

## SEDAR

## Southeast Data, Assessment, and Review

## SEDAR 60

# South Atlantic Red Porgy Stock Assessment Report 

April 2020
SEDAR
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## Table of Contents

Each Section is Numbered Separately

## Section I Introduction <br> Pg. 4

Section II Assessment Report
Pg. 36


## SEDAR

## Southeast Data, Assessment, and Review

## SEDAR 60

# South Atlantic Red Porgy Section I: Introduction 

## April 2020

SEDAR
4055 Faber Place Drive, Suite 201 North Charleston, SC 29405I. Introduction2

1. SEDAR Process Description. ..... 2
2. Management Overview: ..... 3
2.1 SAFMC Fishery Management Plan and Amendments ..... 3
2.1.1 Emergency and Interim Rules ..... 8
2.1.2 Secretarial Amendments ..... 8
2.1.3 Control Date Notices ..... 8
2.1.4 Management Program Specifications. ..... 9
2.1.5 Stock Rebuilding Information ..... 11
2.2 SAFMC Management and Regulatory Timeline ..... 13
2.2.1 Closures Due to Meeting Commercial Quota or Commercial/Recreational ACL ..... 16
2.3 . State Regulatory History ..... 16
2.3.1 North Carolina: ..... 16
2.3.2 South Carolina: ..... 23
2.3.3 Georgia: ..... 23
2.3.4 Florida: ..... 23
3. Assessment History ..... 28
4. Regional Maps ..... 29
5. Abbreviations ..... 30

## I. Introduction

## 1. SEDAR Process Description

SouthEast Data, Assessment, and Review (SEDAR) is a cooperative Fishery Management Council process initiated in 2002 to improve the quality and reliability of fishery stock assessments in the South Atlantic, Gulf of Mexico, and US Caribbean. The improved stock assessments from the SEDAR process provide higher quality information to address fishery management issues. SEDAR emphasizes constituent and stakeholder participation in assessment development, transparency in the assessment process, and a rigorous and independent scientific review of completed stock assessments.

SEDAR is managed by the Caribbean, Gulf of Mexico, and South Atlantic Regional Fishery Management Councils in coordination with NOAA Fisheries and the Atlantic and Gulf States Marine Fisheries Commissions. Oversight is provided by a Steering Committee composed of NOAA Fisheries representatives: Southeast Fisheries Science Center Director and the Southeast Regional Administrator; Regional Council representatives: Executive Directors and Chairs of the South Atlantic, Gulf of Mexico, and Caribbean Fishery Management Councils; a representative from the Highly Migratory Species Division of NOAA Fisheries; and Interstate Commission representatives: Executive Directors of the Atlantic States and Gulf States Marine Fisheries Commissions.

SEDAR 60 addressed the stock assessment for South Atlantic Red Porgy. The assessment process consisted of a series of webinars held from June 2018 - February 2020 and an in person workshop held in Beaufort, North Carolina on December 10-12, 2019. The Stock Assessment Report is organized into 2 sections. Section I -Introduction contains a brief description of the SEDAR Process, Assessment and Management Histories for the species of interest, and the management specifications requested by the Cooperator. Section II is the Assessment Process report. This section details the assessment model, as well as documents any data recommendations that arise for new data sets presented during this assessment process, or changes to data sets used previously.

The final Stock Assessment Reports (SAR) for South Atlantic Red Porgy was disseminated to the public in April 2020. The Council's Scientific and Statistical Committee (SSC) will review the SAR for its stock. The SSCs are tasked with recommending whether the assessments represent Best Available Science, whether the results presented in the SARs are useful for providing management advice and developing fishing level recommendations for the Council. An SSC may request additional analyses be conducted or may use the information provided in the SAR as the basis for their Fishing Level Recommendations (e.g., Overfishing Limit and Acceptable Biological Catch). The South Atlantic Fishery Management Council's SSC will review the assessment at its April 2020 meeting, followed by the Council receiving that information at its June 2020 meeting. Documentation on SSC recommendations is not part of the SEDAR process and is handled through each Council.

## 2. Management Overview:

### 2.1 SAFMC Fishery Management Plan and Amendments

The following summary describes only those management actions that likely affect Red Porgy fisheries and harvest.

## Original SAMFC FMP

The Fishery Management Plan (FMP), Regulatory Impact Review, and Final Environmental Impact Statement for the Snapper Grouper Fishery of the South Atlantic Region, approved in 1983 and implemented in August of 1983, establishes a management regime for the fishery for snappers, groupers and related demersal species of the Continental Shelf of the southeastern United States in the exclusive economic zone (EEZ) under the area of authority of the South Atlantic Fishery Management Council (Council) and the territorial seas of the states, extending from the North Carolina/Virginia border through the Atlantic side of the Florida Keys to $83^{\circ} \mathrm{W}$ longitude. Regulations apply only to federal waters.

