# ALTERNATIVE CATCH RATE INDICES FOR RED SNAPPER (LUTJANUS CAMPECHANUS) LANDED DURING 1981-2003 BY THE U.S. RECREATIONAL FISHERY IN THE GULF OF MEXICO USING MRFSS AND TEXAS PARKS AND WILDLIFE DEPARTMENT DATA SETS 

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#### Abstract

Red snapper (Lutjanus campechanus) is an economically valuable species utilized by recreational and commercial fishermen in the southeastern United States. During 2004, NOAA Fisheries and the Gulf of Mexico Fisheries Management Council will conduct data, assessment and review workshops to complete a stock assessment of red snapper in the U.S. Gulf of Mexico. Abundance or catch rate indices are useful for most stock assessment procedures. For this purpose, catch rate indices for red snapper observed by recreational anglers in the U.S. Gulf of Mexico were constructed using Marine Recreational Fisheries Statistics Survey (MRFSS) and Texas Parks and Wildlife Department (TPWD) data sets. Delta-Poisson indices were constructed using two different approaches. The indices are quite similar. Both demonstrate the influence of strong year classes, and suggest higher catch rates of red snapper after 1990.


## INTRODUCTION

Red snapper is a valuable resource in the U.S. Gulf of Mexico. During 1998-2002, about 9 million pounds were landed annually within the U.S. Gulf of Mexico by commercial and recreational fishermen. While the value of the recreational fishery is difficult to quantify, it is estimated that Gulf wide, approximately 264,000 individual recreational trips target red snapper annually (Holiman, 1999). The commercial catch was valued at approximately $\$ 10$ million annually.

Red snapper are found in the western Atlantic Ocean and Gulf of Mexico, from Massachusetts to the Bay of Campeche, but are infrequent north of Cape Hatteras, NC (Hoese and Moore, 1998). Adults are common in submarine gullies and depressions, and over coral reefs, rock outcrops and gravel bottoms. They are most commonly found at depths of 40-110 meters ${ }^{1}$. Typically, red snapper reach a size of approximately 1000 mm TL, and weights up to

[^0]9.2 kg (Wilson and Nieland, 2001). Although ages in excess of 50 years have been observed, the vast majority of red snapper landed in the Gulf of Mexico are less than 15 years old (Wilson and Nieland, 2001).

In 2004, NOAA Fisheries personnel, Gulf of Mexico Fisheries Management Council (GMFMC) members, scientific experts, fishermen and other interested parties will participate in a series of workshops (Southeast Data Assessment and Review; SEDAR) to determine the status of the Gulf population of red snapper. To accomplish this objective, catch rate indices (or other abundance indices) are useful. This document describes the construction of alternative catch rate indices for the recreational fishery for red snapper in the U.S. Gulf of Mexico. These indices were created for formal consideration during the April 2004 SEDAR data workshop in New Orleans, LA. They are appropriate for use during stock assessment modeling procedures.

## METHODS

## Data Sources

NOAA Fisheries initiated the Marine Recreational Fisheries Statistics Survey (MRFSS) in 1979 in order to obtain standardized estimates of participation, effort, and catch by recreational fishermen in U.S. marine waters. MRFSS data is collected using two approaches: a telephone survey of households in coastal counties, and dockside interviews of fishermen (intercept survey). MRFSS intercept data was used for the construction of catch rate indices.

MRFSS intercept survey sampling coverage has varied over the time series. Initially, the survey covered shore fishing, as well as charter boat (CB), headboat (HB) and private boat (PB) fishing modes in all Gulf States. During 1982-1984, MRFSS discontinued sampling boat modes in Texas. This program was turned over to the Texas Park and Wildlife Department (TPWD) which began sampling Texas boat modes in the summer of 1983. Headboat sampling Gulf wide was transferred to the NOAA Fisheries Headboat Survey (HBS) program in 1986. TPWD continued to survey bay headboats until July, 1991. The MRFSS program no longer recommends the use of data collected during. 1979 and 1980 or wave 4 of 1981-1985. Therefore, these data were not included during the construction of catch rate indices ${ }^{2}$.

Index 1: Two catch rate indices were constructed. The first (hereafter referred to as Index 1) was intended to replicate the recreational index used during the most recent red snapper assessment (Schirripa and Legault, 1999) using a similar technique. This index was constructed using MRFSS intercept data from 1981-2003 and TPWD catch and effort data from 1983-1989. TPWD data was not included after 1989 because strict minimum size and bag limits were mandated in 1990 (Table 1). Unlike MRFSS data which includes fish landed and observed by the interviewer (A), dead fish not observed by the interviewer(B1; e.g., unavailable, filleted, used for bait, discarded dead at sea) and fish released alive (B2), TPWD data only records fish observed by the interviewer (A; presumably most landed fish were available for observation). TPWD data is not appropriate to combine with MRFSS intercept data after the 1990 regulations because the proportion of red snapper discarded by the recreational fishery may have increased significantly.

[^1]All HB, CB and PB trips that fished in "oceanic" areas using hook and line gear were included in the dataset used to construct the first standardized index of abundance. Shore mode and inshore fishing trips were excluded as they very seldom land red snapper. Table 2 summarizes the interviewed trips by year, state and fishing mode.

Index 2: A second index (hereafter referred to as Index 2) was constructed using only MRFSS intercept data. Like the previous approach, all HB, CB and PB trips that fished in "oceanic" areas using hook and line gear were initially included. These trips were examined to identify a list of species associated with red snapper in two Gulf regions, east (FL, AL, MS) and west (LA, TX). Then, trips were excluded if they did not catch a least one red snapper, or a species associate. The final data set was intended to exclude trips with a low probability of observing red snapper due to unrecorded covariates such as depth and location of fishing, bait choice, bottom type and gear configuration. Table 3 is a summary of the interviewed trips identified for index 2 by year, state and fishing mode. Texas was excluded from the analysis due to insufficient sampling ( $\mathrm{n}=59$ ).

The two sets of species associates (east and west) were identified using an association statistic proposed by Heinemann ${ }^{3}$. The association statistic was calculated for each species (species x) reported by $>50$ trips during 1981-2003 (Eq. 1).

$$
\begin{equation*}
\text { Association Statistic }=\frac{\# \text { Trips } \text { with Red Snapper and Species } X}{\# \text { Trips } \text { with } \operatorname{Re} d \text { Snapper }} / \frac{\# \text { Trips with Species } X}{\# \text { Total Trips }} \tag{1}
\end{equation*}
$$

The association statistic does not provide an objective critical value at which to include or exclude a species. A value of 1.0 implies that a given species co-occurs with red snapper exactly as often as random chance would predict. Values $>1.0$ indicate that a species co-occurs more often with red snapper than expected, and values $<1.0$ indicate that a given species co-occurs with red snapper less often than expected. For this analysis, a species was assumed to be associated with red snapper if its association statistic was $\geq 3.0$.

## Index Development

Index 1: Like the previous red snapper recreational index (Schirripa and Legault, 1999), the factors YEAR, MONTH, MODE (PB, HB, CB) and STATE (FL, AL, MS, LA, TX) were considered as possible influences on the probability of catching a red snapper, and the catch rates on positive trips.

