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# Standardized Catch Rates for Greater Amberjack from the Gulf of Mexico Headboat Fishery 1986-2012 

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## INTRODUCTION

The recreational fishery in the Gulf of Mexico is surveyed by the Marine Recreational Fishery Statistics Survey (MRFSS) conducted by NOAA Fisheries, the Texas Marine SportHarvest Monitoring Program conducted by the Texas Parks and Wildlife Department (TPWD), and the Headboat Survey (HBS) conducted by NOAA Fisheries. The HBS has monitored catch and effort from party (head) boats in the Gulf of Mexico since 1986. The purpose of this report is to outline the development of a standardized index of abundance for Gulf of Mexico greater amberjack using the HBS data.

## METHODS

## Headboat Survey

The HBS collects catch and effort data for individual headboat trips. Specific information such as the number of anglers, vessel identification, fishing area, trip type/duration (half, three-quarter, full, and multi-day trips), approximate time of day of fishing, fishing date, and catch by species in number and weight are collected as part of this program.

HBS data were used to characterize abundance trends of greater amberjack in the Gulf of Mexico. Catch per unit effort (CPUE) was calculated on an individual trip basis. CPUE was equal to the number of greater amberjack landed on a given trip divided by the effort, where effort was the product of the number of anglers and the total hours fished. A half-day fishing trip was assumed to be 5 hours, a three-quarter day trip was assumed to be seven hours, and a fullday trip was assumed to be 10 hours. A fishing day was assumed to be 12 hours for multi-day trips. Many individuals fish aboard headboats; therefore, total effort per trip was calculated as the product of the reported number of anglers and the assumed hours fished.

## Data preparation and filtering

The following data preparation and filtering techniques were applied to the HBS dataset:

1. Only full-day trips were retained.
2. HBS observations in the Gulf of Mexico were classified into three regions.
3. Selected trips that reached bag limits for greater amberjack were retained.
4. Trips during the closed season for greater amberjack were excluded.
5. Data from 2010 were excluded.
6. The Stephens MacCall (2004) approach was used to restrict the dataset to those trips that targeting greater amberjack.

The HBS dataset was looked at across different strata to assess the sample size of total trips and successful trips (trips that reported having caught greater amberjack) within each of the strata. Although reported headboat trips ranged in length from half a day to multiple days, trip length was observed to be confounded with region. For example, full day trips made up $93.2 \%$ of all trips in the NW TX region. Therefore, only full day trips were included in the analysis.

In the SEDAR9 benchmark and update assessments, HBS observations in the Gulf of Mexico were classified into five regions. The five regions were: 1) central and southwest Florida, 2) northwest Florida and Alabama, 3) Louisiana, 4) northeast Texas, and 5) central and south Texas. Since the Louisiana and central and southwest Florida regions each had multiple years without any positive trips, the 5 regions were aggregated into only 3 regions by taking into account geographic proximity and individual trends in CPUE over time. The regions used in this analysis were: 1) Florida, Alabama, and Louisiana, 2) northeast Texas, and 3) central and south Texas.

The management of greater amberjack is done by size limits, bag limits, and fishing seasons. The HBS data were explored to determine the number of trips that reached the cumulative bag limit. Between 1990 and 1996, when the bag limit was three fish per person per day, only $0.17 \%$ of positive trips either reached or exceeded the bag limit for greater amberjack. Between 1997 and 2012, when the bag limit was one fish per person per day, $2.76 \%$ of positive trips reached the bag limit while $0.93 \%$ of positive trips exceeded the bag limit. Given so few trips reached the bag limit, they were left in the database for analysis.

Fishing behavior was assumed to have been altered by the implementation of opened and closed seasons (see SEDAR33-RD05 for the management history of greater amberjack). Trips during the closed fishing seasons in 2011, and 2012 were removed from this analysis.

In 2010, there were significant area closures in the Gulf of Mexico from May to November that were related to the Deepwater Horizon/BP Oil Spill (SERO 2013). Total trips, total positive trips, and total landings from the FL, AL, and LA region during the summer (May July) of 2010 declined by $48 \%, 81 \%$, and $90 \%$, respectively, as compared to mean values over the previous three summers (2007-2009). As such, catch rates reported in the FL, AL, and LA region during the 2010 area closures may reflect temporary shifts in targeting and catchability. Since changes in headboat fishing behavior in response to the 2010 area closures are not accounted for in the standardization procedure, data from 2010 were excluded from the analysis.