## SAFMC FMP Amendments affecting Red Porgy

| Description of Action | FMP/Amendment | Effective <br> Date |
| :---: | :---: | :---: |
| -4" Trawl mesh size <br> -Gear limitations (poisons, explosives, fish traps, trawls) <br> -Designated modified habitats or artificial reefs as Special Management Zones | Snapper Grouper FMP | 8/31/1983 |
| -Prohibit trawls to harvest snapper grouper species south of Cape Hatteras, NC and north of Cape Canaveral, FL <br> -Defined directed fishery as vessel with trawl gear and at least 200 pounds of snapper grouper species on board | Amendment 1 | 1/12/1989 |
| -Prohibited gear: fish traps except black sea bass pots north of Cape Canaveral, FL; entanglement nets; longlines inside 50 fathoms; powerheads in designated SMZs off SC <br> -Defined overfishing/overfished established rebuilding timeframe: red porgy <br> $\leq 10$ years (year $1=1991$ ) <br> -Required permits (commercial and for-hire) and specified data collection regulations -No retention of snapper grouper species caught in other fisheries with gear prohibited in snapper grouper fishery if captured snapper grouper had no bag limit or harvest was prohibited. If had a bag limit, could retain only the bag limit; $-12 " \text { TL limit - red porgy }$ | Amendment 4 | 1/1/1992 |


| -Required 100\% logbook coverage upon renewal <br> of commercial permit <br> -Oculina Experimental Closed Area | Amendment 6 | $7 / 27 / 1994$ |
| :--- | :--- | :--- |
| -Required dealer, charter and headboat federal <br> permits <br> -Specified allowable gear and made allowance for <br> experimental gear <br> -Restricted sale/purchase of snapper grouper species <br> -Adjusted requirements for possessing multi-day bag <br> limits | Amendment 7 | $1 / 23 / 1995$ |
| -Established limited entry for commercial <br> snapper grouper fishery |  |  |
| -Increased red porgy minimum size limit to 14" TL <br> (commercial and recreational) | Amendment 8 | $12 / 14 / 1998$ |
| -Specified red porgy bag limit of 5 fish per person <br> per day. <br> -No harvest and possession of red porgy above the <br> bag limit and no purchase or sale in March and <br> April. | Amendment 9 | $2 / 24 / 1999$ |
| -Approved definitions for overfished and <br> overfishing. <br> -MSST = [(1-M) or 0.5, whichever is <br> greater]*B |  |  |
| -MFMT |  |  |
| -Overfished/overfishing evaluations: red porgy <br> overfished (static SPR = 14-19\%) | Amendment 11 | $12 / 2 / 1999$ |
| For Red porgy: <br> -MSY=4.38 mp; OY=45\% static SPR; <br> MFMT=0.43; MSST=7.34 mp; rebuilding <br> timeframe=18 years (1999=year 1); <br> -No sale of red porgy during Jan-April; <br> -1 fish bag limit; <br> -50 lb. bycatch commercial trip limit May- <br> December | Amendment 13A | $4 / 26 / 2004$ |
| -Extended for an indefinite period the regulation <br> prohibiting fishing for and possessing snapper <br> grouper species within the Oculina Experimental <br> Closed Area. | Amendment 12 | $9 / 22 / 2000$ |


