panel is from 1976-1991 data; right panel is from 1992-2009 data.


Figure 11. Estimates of species-specific regression coefficients from Stephens and MacCall method applied to headboat data from areas in the southern region. Left panel is from 1976-1991 data; right panel is from 1992-2009 data.


Figure 12. Non metric multidimensional scaling (NMDS) results. Annual shift in species assemblage from analysis using CPUE of the top snapper grouper species. The two colors represent the default number of medoids (clusters) used prior to final NMDS analysis ( $\mathrm{k}=2$ ).

dat.d
hclust (*, "complete")
Figure 13. Hierarchical cluster analysis results from CPUE of the top snapper grouper species from vessels north of Cape Canaveral, Florida.

Figure 14. Hypothetical example of distinct clustering with NMDS.



## Updated Results for Red Snapper Headboat Logbook Index (1976-2009)

Table 3. The relative nominal CPUE, number of trips, standardized index, and CV for the red snapper headboat logbook data in the south Atlantic from 1976-2009.

|  |  | Nominal | Relative | Standardized |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | N | CPUE | nominal | CPUE | CV |
| 1976 | 876 | 0.5523 | 2.6159 | 2.3652 | 0.0525 |
| 1977 | 900 | 0.4672 | 2.2129 | 2.1644 | 0.0800 |
| 1978 | 1576 | 0.4780 | 2.2640 | 2.1293 | 0.0295 |
| 1979 | 1293 | 0.4647 | 2.2009 | 2.2279 | 0.0503 |
| 1980 | 1409 | 0.3053 | 1.4458 | 1.4517 | 0.0461 |
| 1981 | 1092 | 0.5072 | 2.4023 | 2.9481 | 0.0427 |
| 1982 | 1347 | 0.2043 | 0.9676 | 1.2042 | 0.0519 |
| 1983 | 1579 | 0.3103 | 1.4695 | 1.6414 | 0.0536 |
| 1984 | 1477 | 0.3368 | 1.5953 | 1.4202 | 0.0285 |
| 1985 | 1741 | 0.3518 | 1.6661 | 2.0710 | 0.0477 |
| 1986 | 2185 | 0.1130 | 0.5351 | 0.4773 | 0.0660 |
| 1987 | 2199 | 0.1363 | 0.6453 | 0.5782 | 0.0462 |
| 1988 | 2061 | 0.1552 | 0.7349 | 0.5616 | 0.0575 |
| 1989 | 1438 | 0.1984 | 0.9396 | 0.9001 | 0.0457 |
| 1990 | 1468 | 0.1638 | 0.7758 | 0.8680 | 0.0557 |
| 1991 | 1463 | 0.1370 | 0.6488 | 0.6937 | 0.0444 |
| 1992 | 2156 | 0.0316 | 0.1497 | 0.0776 | 0.0950 |
| 1993 | 1981 | 0.0575 | 0.2721 | 0.1622 | 0.0830 |
| 1994 | 1633 | 0.0889 | 0.4210 | 0.2586 | 0.0450 |
| 1995 | 1523 | 0.0760 | 0.3600 | 0.2778 | 0.0638 |
| 1996 | 1130 | 0.0655 | 0.3103 | 0.2477 | 0.0646 |
| 1997 | 790 | 0.0641 | 0.3038 | 0.2662 | 0.0927 |
| 1998 | 1647 | 0.0626 | 0.2963 | 0.2427 | 0.0756 |
| 1999 | 1706 | 0.0779 | 0.3689 | 0.2891 | 0.0484 |
| 2000 | 1442 | 0.1026 | 0.4859 | 0.4120 | 0.0535 |
| 2001 | 1553 | 0.1712 | 0.8110 | 0.7569 | 0.0675 |
| 2002 | 1466 | 0.2278 | 1.0788 | 0.8778 | 0.0497 |
| 2003 | 1150 | 0.1249 | 0.5917 | 0.5154 | 0.0454 |
| 2004 | 1606 | 0.1631 | 0.7722 | 0.7641 | 0.0374 |
| 2005 | 1290 | 0.1447 | 0.6854 | 0.7582 | 0.0430 |
| 2006 | 1406 | 0.1124 | 0.5325 | 0.4330 | 0.0513 |
| 2007 | 1505 | 0.1089 | 0.5158 | 0.4369 | 0.0823 |
| 2008 | 1551 | 0.3209 | 1.5200 | 1.7092 | 0.0517 |
| 2009 | 1917 | 0.2966 | 1.4047 | 1.8121 | 0.0276 |
|  |  |  |  |  |  |

Figure 15. CPUE binomial residuals for year, area, season, trip type and party size 1976-2009.


Figure 16. The lognormal distribution and qq plot of catch for the south Atlantic red snapper headboat logbook during1976-2009.



Figure 17 The standardized and nominal CPUE index with error bars at (+/-) 2 standard deviations (nominal by area below) computed for red snapper in the south Atlantic using the headboat logbook data during 1976-2009.