Headboat trips can target any number of species on any given trip; therefore, species targeting is generally unknown. The Stephens-McCall approach (2004) was used to identify trips that targeted greater amberjack. This approach uses the species composition of each trip in a logistic regression of species presence/absence to infer if effort on that trip occurred in similar habitat to greater amberjack habitat. If effort on a trip was determined to occur in similar habitat to greater amberjack, or if a trip caught only greater amberjack, then that trip was used in the analysis.

## Standardization

Delta-lognormal modeling methods were used to estimate a relative index of abundance for greater amberjack (Lo et al. 1992). The main advantage of using this method is allowance for the probability of zero catch (Ortiz et al. 2000). The delta-lognormal modeling approach combines separate generalized linear model (GLM) analyses of the proportion of successful trips (trips that landed greater amberjack) and of the catch rates on successful trips to construct a single standardized CPUE index (Lo et al. 1992, Hinton and Maunder 2004, Maunder and Punt 2004).

For each GLM procedure of proportion positive trips, a type-3 model assuming a binomial error distribution was assumed and the logit link was selected. The response variable was the proportion of successful trips across strata. For the analysis of the catch rates on successful trips, a type- 3 model assuming lognormal error distribution was examined. A "normal" linking function was selected and the response variable was calculated as the natural log of CPUE.

A stepwise approach was used to quantify the relative importance of the explanatory factors. First, a GLM model was fit to the null model (only the intercept) and the AIC, deviance and degrees of freedom were calculated. Next, a suite of models was tested where each potential explanatory factor was added to the null model. Again, the AIC, deviance, and degrees of freedom were calculated. The model with the factor that had the lowest AIC became the new base model and the process was repeated by adding factors individually until either the AIC was no longer further reduced or the all the factors were added to the model. In addition to screening using AIC, factors were also screened and not added to the model if the reduction in deviance per degree of freedom was less than one percent. This screening was implemented in order to fit a more parsimonious model, given the fact that factors which reduce the deviance by so little exert little influence on the index trend.

Two-way interactions among significant main effects were examined. YEAR*FACTOR interaction terms were included in the model as random effects. The final delta-lognormal model was fit using a SAS macro, GLIMMIX (Russ Wolfinger, SAS Institute). To facilitate visual comparison, a relative standardized index and relative nominal CPUE series were calculated by dividing each value in the series by the mean value of the entire time-series.

The following factors were examined as possible influences on the proportion of positive trips, and on the catch rates of trips reporting the capture of greater amberjack:

| FACTOR | LEVELS | DESCRIPTION |
| :--- | :---: | :--- |
| YEAR | 26 | $1986-2009$ and 2011-2012 |
| REGION | 3 | Central and South West TX (Area codes 26-27), <br> Northwest TX (Area codes 25), <br> FL, AL, and LA (Area codes 21-22-23-24) |
| SEASON | 4 | Nov-Jan, Feb-Apr, May-July, Aug-Oct |
| NUMBER OF ANGLERS <br> (binomial component only) | 8 | Bins for number anglers: <br> $1-10,11-20,21-30,31-40, ~ 41-50, ~ 51-60, ~ 61-70, ~ 71+~$ |

## Results and discussion

The Stephens-MacCall approach was used to identify trips that targeted greater amberjack. The left panel of Figure 1 shows the critical probability which minimizes the difference between the predicted number and the observed number of trips greater amberjack. The right panel of Figure 1 shows the frequency of trips associated with the critical probability. Given these diagnostics, sufficient trips were retained in the database to develop a standardized index of abundance.

Various factors and first level interactions were tested for significance using the stepwise approach and accordingly included or excluded from the model. The following models resulted from the standardization procedures where Success is a binomial indicating whether or not a group of anglers caught greater amberjack, $\alpha$ represents the parameter estimate of each factor, $\mu$ represents the mean, and $\varepsilon$ represents the error term.

$$
\begin{gathered}
\text { Success }=\mu+(\text { Year }) \alpha_{1}+(\text { Region }) \alpha_{2}+(\text { Year } * \text { Region }) \alpha_{3}+\varepsilon \\
\ln (\text { CPUE })=\mu+(\text { Year }) \alpha_{1}+(\text { Region }) \alpha_{2}+(\text { Season }) \alpha_{3}+(\text { Year } * \text { Region }) \alpha_{4}+\varepsilon
\end{gathered}
$$

Although the interaction term between Year and Region was included in the binomial deviance analysis and in the GLM exercise, this interaction was not included in the final binomial model because the model would not converge.