Index 2: This analysis examined the influence of the factors YEAR, SEASON (Dec-Feb, MarMay, Jun-Aug and Sep-Nov), MODE (PB, HB, CB) and STATE (FL, AL, MS, LA) on the probability of catching a red snapper, and the catch rates on positive trips.

A delta-Poisson approach (Lo et al., 1992) was used to develop the standardized indices of abundance. This method combines separate generalized linear modeling (GLM) analyses of

[^2]the probability of success ${ }^{4}$ (trips that observed red snapper) and the catch on successful trips ${ }^{5}$ to construct a single standardized index of abundance. Parameterization of each model was accomplished using a GLM procedure (GENMOD; Version 8.02 of the SAS System for Windows © 2000. SAS Institute Inc. Cary, NC, USA). For the Poisson models, the response variable, catch, was calculated:
\[

$$
\begin{equation*}
\text { Catch }=A+B 1+B 2 \tag{2}
\end{equation*}
$$

\]

where $\mathrm{A}=$ fish observed, $\mathrm{B} 1=$ dead fish not observed and $\mathrm{B} 2=$ fish released alive. B 1 and B 2 catch, as well as effort (angler hours) were corrected for non-interviewed fishermen. When necessary, catch was rounded to the nearest whole number.

A forward stepwise approach was used during the construction of each GLM. First, a null model was fit. These results reflect the distribution of the nominal data. Next each potential factor was added to the null model individually, and the resulting reduction (\%RED) in deviance per degree of freedom (DEV/DF) was examined. The factor that caused the greatest reduction in deviance per degree of freedom was added to the base model if the factor was significant based upon a Chi-Square test (PROBCHISQ $\leq 0.05$ ), and the reduction in deviance per degree of freedom was $\geq 1 \%$. This model then became the base model, and the process was repeated, adding factors and two-way interaction terms individually until no factor or interaction met the criteria for incorporation into the final model. Higher order interaction terms were not examined.

The final delta-Poisson model was fit using a SAS macro, GLIMMIX (glmm800MaOB.sas: Russ Wolfinger, SAS Institute). All factors were modeled as fixed effects except two-way interaction terms containing YEAR (e.g. YEAR*STATE). These were modeled as random effects. To facilitate visual comparison, a relative index and relative nominal CPUE series were calculated by dividing each value in the series by the mean value of the series.

## RESULTS AND DISCUSSION

## Index 1

This index was intended to replicate and update the 1999 red snapper recreational index reported by Schirripa and Legault (1999). However, Index 1 deviates from a strict replication in that it was constructed using a binomial error structure for the analysis of proportion successful trips and a Poisson error structure for the analysis of the catch rates on successful trips. Also, TPWD after 1989 was excluded.

The stepwise construction of the binomial model on success is summarized in Table 4. The final model was:

$$
\text { SUCCESS }=\text { MODE }+ \text { STATE }+ \text { YEAR }+ \text { MONTH }+ \text { MONTH*STATE }+ \text { YEAR*STATE }
$$

[^3]Although the interaction terms YEAR*MONTH and MODE*STATE were significant, and reduced the $\mathrm{DEV} / \mathrm{DF}$ by $\geq 1 \%$, the validity of models containing these terms was questionable (negative of Hessian not positive definite). Therefore, these interaction terms were excluded.

Annual variations in the proportion of positive trips are summarized in Figure 1. The probable influence of a large year class is evident in 1983, but subsequently, the proportion of positive trips returned to about the 1981-1982 level during 1985-1996. An increase in the proportion of positive trips has occurred recently, from 1997-2003. Diagnostic plots were examined to evaluate the fit of the binomial model. The distributions of the chi-square residuals by the factors MODE, STATE, YEAR and MONTH (Fig 2A-D) indicate an acceptable fit. In general, the residuals are distributed evenly above and below zero, and show no trend in variance with year.

The stepwise construction of the Poisson model on catch during successful trips is summarized in Table 5. The final model was:

$$
\mathrm{CATCH}=\mathrm{YEAR}+\mathrm{STATE}+\mathrm{MODE}+\mathrm{MONTH}+\mathrm{YEAR} * \mathrm{MONTH}+\mathrm{YEAR} * S T A T E ~+~ Y E A R * M O D E
$$

The annual trend in nominal CPUE is shown in Figure 3. The trend in nominal CPUE is quite similar to the trend in proportion positive trips. A large increase occurred in 1983, but CPUE returned to low values in 1984. Higher CPUEs occur in 1991 and 1992 and from 1997 to the present. Diagnostic plots were examined to assess the fit of the Poisson model to catch on positive trips. The distributions of the residuals by the factors MODE, STATE, YEAR and MONTH (Fig 4A-D) indicate over-dispersion. (The Poisson model assumes that Mean = Variance. This assumption is violated.)

The delta-Poisson abundance index, with $95 \%$ confidence intervals, and the nominal CPUE series are shown in Figure 5. To facilitate visual comparison, each series was scaled to its respective mean. The index statistics can be found in Table 6. The standardized abundance index is roughly similar to the nominal CPUE series. Index 1 indicates that, for the recreational fishery, catch rates were lowest during 1984-1990. Since then, the catch rates have improved considerably. This result suggests the population of red snapper has increased since 1990.

## Index 2

Index 2 was constructed using a different approach which, in addition to excluding shore modes and inshore fishing, also excluded trips that did not catch red snapper, or a species associated with red snapper. Lists of the species associates identified for the eastern and western Gulf, and their association statistics are summarized in Tables 7 and 8.

The stepwise construction of the binomial model on success is summarized in Table 9 The final model was:

$$
\text { SUCCESS }=\text { STATE }+ \text { MODE }+ \text { YEAR }+ \text { SEASON }+ \text { SEASON*STATE }
$$

Although the interaction terms MODE*STATE, YEAR*STATE and YEAR*SEASON were significant, and reduced the $\mathrm{DEV} / \mathrm{DF}$ by $\geq 1 \%$, the validity of models containing these terms was
questionable (negative of Hessian not positive definite). Therefore, these interaction terms were excluded.

Annual variations in the proportion of positive trips are shown in Figure 6. In this case, the proportion of positive trips fluctuates around $45 \%$ until 1997 when it increases to $\sim 60 \%$. Then, the proportion of positive trips remains between $60 \%$ and $65 \%$ throughout the remainder of the time series. Diagnostic plots were examined to evaluate the fit of the binomial model. The distributions of the chi-square residuals by the factors MODE, STATE, YEAR and MONTH (Fig 7A-D) indicate an acceptable fit. In general, the residuals are distributed evenly above and below zero, and show no trend in variance with year.

The stepwise construction of the Poisson model on catch during positive trips is summarized in Table 10. The final model was:

$$
\mathrm{CATCH}=\mathrm{YEAR}+\mathrm{STATE}+\mathrm{MODE}+\mathrm{SEASON}+\mathrm{YEAR} * S T A T E+\text { YEAR } * S E A S O N+\text { YEAR } * M O D E
$$

The annual trend in nominal CPUE is shown in Figure 8. The trend in nominal CPUE is quite similar to the trend in proportion positive trips. The lowest values occur during the early part of the time series. A large increase in CPUE occurred in 1990, and CPUE remained high throughout the remainder of the time series. Diagnostic plots were examined to assess the fit of the Poisson model to catch on positive trips. The distributions of the residuals by the factors MODE, STATE, YEAR and MONTH (Fig 9A-D) indicate over-dispersion.