| Red Porgy: <br> Commercial and recreational: <br> -Retained 14" TL size limit and seasonal closure (retention limited to the bag limit); <br> -Specified a commercial quota of $127,000 \mathrm{lbs}$ gw and prohibit sale/purchase and prohibit harvest and/or possession beyond the bag limit when quota is taken and/or during January through April; -Increased commercial trip limit from 50 lbs ww to 120 fish ( 210 lbs gw ) during May through December -Increased recreational bag limit from one to three red porgy per person per day or per trip, whichever is more restrictive. | Amendment 13C | 10/23/2006 |
| :---: | :---: | :---: |
| -Established eight deepwater Type II marine protected areas (MPAs) to protect a portion of the population and habitat of long-lived deepwater snapper grouper species. | Amendment 14 | 2/12/2009 |
| -Updated management reference points and defined rebuilding strategy for red porgy. | Amendment 15A | 3/14/2008 |
| -Established recreational and commercial shallow water grouper spawning closure January through April to address overfishing of gag <br> -Established recreational closed season for vermilion snapper from November through March. <br> -Required venting and dehooking tools when catching snapper grouper species to reduce recreational and commercial bycatch mortality. | Amendment 16 | 2/29/2009 |
| -Prohibited the sale of snapper grouper species harvested or possessed in the EEZ under the bag limits and prohibited the sale of snapper grouper species harvested or possessed under the bag limits by vessels with a Federal charter vessel/headboat permit for South Atlantic snapper grouper regardless of where harvested. <br> -Established allocations for red porgy ( $50 \%$ commercial \& 50\% recreational). Commercial quota $=190,050 \mathrm{lbs}$ gutted weight $(197,652 \mathrm{lbs}$ whole weight). Recreational quota $=190,050 \mathrm{lbs}$ gutted weight. | Amendment 15B | 2/15/2010 |
| -Required use of non-stainless-steel circle hooks when fishing for snapper grouper species with hook-and-line gear north of 28 deg. N latitude in the South Atlantic EEZ <br> -Implemented an area closure for snapper-grouper species. | Amendment 17A | 3/3/2011 |


| -Limit harvest of snapper grouper species in SC <br> SMZs to the bag limit; | Amendment 23 <br> (Comprehensive <br> Ecosystem-based <br> Amendment 2) | $1 / 30 / 2012$ |
| :--- | :--- | :--- |
| -Reorganized FMU into 6 complexes (deepwater, <br> jacks, snappers, grunts, shallow-water groupers, <br> porgies) (see final rule for species list); <br> -Established acceptable biological catch (ABC) <br> control rules and established ABCs, ACLs, and AMs <br> for species not undergoing overfishing; <br> -Established commercial quota as commercial ACL <br> for red porgy <br> and specified recreational ACL (197,652 lbs ww). | Amendment 25 <br> (Comprehensive ACL <br> Amendment) | $4 / 16 / 2012$ |
| -Modified the restriction on retention of bag limit <br> quantities of some snapper grouper species by captain <br> and crew of for-hire vessels; | Amendment 27 | $1 / 27 / 2014$ |
| -Required headboat vessels to report electronically at <br> weekly intervals. | Amendment 31 (Joint <br> South Atlantic and Gulf <br> of Mexico Generic <br> Headboat Reporting <br> Amendment) | $1 / 27 / 2014$ |
| -Modified accountability measures for snapper <br> grouper species, including red porgy | Amendment 34 (Generic <br> Accountability Measures <br> and Dolphin Allocation <br> Amendment) | $2 / 22 / 2016$ |
| Amendment 36 | $7 / 31 / 2017$ |  |

SAFMC Regulatory Amendments affecting Red Porgy

| Description of Action | Amendment | Effective Date |
| :---: | :---: | :---: |
| -Prohibited fishing in SMZs except with hand-held hook-and-line and spearfishing gear | Regulatory Amendment 1 | 3/27/1987 |
| -Allowed multi-gear trips for black sea bass and allowed retention of incidentally-caught snapper grouper species on black sea bass trips <br> -As FYI - from 1990 through 2017, red porgy were incidentally caught in $46 \%$ of hook-and-line black sea bass trips and $13 \%$ of trips with pot gear. | Regulatory Amendment 4 | 7/6/1993 |
| -Eliminated closed area for snapper grouper species approved in Amendment 17A. | Regulatory Amendment 10 | 5/31/2011 |
| $\begin{gathered} -\mathrm{MSY}=834,000 \mathrm{lbs} \text { whole weight } \\ -\mathrm{OY}=\mathrm{ACL}=\mathrm{ABC} \\ 2013=306,000 \mathrm{lbs} \text { ww } \\ 2014=309,000 \mathrm{lbs} \text { ww } \end{gathered}$ | Regulatory Amendment 18 | 9/5/2013 |
| 2015 and subsequent years=328,000 lbs ww; -Revised commercial/recreational ACL (as FYI - gutted weight determined with conversion factor of 1.04 from commercial logbooks): $\begin{aligned} & 2013=147,115 \mathrm{lbs} \text { gw }(153,000 \mathrm{lbs} \text { ww }) \\ & 2014=148,558 \mathrm{lbs} \text { gw }(154,500 \mathrm{lbs} \text { ww }) \end{aligned}$ <br> 2015 and subsequent years $=157,692 \mathrm{lbs} \mathrm{gw}(164,000 \mathrm{lbs}$ ww) <br> -Removed vermilion snapper November through March recreational closure |  |  |
| -Modified the gag commercial AM to remove the requirement that all other shallow water groupers (black grouper, red grouper, scamp, red hind, rock hind, graysby, coney, yellowmouth grouper, and yellowfin grouper) are prohibited from harvest in the South Atlantic when the gag commercial ACL is met or projected to be met. | Regulatory Amendment 15 | 9/12/2013 |
| -Implemented an annual closure on the use of black sea bass pots from November 1 to April 30. | Regulatory Amendment 19 | 10/23/13 |
| -Modified the definition of the overfished threshold (MSST) for several snapper grouper species, including red porgy. MSST $=75 \%$ SSB $_{\mathrm{MSY}}$ | Regulatory Amendment 21 | 11/6/2014 |
| -Revise the area where fishing with black sea bass pots is prohibited from Nov.1-April 30. | Regulatory Amendment 16 | 12/29/2016 |