Table 2 summarizes the standardized index and corresponding coefficients of variation, upper confidence limits, lower confidence limits, and nominal CPUE. Final deviance tables are included in Table 3.

Results for the greater amberjack headboat index standardization show that the highest value was at the start of the time series in 1986 and was followed by an overall decline through 1990. The index increased until 1992, declined until 1994, and then remained relatively stable through 2000. After 2000, the index increased until 2003 and then declined until 2005. The most recent years of the time series are marked by a brief and potentially spurious peak centered in 2008. In the last two years of the time series the index is stable (Figure 2).

The headboat index developed here for the SEDAR 33 assessment has similar trends as the headboat indices developed for the SEDAR9 benchmark and update assessments (Figure 3).

## Acknowledgements

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## Tables

Table 1. Stephens and MacCall regression coefficients for species occurring in at least $1 \%$ of headboat trips in the Gulf of Mexico.

| Species | Species Code | Estimate | Std. Error | z value | $\operatorname{Pr}(>\|\mathbf{z}\|)$ |
| :--- | :--- | :---: | :---: | :---: | :---: |
| Vermilion Snapper | sp10 | 1.03 | 0.03 | 38.87 | $<0.0001$ |
| Cobia | sp55 | 0.82 | 0.03 | 28.79 | $<0.0001$ |
| Scamp | sp30 | 0.69 | 0.03 | 23.03 | $<0.0001$ |
| Warsaw Grouper | sp23 | 0.43 | 0.04 | 11.46 | $<0.0001$ |
| Gray Snapper | sp18 | 0.30 | 0.03 | 11.21 | $<0.0001$ |
| Tomtate | sp51 | 0.15 | 0.04 | 3.87 | 0.0001 |
| Little Tunny | sp116 | 0.13 | 0.03 | 3.80 | 0.0001 |
| Red Snapper | sp11 | 0.03 | 0.03 | 1.09 | 0.2737 |
| King Mackerel | sp74 | 0.01 | 0.02 | 0.27 | 0.7845 |
| Gag Grouper | sp29 | -0.01 | 0.03 | -0.43 | 0.6641 |
| Dolphin | sp117 | -0.08 | 0.04 | -2.22 | 0.0265 |
| Lane Snapper | sp16 | -0.12 | 0.02 | -5.02 | $<0.0001$ |
| Gray Triggerfish | sp77 | -0.14 | 0.03 | -5.53 | $<0.0001$ |
| Red Grouper | sp22 | -0.14 | 0.04 | -4.03 | 0.0001 |
| Atlantic Sharpnose Shark | sp230 | -0.26 | 0.03 | -8.61 | $<0.0001$ |
| Black Sea Bass | sp33 | -0.58 | 0.08 | -7.52 | $<0.0001$ |
| Red Porgy | sp1 | -0.64 | 0.03 | -18.77 | $<0.0001$ |
| White Grunt | sp50 | -1.03 | 0.05 | -20.71 | $<0.0001$ |
| Intercept | Intercept | -1.74 | 0.03 | -54.94 | $<0.0001$ |

Table 2. Gulf of Mexico greater amberjack standardized index values, coefficients of variation, upper confidence limits, lower confidence limits, and nominal CPUE values from the headboat fishery.