Index 2 results are summarized in Figure 10 and Table 11. The standardized abundance index is quite similar to the nominal CPUE series. Like Index 1, Index 2 results indicate an increase in the population of red snapper since 1990. Figure 11 summarizes all the indices discussed in this manuscript, including the index used in the previous assessment (Schirripa and Legault, 1999). Although the annual index values are not identical, all the indices have the same overall trend, and all agree that the population red snapper increased after 1990.

## LITERATURE CITED

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Table 1. History of management for the Gulf of Mexico recreational sector.
Changes in recreational red snapper size limits, bag limits, and season length.

| Year | Size Limit <br> (Inches TL) | Daily Bag Limit <br> (Number of Fish) $^{2}$ | Season length <br> (days) |
| :--- | :--- | :--- | :--- |
| 1984 | $13^{1}$ | no bag limit $^{2}$ | 365 |
| 1990 | 13 | 7 | 365 |
| 1994 | 14 | 7 | 365 |
| 1995 | 15 | 5 | 365 |
| 1996 | 15 | 5 | 365 |
| 1997 | 15 | 5 | $330^{3}$ |
| 1998 | 15 | $4^{6}$ | $272^{4}$ |
| 1999 | $15^{7}$ | 4 | $240^{5}$ |
| 2000 | 16 | 4 | 194 |
| 2001 | 16 | 4 | 194 |
| 2002 | 16 | 4 | 194 |
| 2003 | 16 | 4 | 194 |
|  |  |  |  |

${ }^{1}$ for-hire boats exempted until 1987
${ }^{2}$ Allowed to keep 5 undersized fish per day
${ }^{3}$ Fishery closed on November 27, 1997.
${ }^{4}$ Fishery closed on September 30, 1998.
${ }^{5}$ Fishery closed on August 29, 1999.
${ }_{7}^{6}$ Bag limit was 5 fish from January through April, 1998.
${ }^{7}$ Size limit was 18 inches from June 4 through August 29, 1999.

Table 2. Total trips in the analysis dataset used for Index 1. Private boat (PB), charter boat (CB) and headboat (HB) trips fishing using hook and line in "oceanic" areas were included.

|  |  |  |  | MRFSS |  |  | TPWD |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | STATE |  |  | STATE |
| YEAR | MODE | AL | FL | LA | MS | TX | TX |
| 1981 | CB | 28 | 137 | 44 | 17 | 4 | 0 |
|  | HB | 6 | 314 | 5 | 0 | 25 | 0 |
|  | PB | 102 | 680 | 163 | 69 | 102 | 0 |
| 1982 | CB | 31 | 60 | 17 | 69 | 0 | 0 |
|  | HB | 51 | 202 | 24 | 7 | 0 | 0 |
|  | PB | 468 | 1276 | 250 | 293 | 0 | 0 |
| 1983 | CB | 48 | 107 | 161 | 39 | 0 | 53 |
|  | HB | 90 | 687 | 70 | 0 | 0 | 66 |
|  | PB | 206 | 616 | 136 | 111 | 0 | 584 |
| 1984 | CB | 65 | 156 | 109 | 107 | 0 | 36 |
|  | HB | 137 | 697 | 27 | 0 | 0 | 67 |
|  | PB | 231 | 698 | 276 | 136 | 0 | 1015 |
| 1985 | CB | 79 | 145 | 24 | 62 | 40 | 42 |
|  | HB | 34 | 693 | 46 | 0 | 194 | 0 |
|  | PB | 260 | 874 | 320 | 61 | 198 | 1097 |
| 1986 | CB | 117 | 790 | 190 | 107 | 0 | 29 |
|  | PB | 416 | 2769 | 1555 | 80 | 0 | 699 |
| 1987 | CB | 125 | 599 | 112 | 135 | 0 | 37 |
|  | PB | 618 | 4141 | 579 | 294 | 0 | 691 |
| 1988 | CB | 132 | 551 | 48 | 160 | 0 | 24 |
|  | PB | 247 | 3433 | 413 | 320 | 0 | 637 |
| 1989 | CB | 126 | 323 | 39 | 157 | 0 | 22 |
|  | PB | 239 | 2163 | 367 | 145 | 0 | 509 |
| 1990 | CB | 50 | 229 | 132 | 158 | 0 | 0 |
|  | PB | 279 | 1772 | 350 | 199 | 0 | 0 |
| 1991 | CB | 164 | 277 | 84 | 109 | 0 | 0 |
|  | PB | 359 | 1598 | 448 | 261 | 0 | 0 |
| 1992 | CB | 192 | 514 | 140 | 223 | 0 | 0 |
|  | PB | 484 | 3973 | 647 | 351 | 0 | 0 |
| 1993 | CB | 85 | 542 | 80 | 113 | 0 | 0 |
|  | PB | 375 | 3381 | 343 | 145 | 0 | 0 |
| 1994 | CB | 100 | 528 | 59 | 86 | 0 | 0 |
|  | PB | 407 | 4010 | 268 | 103 | 0 | 0 |
| 1995 | CB | 88 | 393 | 66 | 38 | 0 | 0 |
|  | PB | 336 | 3654 | 304 | 60 | 0 | 0 |
| 1996 | CB | 109 | 458 | 73 | 81 | 0 | 0 |
|  | PB | 647 | 4295 | 221 | 128 | 0 | 0 |
| 1997 | CB | 132 | 920 | 93 | 97 | 0 | 0 |
|  | PB | 615 | 4103 | 491 | 287 | 0 | 0 |
| 1998 | CB | 176 | 1734 | 110 | 162 | 0 | 0 |
|  | PB | 575 | 4288 | 180 | 208 | 0 | 0 |
| 1999 | CB | 322 | 3728 | 73 | 238 | 0 | 0 |
|  | PB | 809 | 6156 | 304 | 348 | 0 | 0 |
| 2000 | CB | 303 | 3960 | 95 | 252 | 0 | 0 |
|  | PB | 693 | 4815 | 225 | 176 | 0 | 0 |
| 2001 | CB | 262 | 2987 | 104 | 122 | 0 | 0 |
|  | PB | 759 | 5530 | 200 | 86 | 0 | 0 |
| 2002 | CB | 260 | 3216 | 152 | 82 | 0 | 0 |
|  | PB | 614 | 5931 | 302 | 83 | 0 | 0 |
| 2003 | CB | 189 | 3749 | 105 | 133 | 0 | 0 |
|  | PB | 621 | 5244 | 146 | 61 | 0 | 0 |

Table 3. Total trips (hook and line only) in the analysis dataset used for Index 2. Trips by fishing modes other than HB, PB and CB and "inshore" fishing trips were excluded. Trips were also excluded if they did not land at least one red snapper or a species associate. Finally, TX trips were excluded due to insufficient sampling.