### 2.1.1 Emergency and Interim Rules

- For Black Seabass - modified definition of bsb pot; allowed multi-gear trips for bsb; allowed retention of incidentally-caught fish on bsb trips Initial emergency rule 8/31/1992; emergency rule extension 11/30/1992
- Prohibited harvest or possession of red porgy effective 9/8/1999 - Rule expired on 8/28/2000


### 2.1.2 Secretarial Amendments

None

### 2.1.3 Control Date Notices

Notice of Control Date ( $\mathbf{0 7 / 3 0 / 9 1} 56$ FR 36052) - Anyone entering federal snapper grouper fishery (other than for wreckfish) in the EEZ off S. Atlantic states after 07/30/91 was not assured of future access if limited entry program developed.

Notice of Control Date (10/14/05 70 FR 60058) - Anyone entering federal snapper grouper fishery off
S. Atlantic states after 10/14/05 was not assured of future access if limited entry program developed.

Notice of Control Date (3/8/07 72 FR 60794) - Considered measures to limit participation in the snapper grouper for-hire sector effective 3/8/07.

Notice of Control Date ( $\mathbf{0 1 / 3 1 / 1 1} \mathbf{7 6}$ FR 5325) - Anyone entering federal snapper grouper fishery off S. Atlantic states after 09/17/10 was not assured of future access if limited entry program developed.

Notice of Control Date ( $\mathbf{0 6} / \mathbf{1 5} / \mathbf{2 0 1 6} 81$ FR 66244) - fishermen who enter the federal for-hire recreational sector for the Snapper Grouper fishery after June 15, 2016, will not be assured of future access should a management regime that limits participation in the sector be prepared and implemented.

### 2.1.4 Management Program Specifications

Table 2.1.4.1. General Management Information South Atlantic

| Species | Red Porgy (Pagrus pagrus) |
| :--- | :--- |
| Management Unit | Southeastern US |
| Management Unit Definition | All waters within South Atlantic Fishery <br> Management Council Boundaries |
| Management Entity | South Atlantic Fishery Management Council |
| Management Contacts <br> SERO / Council | SAFMC: Myra Brouwer |
| Current stock exploitation status | SERO: Rick DeVictor |

Table 2.1.4.2. Management Parameters

| Criteria |  | South Atlantic - Current (2012 SEDAR 1 Update) |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Definition | Base Run Values | Units | Median of Base <br> Run MCBs |
| MSST | $(1-\mathrm{M}) * \mathrm{SSB}_{\mathrm{MSY}}$ | 3,048 | mt of all mature fish |  |
| MFMT | $\mathrm{F}_{\mathrm{MSY}}$, if available; $\mathrm{F}_{\text {MSY }}$ proxy if not | 0.17 | per year |  |
| $\mathrm{F}_{\text {MSY }}$ | $\mathrm{F}_{\text {MSY }}$ | 0.17 | per year |  |
| MSY | Yield at $\mathrm{F}_{\mathrm{MSY}}$, landings | 834 | 1,000 lbs. |  |
| $\mathrm{B}_{\mathrm{MS}_{\mathrm{Y}}}{ }^{1}$ | Spawning stock biomass | 4,254 | mt of all mature fish |  |
| $\mathrm{R}_{\text {MSY }}$ | Recruits at MSY | 2,222 | 1,000 age-0 fish |  |
| F Target | 75\% F MSY | 0.13 | per year |  |
| Yield at $\mathrm{F}_{\text {TARGET }}$ (equilibrium) | Landings | 810 | 1,000 lbs. |  |
| M | Natural mortality, constant across ages | 0.225 | per year |  |
| $\mathrm{F}_{\text {Current }}$ (2009-2011) | Geometric mean of F in last 3 years | 0.11 | per year |  |
| Terminal Biomass $(2011)^{1}$ | Spawning Stock Biomass in terminal year | 2,018 | mt of all mature fish |  |
| Exploitation Status | $\mathrm{F}_{2009-2011 / \mathrm{F}_{\text {MSY }}}$ | 0.64 |  |  |
| Biomass Status ${ }^{1}$ | $\begin{aligned} & \hline \text { SSB }_{2011} / \mathrm{MSST}_{2} \\ & \text { SSB }_{2011} / \text { SSB }_{\mathrm{MSY}} \end{aligned}$ | $\begin{aligned} & 0.61 \\ & 0.47 \end{aligned}$ |  |  |
| Generation Time |  | 8 | years |  |
| TREBUILD (if appropriate) |  | 18* | years |  |