| Year | Standardized Index | CV | Lower 95\% CI | Upper 95\% CI | Nominal CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 3.546 | 0.350 | 1.797 | 6.997 | 3.408 |
| 1987 | 1.774 | 0.384 | 0.845 | 3.724 | 1.778 |
| 1988 | 1.905 | 0.372 | 0.928 | 3.913 | 2.263 |
| 1989 | 1.493 | 0.385 | 0.710 | 3.139 | 1.494 |
| 1990 | 0.576 | 0.454 | 0.242 | 1.370 | 0.752 |
| 1991 | 0.728 | 0.433 | 0.318 | 1.668 | 0.791 |
| 1992 | 1.213 | 0.386 | 0.576 | 2.554 | 1.320 |
| 1993 | 0.735 | 0.401 | 0.340 | 1.591 | 0.641 |
| 1994 | 0.577 | 0.423 | 0.257 | 1.298 | 0.466 |
| 1995 | 0.681 | 0.416 | 0.306 | 1.513 | 0.534 |
| 1996 | 0.778 | 0.407 | 0.355 | 1.704 | 0.761 |
| 1997 | 0.597 | 0.446 | 0.255 | 1.399 | 0.526 |
| 1998 | 0.409 | 0.469 | 0.167 | 0.997 | 0.309 |
| 1999 | 0.547 | 0.493 | 0.215 | 1.390 | 0.576 |
| 2000 | 0.521 | 0.486 | 0.208 | 1.308 | 0.384 |
| 2001 | 0.916 | 0.426 | 0.405 | 2.073 | 0.878 |
| 2002 | 1.059 | 0.441 | 0.456 | 2.462 | 0.993 |
| 2003 | 1.425 | 0.417 | 0.640 | 3.172 | 1.230 |
| 2004 | 1.084 | 0.417 | 0.487 | 2.413 | 0.906 |
| 2005 | 0.482 | 0.470 | 0.197 | 1.179 | 0.389 |
| 2006 | 0.692 | 0.476 | 0.280 | 1.710 | 0.552 |
| 2007 | 0.420 | 0.486 | 0.167 | 1.054 | 0.436 |
| 2008 | 1.506 | 0.496 | 0.589 | 3.846 | 1.858 |
| 2009 | 0.729 | 0.445 | 0.311 | 1.705 | 0.987 |
| 2010 |  |  |  |  |  |
| 2011 | 0.865 | 0.540 | 0.314 | 2.381 | 0.898 |
| 2012 | 0.742 | 0.537 | 0.271 | 2.031 | 0.869 |

Table 3. Final deviance tables for the Gulf of Mexico greater amberjack regressions from the headboat fishery. The table shows the order of the factors as they were sequentially added to each model. Fit diagnostics listed for each factor were the diagnostics from a model that included that factor and all of the factors listed above it in the tables below. Although the interaction term between Year and Region (highlighted in gray) was included in the binomial deviance analysis and in the GLM exercise, this interaction was not included in the final model because it did not converge.

| Binomial |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Factor | DF | Deviance | $\begin{gathered} \text { Residual } \\ \text { Df } \end{gathered}$ | Residual Deviance | AIC | \% Deviance Reduced | Log likelihood | Likelihood Ratio Test |
| Null | 1 | 17001.60 | 12418 | 17001.60 | 17001.60 | - | -8500.80 | - |
| Year | 25 | 16371.30 | 12393 | 630.30 | 16371.20 | 3.71 | -8185.60 | 630.40 |
| Region | 2 | 16208.20 | 12391 | 163.10 | 16208.20 | 1.00 | -8104.10 | 163.00 |
| Year*Region | 50 | 15691.00 | 12341 | 517.20 | 15691.00 | 3.19 | -7845.50 | 517.20 |
| Lognormal |  |  |  |  |  |  |  |  |
| Factor | DF | Deviance | $\begin{gathered} \text { Residual } \\ \text { Df } \end{gathered}$ | Residual Deviance | AIC | \% Deviance Reduced | $\begin{gathered} \log \\ \text { likelihood } \end{gathered}$ | Likelihood Ratio Test |
| Null | 1 | 63799.50 | 12418 | 63799.50 | 55567.60 | - | -27783.80 |  |
| Region | 2 | 62307.50 | 12416 | 1492.00 | 55273.60 | 2.34 | -27636.80 | 294.00 |
| Year | 25 | 60905.70 | 12391 | 1401.80 | 54991.00 | 2.25 | -27495.50 | 282.60 |
| Season | 3 | 60238.60 | 12388 | 667.10 | 54854.20 | 1.10 | -27427.10 | 136.80 |
| Year*Region | 50 | 58777.20 | 12338 | 2128.50 | 54549.20 | 2.43 | -27274.60 | 441.80 |

## FIGURES



Figure 1. The left panel shows the difference between the number of records in which greater amberjack were observed and the number in which they were predicted. A critical value of 0.32 minimizes the difference. The right panel shows a histogram of the frequency of probabilities generated by the species regression.