|  |  |  |  | STATE |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | MODE | AL | FL | LA | MS | TX |
| 1981 | CB | 9 | 26 | 28 | 0 | 0 |
|  | HB | 6 | 77 | 5 | 0 | 6 |
|  | PB | 10 | 75 | 26 | 1 | 25 |
| 1982 | CB | 17 | 24 | 8 | 1 | 0 |
|  | HB | 50 | 101 | 19 | 0 | 0 |
|  | PB | 17 | 109 | 58 | 4 | 0 |
| 1983 | CB | 36 | 35 | 113 | 1 | 0 |
|  | HB | 90 | 273 | 57 | 0 | 0 |
|  | PB | 20 | 38 | 36 | 1 | 0 |
| 1984 | CB | 34 | 33 | 87 | 1 | 0 |
|  | HB | 105 | 189 | 17 | 0 | 0 |
|  | PB | 14 | 43 | 29 | 1 | 0 |
| 1985 | CB | 59 | 33 | 8 | 1 | 0 |
|  | HB | 32 | 219 | 23 | 0 | 19 |
|  | PB | 7 | 45 | 36 | 2 | 9 |
| 1986 | CB | 32 | 328 | 93 | 7 | 0 |
|  | PB | 20 | 171 | 58 | 1 | 0 |
| 1987 | CB | 64 | 202 | 45 | 2 | 0 |
|  | PB | 27 | 282 | 34 | 8 | 0 |
| 1988 | CB | 74 | 150 | 12 | 11 | 0 |
|  | PB | 17 | 253 | 43 | 6 | 0 |
| 1989 | CB | 78 | 93 | 10 | 12 | 0 |
|  | PB | 17 | 220 | 67 | 4 | 0 |
| 1990 | CB | 40 | 66 | 39 | 6 | 0 |
|  | PB | 36 | 88 | 55 | 20 | 0 |
| 1991 | CB | 106 | 99 | 43 | 14 | 0 |
|  | PB | 63 | 122 | 46 | 11 | 0 |
| 1992 | CB | 178 | 153 | 64 | 66 | 0 |
|  | PB | 117 | 302 | 87 | 69 | 0 |
| 1993 | CB | 72 | 148 | 30 | 20 | 0 |
|  | PB | 80 | 228 | 53 | 25 | 0 |
| 1994 | CB | 91 | 174 | 34 | 24 | 0 |
|  | PB | 79 | 227 | 53 | 22 | 0 |
| 1995 | CB | 75 | 54 | 25 | 11 | 0 |
|  | PB | 81 | 151 | 45 | 9 | 0 |
| 1996 | CB | 82 | 73 | 22 | 13 | 0 |
|  | PB | 90 | 228 | 36 | 14 | 0 |
| 1997 | CB | 113 | 252 | 29 | 21 | 0 |
|  | PB | 111 | 191 | 78 | 37 | 0 |
| 1998 | CB | 149 | 480 | 42 | 26 | 0 |
|  | PB | 95 | 247 | 42 | 33 | 0 |
| 1999 | CB | 218 | 806 | 33 | 38 | 0 |
|  | PB | 199 | 311 | 85 | 19 | 0 |
| 2000 | CB | 236 | 971 | 39 | 28 | 0 |
|  | PB | 182 | 208 | 75 | 7 | 0 |
| 2001 | CB | 189 | 733 | 34 | 16 | 0 |
|  | PB | 214 | 363 | 53 | 13 | 0 |
| 2002 | CB | 208 | 779 | 96 | 23 | 0 |
|  | PB | 203 | 377 | 56 | 40 | 0 |
| 2003 | CB | 153 | 987 | 67 | 38 | 0 |
|  | PB | 162 | 345 | 42 | 30 | 0 |

Table 4. A summary of formulation of the binomial model for INDEX 1. Factors were added to the model if PROBCHISQ $\leq 0.05$ and the reduction in DEV/DF (\%RED) $\geq 1.0 \%$ (bold blue font).

| There are no explanatory FACTOR | actors in DEGF | n the base DEVIANCE | ode1 (NU DEV/DF | NULL MODEL). \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BASE | 141656 | 80891.2 | 0.5710 |  | -40445.6 |  |  |
| MODE | 141654 | 71462.7 | 0.5045 | $5 \quad 11.65$ | -35731.4 | 9428.49 | <0.0001 |
| STATE | 141652 | 74396.4 | 0.5252 | 8.03 | -37198.2 | 6494.85 | <0.0001 |
| YEAR | 141634 | 79444.3 | 0.5609 | 1.77 | -39722.1 | 1446.93 | <0.0001 |
| MONTH | 141645 | 80131.2 | 0.5657 | 0.93 | -40065.6 | 759.98 | <0.0001 |
| The explanatory factors in | the base | e mode1 are: | MODE |  |  |  |  |
| FACTOR | DEGF | DEVIANCE | DEV/DF | \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 141654 | 71462.7 | 0.5045 |  | -35731.4 |  |  |
| STATE | 141650 | 63150.6 | 0.4458 | $8 \quad 11.63$ | -31575.3 | 8312.09 | $<0.0001$ |
| MONTH | 141643 | 70067.1 | 0.4947 | 1.95 | -35033.5 | 1395.67 | <0.0001 |
| YEAR | 141632 | 70979.0 | 0.5012 | 0.66 | -35489.5 | 483.78 | <0.0001 |
| The explanatory factors in FACTOR | the base DEGF | mode1 are DEVIANCE | $\begin{aligned} & \text { MODE S } \\ & \text { DEV/DF } \end{aligned}$ | STATE <br> \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 141650 | 63150.6 | 0.4458 |  | -31575.3 |  |  |
| YEAR | 141628 | 62170.2 | 0.4390 | $0 \quad 1.54$ | -31085.1 | 980.41 | $<0.0001$ |
| MONTH_CHAR | 141639 | 62311.7 | 0.4399 | 1.32 | -31155.8 | 838.95 | <0.0001 |
| The explanatory factors in FACTOR | the base DEGF | mode1 are: DEVIANCE | MODE | STATE YEAR \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 141628 | 62170.2 | 0.4390 |  | -31085.1 |  |  |
| MONTH | 141617 | 61148.5 | 0.4318 | $8 \quad 1.64$ | -30574.2 | 1021.78 | <0.0001 |
| The explanatory factors in FACTOR | the base DEGF | mode1 are: DEVIANCE | MODE ST DEV/DF | STATE YEAR MONTH \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 141617 | 61148.5 | 0.4318 |  | -30574.2 |  |  |
| MONTH*STATE | 141573 | 59418.1 | 0.4197 | 7 2.80 | -29709.0 | 1730.40 | $<0.0001$ |
| YEAR*STATE | 141544 | 59719.7 | 0.4219 | 2.29 | -29859.9 | 1428.72 | $<0.0001$ |
| MONTH*MODE | 141595 | 60459.2 | 0.4270 | 1.11 | -30229.6 | 689.22 | <0.0001 |
| YEAR*MODE | 141591 | 60875.7 | 0.4299 | 0.43 | -30437.9 | 272.72 | <0.0001 |
| The explanatory factors in FACTOR | the base DEGF | mode1 are: DEVIANCE | MODE DEV/DF | STATE YEAR MONTH \%REDUCTION | MONTH*STATE LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 141573 | 59418.1 | 0.4197 |  | -29709.0 |  |  |
| YEAR*STATE | 141500 | 58108.3 | 0.4107 | $7 \quad 2.15$ | -29054.2 | 1309.73 | <0.0001 |
| MONTH*MODE | 141551 | 59062.6 | 0.4173 | 0.58 | -29531.3 | 355.46 | <0.0001 |
| YEAR*MODE | 141547 | 59158.2 | 0.4179 | 0.42 | -29579.1 | 259.90 | <0.0001 |
| The explanatory factors in | the base | e mode1 are: | MODE | STATE YEAR MONTH | MONTH*STATE | YEAR*STATE |  |
| FACTOR | DEGF | DEVIANCE | DEV/DF | \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 141500 | 58108.3 | 0.4107 |  | -29054.2 |  |  |
| MONTH*MODE | 141478 | 57764.4 | 0.4083 | 0.58 | -28882.2 | 343.88 | <0.0001 |
| YEAR*MODE | 141474 | 57852.6 | 0.4089 | 0.42 | -28926.3 | 255.70 | <0.0001 |