Figure 2. Nominal CPUE, standardized index, and the $95 \%$ confidence intervals for the Gulf of Mexico greater amberjack from the headboat fishery. The standardized index and nominal CPUE values were normalized by their respective means over the time series.


Figure 3. Standardized headboat indices for Gulf of Mexico greater amberjack from the current assessment (SEDAR 33) and from previous assessments (SEDAR 9 and the SEDAR 9 update). Indices were normalized by their respective means during the overlapping period.

Appendix A: Diagnostic plots for the headboat index of Gulf of Mexico greater amberjack


Figure 4. Frequency distribution of catch rates on positive trips. The red line is the expected normal distribution.


Figure 5. Q-Q plot of CPUE.


Figure 6. Residuals from the binomial model on proportion of positive trips, by year (left) and by region (right).

## Appendix B: Number of total trip and trips that reported having caught greater amberjack across strata

Table 4. The total trips, number of positive trips, and percentage of positive trips by year for a full-day trips only dataset and for a Stephens and MacCall selected trips dataset, as well as the percentage of trips retained by the Stephens and MacCall trip selection procedure for the headboat fishery in the Gulf of Mexico.

|  |  |  | Percent of |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Total <br> Trips | Positive <br> Trips | Positive <br> Trips | Selected <br> Trips | Positive <br> Selected <br> Trips | Percent of <br> Positive <br> Selected <br> Trips | Percent of <br> Trips <br> Selected |
| 1986 | 2248 | 1012 | 45.02 | 509 | 367 | 72.10 | 22.64 |
| 1987 | 2666 | 961 | 36.05 | 523 | 299 | 57.17 | 19.62 |
| 1988 | 2829 | 821 | 29.02 | 518 | 317 | 61.20 | 18.31 |
| 1989 | 2468 | 863 | 34.97 | 427 | 258 | 60.42 | 17.30 |
| 1990 | 3178 | 418 | 13.15 | 496 | 182 | 36.69 | 15.61 |
| 1991 | 2882 | 444 | 15.41 | 558 | 209 | 37.46 | 19.36 |
| 1992 | 3265 | 764 | 23.40 | 661 | 320 | 48.41 | 20.25 |
| 1993 | 3398 | 699 | 20.57 | 758 | 302 | 39.84 | 22.31 |
| 1994 | 4011 | 659 | 16.43 | 804 | 256 | 31.84 | 20.04 |
| 1995 | 3071 | 640 | 20.84 | 707 | 267 | 37.77 | 23.02 |
| 1996 | 3229 | 579 | 17.93 | 691 | 277 | 40.09 | 21.40 |
| 1997 | 2036 | 333 | 16.36 | 474 | 180 | 37.97 | 23.28 |
| 1998 | 2535 | 347 | 13.69 | 597 | 167 | 27.97 | 23.55 |
| 1999 | 1752 | 218 | 12.44 | 374 | 127 | 33.96 | 21.35 |
| 2000 | 2438 | 363 | 14.89 | 469 | 142 | 30.28 | 19.24 |
| 2001 | 2104 | 410 | 19.49 | 466 | 208 | 44.64 | 22.15 |
| 2002 | 1765 | 461 | 26.12 | 417 | 194 | 46.52 | 23.63 |
| 2003 | 1548 | 470 | 30.36 | 373 | 202 | 54.16 | 24.10 |
| 2004 | 1803 | 441 | 24.46 | 371 | 206 | 55.53 | 20.58 |
| 2005 | 1943 | 310 | 15.95 | 454 | 167 | 36.78 | 23.37 |
| 2006 | 1790 | 324 | 18.10 | 416 | 156 | 37.50 | 23.24 |
| 2007 | 1709 | 296 | 17.32 | 495 | 156 | 31.52 | 28.96 |
| 2008 | 993 | 248 | 24.97 | 200 | 98 | 49.00 | 20.14 |
| 2009 | 1177 | 324 | 27.53 | 308 | 158 | 51.30 | 26.17 |
| 2010 | 865 | 169 | 19.54 |  |  |  |  |
| 2011 | 507 | 160 | 31.56 | 169 | 88 | 52.07 | 33.33 |
| 2012 | 694 | 200 | 28.82 | 184 | 91 | 49.46 | 26.51 |