Table 5. A summary of formulation of the Poisson model for INDEX 1. Factors were added to the model if PROBCHISQ $\leq 0.05$ and the reduction in DEV/DF (\%RED) $\geq 1.0 \%$ (bold blue font).


| The explanatory factors in FACTOR | the bas DEGF | mode1 a DEVIANCE | YEAR S DEV/DF | STATE MODE MONTH \%REDUCTION | $\begin{aligned} & \text { YEAR*MONTH } \\ & \text { LOGLIKE } \end{aligned}$ | $\begin{gathered} \text { YEAR*STATE } \\ \text { CHISQ } \end{gathered}$ | PROBCHISQ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BASE | 11382 | 192757.8 | 16.9353 |  | 850420.1 |  |  |
| YEAR*MODE | 11356 | 190132.2 | 16.7429 | -1.14 | 851732.9 | 2625.55 | <0.0001 |
| MODE*STATE | 11375 | 190912.8 | 16.7835 | 0.90 | 851342.6 | 1844.94 | <0.0001 |
| MONTH*STATE | 11338 | 190645.4 | 16.8147 | 0.71 | 851476.3 | 2112.34 | <0.0001 |
| MONTH*MODE | 11360 | 191170.5 | 16.8284 | 0.63 | 851213.8 | 1587.30 | <0.0001 |


| The explanatory factors i FACTOR | the ba DEGF | e mode1 are DEVIANCE | YEAR DEV/DF | STATE MODE MONTH \%REDUCTION | $\begin{aligned} & \text { YEAR*MONTH } \\ & \text { LOGLIKE } \end{aligned}$ | $\begin{gathered} \text { YEAR*STATE } \\ \text { CHISQ } \end{gathered}$ | YEAR*MODE PROBCHISQ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BASE | 11356 | 190132.2 | 16.7429 |  | 851732.9 |  |  |
| MODE*STATE | 11349 | 188411.7 | 16.6016 | 0.84 | 852593.2 | 1720.55 | <0.0001 |
| MONTH*STATE | 11312 | 187938.9 | 16.6141 | 0.77 | 852829.6 | 2193.30 | <0.0001 |
| MONTH*MODE | 11334 | 188783.1 | 16.6564 | 0.52 | 852407.5 | 1349.14 | <0.0001 |

Table 6. Relative nominal CPUE, number of trips, number of positive trips, proportion positive trips (PPT) and abundance index statistics (INDEX 1).

| YEAR | TRIPS | POS <br> TRIPS | PPT | Rel. <br> Nominal <br> CPUE | Rel. <br> Index | Lower <br> $\mathbf{9 5 \%} \mathbf{C I}$ | Upper <br> 95\% CI | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 1696 | 105 | 0.062 | 0.834 | 0.472 | -0.197 | 1.142 | 0.724 |
| 1982 | 2748 | 177 | 0.064 | 0.405 | 0.365 | -0.159 | 0.888 | 0.733 |
| 1983 | 2974 | 561 | 0.189 | 2.397 | 1.033 | 0.144 | 1.922 | 0.439 |
| 1984 | 3757 | 353 | 0.094 | 0.921 | 0.429 | -0.103 | 0.961 | 0.632 |
| 1985 | 4169 | 296 | 0.071 | 0.642 | 0.285 | -0.149 | 0.719 | 0.777 |
| 1986 | 6752 | 439 | 0.065 | 0.513 | 0.380 | -0.110 | 0.870 | 0.658 |
| 1987 | 7331 | 341 | 0.047 | 0.357 | 0.401 | -0.115 | 0.917 | 0.656 |
| 1988 | 5965 | 301 | 0.050 | 0.353 | 0.480 | -0.102 | 1.062 | 0.618 |
| 1989 | 4090 | 234 | 0.057 | 0.394 | 0.416 | -0.120 | 0.952 | 0.657 |
| 1990 | 3169 | 153 | 0.048 | 0.461 | 0.573 | -0.157 | 1.302 | 0.650 |
| 1991 | 3300 | 265 | 0.080 | 1.371 | 1.176 | 0.028 | 2.324 | 0.498 |
| 1992 | 6524 | 543 | 0.083 | 1.421 | 2.064 | 0.456 | 3.671 | 0.397 |
| 1993 | 5064 | 325 | 0.064 | 0.914 | 1.455 | 0.186 | 2.723 | 0.445 |
| 1994 | 5561 | 339 | 0.061 | 0.843 | 1.445 | 0.212 | 2.678 | 0.435 |
| 1995 | 4939 | 210 | 0.043 | 0.486 | 0.958 | -0.058 | 1.974 | 0.541 |
| 1996 | 6012 | 271 | 0.045 | 0.763 | 0.948 | -0.024 | 1.920 | 0.523 |
| 1997 | 6738 | 522 | 0.077 | 1.436 | 1.694 | 0.299 | 3.090 | 0.420 |
| 1998 | 7433 | 676 | 0.091 | 1.354 | 1.378 | 0.198 | 2.559 | 0.437 |
| 1999 | 11978 | 1093 | 0.091 | 1.435 | 1.384 | 0.207 | 2.561 | 0.434 |
| 2000 | 10519 | 1211 | 0.115 | 1.508 | 1.181 | 0.119 | 2.244 | 0.459 |
| 2001 | 10050 | 992 | 0.099 | 1.217 | 1.128 | 0.090 | 2.166 | 0.469 |
| 2002 | 10640 | 1158 | 0.109 | 1.566 | 1.764 | 0.411 | 3.117 | 0.391 |
| 2003 | 10248 | 1162 | 0.113 | 1.410 | 1.591 | 0.338 | 2.843 | 0.402 |

Table 7. Results of calculations used to identify species associated with red snapper in the eastern GOM (FL,AL,MS). Species were assumed to be associated with red snapper if the association statistic was $\geq 3.0$. $\% \mathrm{CO}$ is the percent common occurrence.

| Common Name | Scientific Name | Trips with Red Snapper and Species X | Trips with Species X | Total Red <br> Snapper <br> Trips | Total Trips | Association Statistic | \%CO |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Red snapper | Lutjanus campechanus | 9773 | 9773 | 9773 | 91666 | 9.38 | 100.0 |
| Banded rudderfish | Seriola zonata | 282 | 344 | 9773 | 91666 | 7.69 | 82.0 |
| Whitebone porgy | Calamus leucosteus | 208 | 266 | 9773 | 91666 | 7.33 | 78.2 |
| Red porgy | Pagrus pagrus | 1567 | 2036 | 9773 | 91666 | 7.22 | 77.0 |
| Vermilion snapper | Rhomboplites aurorubens | 3281 | 4304 | 9773 | 91666 | 7.15 | 76.2 |
| Warsaw grouper | Epinephelus nigritus | 132 | 180 | 9773 | 91666 | 6.88 | 73.3 |
| Almaco jack | Seriola rivoliana | 530 | 750 | 9773 | 91666 | 6.63 | 70.7 |
| Gray triggerfish | Balistes capriscus | 4310 | 6102 | 9773 | 91666 | 6.62 | 70.6 |
| Scamp | Mycteroperca phenax | 732 | 1060 | 9773 | 91666 | 6.48 | 69.1 |
| Snowy grouper | Epinephelus niveatus | 73 | 110 | 9773 | 91666 | 6.22 | 66.4 |
| Lesser amberjack | Seriola fasciata | 91 | 145 | 9773 | 91666 | 5.89 | 62.8 |
| Queen triggerfish | Balistes vetula | 115 | 184 | 9773 | 91666 | 5.86 | 62.5 |
| Greater amberjack | Seriola dumerili | 2150 | 3924 | 9773 | 91666 | 5.14 | 54.8 |
| Bank sea bass | Centropristis ocyurus | 382 | 704 | 9773 | 91666 | 5.09 | 54.3 |
| Amberjack genus | Seriola spp. | 295 | 629 | 9773 | 91666 | 4.40 | 46.9 |
| Bigeye | Priacanthus arenatus | 23 | 50 | 9773 | 91666 | 4.31 | 46.0 |
| Sea bass genus | Centropristis spp. | 62 | 137 | 9773 | 91666 | 4.24 | 45.3 |
| Tomtate | Haemulon aurolineatum | 358 | 813 | 9773 | 91666 | 4.13 | 44.0 |
| Moray family | Muraenidae | 23 | 56 | 9773 | 91666 | 3.85 | 41.1 |
| Speckled hind | Epinephelus drummondhayi | 39 | 96 | 9773 | 91666 | 3.81 | 40.6 |
| Black snapper | Apsilus dentatus | 22 | 58 | 9773 | 91666 | 3.56 | 37.9 |
| Sharksucker | Echeneis naucrates | 48 | 130 | 9773 | 91666 | 3.46 | 36.9 |
| Atlantic spadefish | Chaetodipterus faber | 174 | 496 | 9773 | 91666 | 3.29 | 35.1 |
| Remora | Remora remora | 120 | 351 | 9773 | 91666 | 3.21 | 34.2 |
| Squirrelfish | Holocentrus adscensionis | 72 | 222 | 9773 | 91666 | 3.04 | 32.4 |


| Lane snapper | Lutjanus <br> synagris | 824 | 2543 | 9773 | 91666 | 3.04 | 32.4 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 8. Results of calculations used to identify species associated with red snapper in the western GOM (LA,TX). Species were assumed to be associated with red snapper if the association statistic was $\geq 3.0 . \% \mathrm{CO}$ is the percent common occurrence.

| Common Name | Scientific Name | Trips with Red Snapper and Species X | Trips with Species X |  | Total Trips | Association Statistic | \%CO |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| red snapper | Lutjanus campechanus | 1239 | 1239 | 1239 | 9265 | 7.48 | 100.0 |
| lane snapper | Lutjanus synagris | 193 | 205 | 1239 | 9265 | 7.04 | 94.1 |
| gag | Mycteroperca microlepis | 102 | 123 | 1239 | 9265 | 6.20 | 82.9 |
| vermilion snapper | Rhomboplites aurorubens | 83 | 104 | 1239 | 9265 | 5.97 | 79.8 |
| almaco jack | Seriola rivoliana | 47 | 60 | 1239 | 9265 | 5.86 | 78.3 |
| $\begin{array}{\|l\|} \hline \text { gray } \\ \text { triggerfish } \end{array}$ | Balistes capriscus | 320 | 409 | 1239 | 9265 | 5.85 | 78.2 |
| greater amberjack | Seriola dumerili | 265 | 363 | 1239 | 9265 | 5.46 | 73.0 |
| atlantic sharpnose shark | Rhizoprionodon terraenovae | 68 | 95 | 1239 | 9265 | 5.35 | 71.6 |
| cobia | Rachycentron canadum | 274 | 391 | 1239 | 9265 | 5.24 | 70.1 |
| great barracuda | Sphyraena barracuda | 38 | 56 | 1239 | 9265 | 5.07 | 67.9 |
| gray snapper | Lutjanus griseus | 188 | 281 | 1239 | 9265 | 5.00 | 66.9 |
| king mackerel | Scomberomorus cavalla | 164 | 310 | 1239 | 9265 | 3.96 | 52.9 |
| silver seatrout | Cynoscion nothus | 71 | 135 | 1239 | 9265 | 3.93 | 52.6 |
| blue runner | Caranx crysos | 122 | 236 | 1239 | 9265 | 3.87 | 51.7 |
| pinfish | Lagodon rhomboides | 99 | 210 | 1239 | 9265 | 3.53 | 47.1 |
| bluefish | Pomatomus saltatrix | 214 | 465 | 1239 | 9265 | 3.44 | 46.0 |
| requiem shark family | Carcharhinidae | 23 | 51 | 1239 | 9265 | 3.37 | 45.1 |
| atlantic spadefish | Chaetodipterus faber | 63 | 146 | 1239 | 9265 | 3.23 | 43.2 |
| blacktip shark | Carcharhinus limbatus | 107 | 266 | 1239 | 9265 | 3.01 | 40.2 |
| little tunny | Euthynnus alletteratus | 71 | 177 | 1239 | 9265 | 3.00 | 40.1 |

Table 9. A summary of formulation of the binomial model for INDEX 2. Factors were added to the model if PROBCHISQ $\leq 0.05$ and the reduction in DEV/DF (\%RED) $\geq 1.0 \%$ (bold blue font).