Table 5. Total trips, number of positive trips and percentage of positive trips by year and region for headboat fishing trips in the Gulf of Mexico selected by the Stephens and MacCall trip selection procedure. Some data excluded from table due to confidentiality.

| Year | Selected Trips by Region |  |  | Positive Selected Trips by Region |  |  | Percent of Positive Selected Trips by Region |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CW TX, SW TX | NW TX | FL, AL, LA | CW TX, SW TX | NW TX | FL, AL, LA | CW TX, SW TX | NW TX | FL, AL, LA |
| 1986 | 56 | 293 | 160 | 42 | 207 | 118 | 75.00 | 70.65 | 73.75 |
| 1987 | 85 | 326 | 112 | 51 | 190 | 58 | 60.00 | 58.28 | 51.79 |
| 1988 | 82 | 265 | 171 | 53 | 161 | 103 | 64.63 | 60.75 | 60.23 |
| 1989 | 98 | 155 | 174 | 70 | 99 | 89 | 71.43 | 63.87 | 51.15 |
| 1990 | 148 | 167 | 181 | 107 | 48 | 27 | 72.30 | 28.74 | 14.92 |
| 1991 | 104 | 219 | 235 | - | - | 75 | - | - | 31.91 |
| 1992 | 89 | 285 | 287 | 70 | 87 | 163 | 78.65 | 30.53 | 56.79 |
| 1993 | 138 | 314 | 306 | 90 | 123 | 89 | 65.22 | 39.17 | 29.08 |
| 1994 | 137 | 284 | 383 | 74 | 115 | 67 | 54.01 | 40.49 | 17.49 |
| 1995 | 186 | 261 | 260 | 97 | 124 | 46 | 52.15 | 47.51 | 17.69 |
| 1996 | 117 | 293 | 281 | 49 | 115 | 113 | 41.88 | 39.25 | 40.21 |
| 1997 | 84 | 176 | 214 | 41 | 75 | 64 | 48.81 | 42.61 | 29.91 |
| 1998 | 84 | 264 | 249 | 35 | 84 | 48 | 41.67 | 31.82 | 19.28 |
| 1999 | 34 | 174 | 166 | 14 | 46 | 67 | 41.18 | 26.44 | 40.36 |
| 2000 | 69 | 237 | 163 | 32 | 79 | 31 | 46.38 | 33.33 | 19.02 |
| 2001 | 74 | 228 | 164 | 28 | 86 | 94 | 37.84 | 37.72 | 57.32 |
| 2002 | 151 | 212 | 54 | 63 | 110 | 21 | 41.72 | 51.89 | 38.89 |
| 2003 | 106 | 171 | 96 | 57 | 91 | 54 | 53.77 | 53.22 | 56.25 |
| 2004 | 128 | 163 | 80 | 70 | 97 | 39 | 54.69 | 59.51 | 48.75 |
| 2005 | 125 | 233 | 96 | 49 | 105 | 13 | 39.20 | 45.06 | 13.54 |
| 2006 | 139 | 227 | 50 | 62 | 77 | 17 | 44.60 | 33.92 | 34.00 |
| 2007 | 231 | 176 | 88 | 72 | 61 | 23 | 31.17 | 34.66 | 26.14 |
| 2008 | 41 | 39 | 120 | - | - | 62 | - | - | 51.67 |
| 2009 | 101 | 98 | 109 | 56 | 27 | 75 | 55.45 | 27.55 | 68.81 |
| 2010 |  |  |  |  |  |  |  |  |  |
| 2011 | 112 | 32 | 25 | 58 | 12 | 18 | 51.79 | 37.50 | 72.00 |
| 2012 | - | - | 37 | - | - | 26 | - | - | 70.27 |

Table 6. Total trips, number of positive trips and percentage of positive trips by year and season for headboat fishing trips in the Gulf of Mexico selected by the Stephens and MacCall trip selection procedure. Some data excluded from table due to confidentiality.