| There are no explanatory fa FACTOR | $\begin{gathered} \text { actors il } \\ \text { DEGF } \end{gathered}$ | n the base DEVIANCE | $\begin{gathered} \text { ode1 (NU } \\ \text { DEV/DF } \end{gathered}$ | ULL MODEL). \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BASE | 19710 | 27083.4 | 1.3741 |  | -13541.7 |  |  |
| STATE | 19707 | 23941.4 | 1.2149 | 11.59 | -11970.7 | 3142.06 | <0.0001 |
| MODE | 19708 | 25104.0 | 1.2738 | 7.30 | -12552.0 | 1979.42 | <0.0001 |
| YEAR | 19688 | 26121.8 | 1.3268 | 3.44 | -13060.9 | 916.60 | <0.0001 |
| SEASON | 19707 | 26767.0 | 1.3582 | 1.15 | -13383.5 | 316.44 | <0.0001 |
| The explanatory factors in FACTOR | the bas DEGF | e model ar DEVIANCE | STATE DEV/DF | \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 19707 | 23941.4 | 1.2149 |  | -11970.7 |  |  |
| MODE | 19705 | 21417.2 | 1.0869 | 10.53 | -10708.6 | 2524.14 | <0.0001 |
| YEAR | 19685 | 22700.7 | 1.1532 | 5.08 | -11350.3 | 1240.70 | <0.0001 |
| SEASON | 19704 | 23573.3 | 1.1964 | 1.52 | -11786.6 | 368.07 | <0.0001 |
| The explanatory factors in FACTOR | the bas DEGF | e model ar DEVIANCE | STATE | $\begin{aligned} & \text { MODE } \\ & \text { \%REDUCTION } \end{aligned}$ | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 19705 | 21417.2 | 1.0869 |  | -10708.6 |  |  |
| YEAR | 19683 | 20797.0 | 1.0566 | 2.79 | -10398.5 | 620.27 | <0.0001 |
| SEASON | 19702 | 21052.2 | 1.0685 | 1.69 | -10526.1 | 365.01 | <0.0001 |
| The explanatory factors in FACTOR | the bas DEGF | model ar DEVIANCE | STATE <br> DEV/DF | MODE YEAR \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 19683 | 20797.0 | 1.0566 |  | -10398.5 |  |  |
| SEASON | 19680 | 20406.1 | 1.0369 | 1.86 | -10203.1 | 390.83 | <0.0001 |
| The explanatory factors in | the bas | e mode1 ar | STATE | MODE YEAR SEASON |  |  |  |
| FACTOR | DEGF | DEVIANCE | DEV/DF | \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 19680 | 20406.1 | 1.0369 |  | -10203.1 |  |  |
| SEASON*STATE | 19671 | 19966.8 | 1.0150 | 2.11 | -9983.4 | 439.30 | $<0.0001$ |
| SEASON*MODE | 19674 | 20108.0 | 1.0221 | 1.43 | -10054.0 | 298.12 | <0.0001 |
| YEAR*MODE | 19654 | 20211.8 | 1.0284 | 0.82 | -10105.9 | 194.28 | <0.0001 |
| The explanatory factors in | the bas | e mode1 ar | STATE | MODE YEAR SEASON | SEASON*S |  |  |
| FACTOR | DEGF | DEVIANCE | DEV/DF | \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| BASE | 19671 | 19966.8 | 1.0150 |  | -9983.4 |  |  |
| SEASON*MODE | 19665 | 19793.5 | 1.0065 | 0.84 | -9896.7 | 173.34 | <0.0001 |
| YEAR*MODE | 19645 | 19798.9 | 1.0078 | 0.71 | -9899.4 | 167.98 | <0.0001 |

Table 10. A summary of formulation of the Poisson model for INDEX 2. Factors were added to the model if PROBCHISQ $\leq 0.05$ and the reduction in DEV/DF (\%RED) $\geq 1.0 \%$ (bold blue font).

| There are no explanatory f FACTOR | ors DEGF | n the base DEVIANCE | de1. DEV/DF | \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BASE | 10945 | 222932.6 | 20.3684 |  | 712770.5 |  |  |
| YEAR | 10923 | 207656.8 | 19.0110 | 6.66 | 720408.4 | 15275.80 | <0.0001 |
| STATE | 10942 | 212373.7 | 19.4090 | 4.71 | 718050.0 | 10558.97 | <0.0001 |
| SEASON | 10942 | 219353.1 | 20.0469 | 1.58 | 714560.3 | 3579.52 | <0.0001 |
| MODE | 10943 | 220713.2 | 20.1694 | 0.98 | 713880.2 | 2219.43 | <0.0001 |
| The explanatory factors in FACTOR | the base model are DEGF DEVIANCE |  | YEAR <br> DEV/DF | \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
|  |  |  |  |  |  |  |  |
| BASE | 10923 | 207656.8 | 19.0110 |  | 720408.4 |  |  |
| STATE | 10920 | 197414.2 | 18.0782 | 4.91 | 725529.8 | 10242.62 | <0.0001 |
| SEASON | 10920 | 205433.9 | 18.8126 | 1.04 | 721519.9 | 2222.88 | <0.0001 |
| MODE | 10921 | 206629.3 | 18.9204 | 0.48 | 720922.2 | 1027.48 | <0.0001 |
| The explanatory factors in FACTOR | the base model are DEGF DEVIANCE |  | YEAR STATE |  | LOGLIKE | CHISQ | PROBCHISQ |
|  |  |  | DEV/DF | \%REDUCTION |  |  |  |
| BASE | 10920 | 197414.2 | 18.0782 |  | 725529.8 |  |  |
| MODE | 10918 | 194925.3 | 17.8536 | 1.24 | 726774.2 | 2488.89 | <0.0001 |
| SEASON | 10917 | 194921.6 | 17.8549 | 1.24 | 726776.1 | 2492.64 | <0.0001 |
| The explanatory factors in FACTOR | the bas DEGF | mode1 ar DEVIANCE | YEAR STATE MODE <br> DEV/DF \%REDUCTION |  | LOGLIKE | CHISQ | PROBCHISQ |
|  |  |  |  |  |  |  |  |  |
| BASE | 10918 | 194925.3 | 17.8536 |  | 726774.2 |  |  |
| SEASON | 10915 | 192605.3 | 17.6459 | 1.16 | 727934.2 | 2319.99 | <0.0001 |
| The explanatory factors in FACTOR | the bas DEGF | mode1 are DEVIANCE | $\begin{aligned} & \text { YEAR S } \\ & \text { DEV/DF } \end{aligned}$ | ate mode sea \%REDUCTION | LOGLIKE | CHISQ | PROBCHISQ |
|  |  |  |  |  |  |  |  |
| BASE | 10915 | 192605.3 | 17.6459 |  | 727934.2 |  |  |
| YEAR*STATE | 10850 | 182080.6 | 16.7816 | 4.90 | 733196.6 | 10524.70 | <0.0001 |
| YEAR*SEASON | 10849 | 185757.5 | 17.1221 | 2.97 | 731358.1 | 6847.78 | <0.0001 |
| YEAR*MODE | 10889 | 188642.9 | 17.3242 | 1.82 | 729915.4 | 3962.44 | <0.0001 |
| MODE*STATE | 10910 | 189707.4 | 17.3884 | 1.46 | 729383.1 | 2897.90 | <0.0001 |
| SEASON*STATE | 10906 | 191732.6 | 17.5805 | 0.37 | 728370.6 | 872.75 | <0.0001 |
| SEASON*MODE | 10909 | 191915.3 | 17.5924 | 0.30 | 728279.2 | 690.01 | <0.0001 |
| The explanatory factors in FACTOR | the base model are DEGF DEVIANCE |  | YEAR DEV/DF | TATE MODE SEASON | YEAR*STATE <br> LOGLIKE | CHISQ | PROBCHISQ |
|  |  |  | \%REDUCTION |  |  |  |  |
| BASE | 10850 | 182080.6 |  | 16.7816 |  | 733196.6 | 6099.37 | <0.0001 |
| YEAR*SEASON |  | 175981.2 | 16.3187 | 2.76 | 736246.2 |  |  |  |
| YEAR*MODE | 10824 | 179343.3 | 16.5690 | 1.27 | 734565.2 | 2737.30 | <0.0001 |  |
| MODE*STATE | 10845 | 180622.2 | 16.6549 | 0.76 | 733925.7 | 1458.35 | <0.0001 |  |
| SEASON*STATE | 10841 | 181191.6 | 16.7135 | 0.41 | 733641.1 | 889.03 | <0.0001 |  |
| SEASON*MODE | 10844 | 181523.9 | 16.7396 | 0.25 | 733474.9 | 556.71 | <0.0001 |  |
| The explanatory factors in | the base model are |  | $\begin{aligned} & \text { YEAR S } \\ & \text { DEV/DF } \end{aligned}$ | TATE MODE SEASON \%REDUCTION | YEAR*STATE <br> LOGLIKE | YEAR*SEASONCHISQ | PROBCHISQ |  |
| FACTOR |  |  |  |  |  |  |  |  |  |
| BASE | 10784 | 175981.2 | 16.3187 |  | 736246.2737362.1 | 223174 | <0.0001 |  |
| YEAR*MODE | 10758 | 173749.5 | 16.1507 | 1.03 |  |  |  |  |
| MODE*STATE | 10779 | 174521.1 | 16.1908 | 0.78 | 736976.3 | 1460.08 | <0.0001 |  |
| SEASON*STATE | 10775 | 175163.3 | 16.2565 | 0.38 | 736655.2 | 817.88 | <0.0001 |  |
| SEASON*MODE | 10778 | 175801.3 | 16.3111 | 0.05 | 736336.2 | 179.92 | <0.0001 |  |
| The explanatory factors in FACTOR | the bas DEGF | mode1 ar DEVIANCE | $\begin{aligned} & \text { YEAR § } \\ & \text { DEV/DF } \end{aligned}$ | TATE MODE SEASON \%REDUCTION | YEAR*STATE <br> LOGLIKE | $\begin{aligned} & \text { YEAR*SEASON } \\ & \text { CHISQ } \end{aligned}$ | YEAR*MODE PROBCHISQ |  |
| BASE | 10758 | 173749.5 | 16.1507 |  | 737362.1 |  |  |  |
| MODE*STATE | 10753 | 172633.0 | 16.0544 | 0.60 | 737920.4 | 1116.48 | <0. 0001 |  |
| SEASON*STATE | 10749 | 172998.1 | 16.0943 | 0.35 | 737737.8 | 751.41 | <0.0001 |  |
| SEASON*MODE | 10752 | 173514.9 | 16.1379 | 0.08 | 737479.4 | 234.61 | <0.0001 |  |

Table 11. Relative nominal CPUE, number of trips, number of positive trips, proportion positive trips (PPT) and abundance index statistics (INDEX 2).

| YEAR | TRIPS | POS <br> TRIPS | PPT | Rel. <br> Nominal <br> CPUE | Rel. <br> Index | Lower <br> $95 \%$ Cl | Upper <br> $\mathbf{9 5 \%} \mathbf{C l}$ | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 263 | 99 | 0.376 | 0.761 | 0.835 | 0.331 | 1.340 | 0.308 |
| 1982 | 408 | 177 | 0.434 | 0.416 | 0.440 | 0.180 | 0.700 | 0.301 |
| 1983 | 700 | 417 | 0.596 | 1.200 | 1.046 | 0.474 | 1.617 | 0.279 |
| 1984 | 553 | 198 | 0.358 | 0.547 | 0.464 | 0.175 | 0.753 | 0.318 |
| 1985 | 465 | 165 | 0.355 | 0.714 | 0.646 | 0.245 | 1.046 | 0.317 |
| 1986 | 710 | 362 | 0.510 | 0.650 | 0.714 | 0.306 | 1.122 | 0.291 |
| 1987 | 664 | 251 | 0.378 | 0.499 | 0.645 | 0.270 | 1.019 | 0.296 |
| 1988 | 566 | 198 | 0.350 | 0.431 | 0.649 | 0.265 | 1.033 | 0.302 |
| 1989 | 501 | 159 | 0.317 | 0.393 | 0.450 | 0.167 | 0.734 | 0.321 |
| 1990 | 350 | 153 | 0.437 | 0.636 | 0.624 | 0.247 | 1.001 | 0.308 |
| 1991 | 504 | 265 | 0.526 | 1.369 | 1.104 | 0.494 | 1.714 | 0.282 |
| 1992 | 1036 | 543 | 0.524 | 1.364 | 1.417 | 0.685 | 2.148 | 0.263 |
| 1993 | 656 | 325 | 0.495 | 1.076 | 1.219 | 0.579 | 1.859 | 0.268 |
| 1994 | 704 | 339 | 0.482 | 1.016 | 0.940 | 0.437 | 1.442 | 0.273 |
| 1995 | 451 | 210 | 0.466 | 0.812 | 0.883 | 0.380 | 1.386 | 0.290 |
| 1996 | 558 | 271 | 0.486 | 1.254 | 1.060 | 0.488 | 1.631 | 0.275 |
| 1997 | 832 | 522 | 0.627 | 1.774 | 1.658 | 0.802 | 2.513 | 0.263 |
| 1998 | 1114 | 676 | 0.607 | 1.378 | 1.410 | 0.682 | 2.138 | 0.263 |
| 1999 | 1709 | 1093 | 0.640 | 1.535 | 1.551 | 0.764 | 2.339 | 0.259 |
| 2000 | 1746 | 1211 | 0.694 | 1.386 | 1.354 | 0.657 | 2.050 | 0.263 |
| 2001 | 1615 | 992 | 0.614 | 1.155 | 1.228 | 0.588 | 1.868 | 0.266 |
| 2002 | 1782 | 1158 | 0.650 | 1.426 | 1.381 | 0.680 | 2.082 | 0.259 |
| 2003 | 1824 | 1162 | 0.637 | 1.209 | 1.282 | 0.625 | 1.940 | 0.262 |



Figure 1. Proportion positive trips 1981-2003 (Index 1).


Figure 2. Chi-square residuals for the binomial model (Index 1) by Mode (A), STATE (B), MONTH (C) and Year (D).


Figure 3. Nominal CPUE (Index 1) 1981-2003.


Figure 4. Residuals for the Poisson model (Index 1) by Mode (A), STATE (B), MONTH (C) and Year (D).


Figure 5. Index 1: Nominal CPUE, scaled to the mean (gray triangles), and the standardized index, also scaled to the mean (black circles) with upper and lower $95 \%$ confidence intervals (dashed gray lines).


Figure 6. Proportion positive trips 1981-2003 (Index 2).


Figure 7. Chi-square residuals for the binomial model (Index 2) by Mode (A), STATE (B), SEASON (C) and YEAR (D).


Figure 8. Nominal CPUE (Index 2) 1981-2003.


Figure 9. Residuals for the Poisson model (Index 2) by Mode (A), STATE (B), SEASON (C) and Year (D).


Figure 10. Index 2: Nominal CPUE, scaled to the mean (gray triangles), and the standardized index, also scaled to the mean (black circles) with upper and lower $95 \%$ confidence intervals (dashed gray lines).

Index Comparison


Figure 11. Comparison of all indices, including Schirripa and Legault ,1999. All indices are scaled to their common mean.


[^0]:    ${ }^{1}$ NOAA Fisheries, Southeast Fisheries Science Center, Panama City Laboratory.

[^1]:    ${ }^{2}$ Patty Phares. Personal communication. NOAA Fisheries, Southeast Fisheries Science Center. Miami Laboratory.

[^2]:    ${ }^{3}$ Heinemann, Dennis. The Ocean Conservancy, 1725 DeSales Street, Suite 600, Washington, D.C. 20036

[^3]:    ${ }^{4}$ Type- 3 model, error $=$ binomial, link $=$ logit, response variable $=$ success $($ where success $=1$ if red snapper catch $>$ 0 , else success $=0$ )
    ${ }^{5}$ Type- 3 model, error $=$ Poisson, link $=\log$, offset $=\log ($ angler hours $)$, response variable $=$ catch $($ where catch $\neq 0)$.