|  | Selected Trips by Season |  |  |  | Positive Selected Trips by Season |  |  |  |  | Percent of Positive Selected Trips by Season |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Winter | Spring | Summer | Fall | Winter | Spring | Summer | Fall | Winter | Spring | Summer | Fall |
| 1986 | 43 | 76 | 225 | 165 | 32 | 40 | 161 | 134 | 74.42 | 52.63 | 71.56 | 81.21 |
| 1987 | 50 | 108 | 227 | 138 | 24 | 50 | 136 | 89 | 48.00 | 46.30 | 59.91 | 64.49 |
| 1988 | 48 | 117 | 214 | 139 | 28 | 56 | 144 | 89 | 58.33 | 47.86 | 67.29 | 64.03 |
| 1989 | 33 | 82 | 167 | 145 | 17 | 37 | 104 | 100 | 51.52 | 45.12 | 62.28 | 68.97 |
| 1990 | 48 | 100 | 183 | 165 | 28 | 28 | 76 | 50 | 58.33 | 28.00 | 41.53 | 30.30 |
| 1991 | 68 | 92 | 223 | 175 | 37 | 29 | 93 | 50 | 54.41 | 31.52 | 41.70 | 28.57 |
| 1992 | 47 | 121 | 288 | 205 | 23 | 51 | 156 | 90 | 48.94 | 42.15 | 54.17 | 43.90 |
| 1993 | 46 | 171 | 279 | 262 | 18 | 67 | 128 | 89 | 39.13 | 39.18 | 45.88 | 33.97 |
| 1994 | 66 | 150 | 338 | 250 | 15 | 35 | 134 | 72 | 22.73 | 23.33 | 39.64 | 28.80 |
| 1995 | 63 | 142 | 273 | 229 | 22 | 54 | 120 | 71 | 34.92 | 38.03 | 43.96 | 31.00 |
| 1996 | 53 | 99 | 322 | 217 | 17 | 28 | 153 | 79 | 32.08 | 28.28 | 47.52 | 36.41 |
| 1997 | 23 | 101 | 246 | 104 | 7 | 33 | 105 | 35 | 30.43 | 32.67 | 42.68 | 33.65 |
| 1998 | 35 | 99 | 313 | 150 | 6 | 27 | 108 | 26 | 17.14 | 27.27 | 34.50 | 17.33 |
| 1999 | 11 | 100 | 192 | 71 | 3 | 21 | 78 | 25 | 27.27 | 21.00 | 40.63 | 35.21 |
| 2000 | 14 | 77 | 257 | 121 | 6 | 30 | 75 | 31 | 42.86 | 38.96 | 29.18 | 25.62 |
| 2001 | 6 | 49 | 262 | 149 | 4 | 19 | 138 | 47 | 66.67 | 38.78 | 52.67 | 31.54 |
| 2002 | 27 | 61 | 204 | 125 | 8 | 26 | 118 | 42 | 29.63 | 42.62 | 57.84 | 33.60 |
| 2003 | 12 | 63 | 196 | 102 | - | - | 120 | 51 | - | - | 61.22 | 50.00 |
| 2004 | 15 | 39 | 191 | 126 | 6 | 20 | 118 | 62 | 40.00 | 51.28 | 61.78 | 49.21 |
| 2005 | 5 | 54 | 280 | 115 | 3 | 15 | 101 | 48 | 60.00 | 27.78 | 36.07 | 41.74 |
| 2006 | 12 | 48 | 224 | 132 | 4 | 16 | 88 | 48 | 33.33 | 33.33 | 39.29 | 36.36 |
| 2007 | 8 | 62 | 291 | 134 | 5 | 24 | 94 | 33 | 62.50 | 38.71 | 32.30 | 24.63 |
| 2008 | 8 | 17 | 135 | 40 | 4 | 13 | 62 | 19 | 50.00 | 76.47 | 45.93 | 47.50 |
| 2009 | - | - | 204 | 74 | - | - | 105 | 36 | - | - | 51.47 | 48.65 |
| 2010 |  |  |  |  |  |  |  |  |  |  |  |  |
| 2011 | - | - | 28 | 74 | - | - | 15 | 40 | - | - | 53.57 | 54.05 |
| 2012 | 33 | 41 | 38 | 72 | 15 | 12 | 20 | 44 | 45.45 | 29.27 | 52.63 | 61.11 |