* Am 12 established 18 -year rebuilding schedule with 1991 = year 1

Table 2.1.4.2 Continued Management Parameters

| Criteria | South Atlantic - Proposed (values from SEDAR 60) |  |  |
| :--- | :--- | :--- | :--- |
|  | Definition | Base Run <br> Values | Median of Base Run <br> MCBs |
| MSST $^{1}$ | $75 \%$ SSB $_{\text {MSY }}$ |  |  |

1. Biomass values reported for management parameters and status determinations should be based on the biomass metric recommended through the Assessment process and SSC. This may be total, spawning stock or some measure thereof, and should be applied consistently in this table.

NOTE: "Proposed" columns are for indicating any definitions that may exist in FMPs or amendments that are currently under development and should therefore be evaluated in the current assessment. Please clarify whether landings parameters are 'landings' or 'catch' (Landings + Discard). If 'landings', please indicate how discards are addressed.

### 2.1.5 Stock Rebuilding Information

Amendment 12 (SAFMC 2000) established an 18-year rebuilding schedule for red porgy with 1991 being year 1 .

The most recent assessment update (SEDAR 1 update 2012) included data through 2011, adding an additional six years of landings information to the 2006 update. The South Atlantic Council's SSC reviewed the 2012 assessment update for red porgy in October 2012. The National Standard 1 Guidelines state that, for overfished stocks and stock complexes, a rebuilding ABC must be set to reflect the annual catch that is consistent with the schedule of fishing mortality rates in the rebuilding plan. None of the projection scenarios in the assessment update demonstrated that red porgy could be rebuilt by the end of the rebuilding schedule (2018) even in the absence of fishing mortality. Hence, the SSC recommended using a provision of the NMFS National Standard 1 (NS1) that states "if the stock or stock complex has not rebuilt by $\mathrm{T}_{\text {MAX }}$, then the fishing mortality rate should be maintained at $\mathrm{F}_{\text {rebuild }}$ or $75 \%$ of the maximum fishing mortality threshold (MFMT), whichever is less." Since F at 75\% of $\mathrm{F}_{\text {MSY }}$ estimated in the model is very close to the level associated with red porgy bycatch harvest, the SSC recommended using this value in setting the acceptable biological catch (ABC).

Table 2.1.5.1. General Projection Specifications
South Atlantic

| First Year of Management | Assume management begins in 2020. <br> However, if there are no changes to the <br> reference points, a projection with the <br> revised ABC and OFL should be provided <br> assuming that landings limits are changed <br> in <br> the 2019 fishing year. |  |  |
| :--- | :--- | :---: | :---: |
| Interim basis | SEDAR 60 ToR ask the Panel to <br> provide guidance on appropriate <br> assumptions to address harvest and <br> mortality levels in <br> interim years; recent SEDAR assessments |  |  |
|  | have asked for ACL, if ACL is met <br> Average exploitation, if ACL is not met |  |  |
|  |  |  |  |
| Projection Outputs | Pounds and numbers |  |  |
| Landings | Pounds and numbers |  |  |
| Discards | F \& Probability F>MFMT |  |  |
| Exploitation | B \& Probability B>MSST <br> (and Prob. B>B MSY if under rebuilding <br> plan) |  |  |
| Biomass (total or SSB, as <br> appropriate) | Number |  |  |
| Recruits |  |  |  |

Table 2.1.5.2 Base Run Projections Specifications. Long Term and Equilibrium conditions.

| Criteria | Definition | If overfished | If overfishing | Neither <br> overfished nor <br> overfishing |
| :--- | :--- | :---: | :---: | :---: |
| Projection <br> Values | Years | T $_{\text {REBUILD }}$ | X | 10 |

NOTE: Exploitation rates for projections may be based upon point estimates from the base run (current process) or upon the median of such values from the MCBs evaluation of uncertainty. The critical point is that the projections be based on the same criteria as the management specifications.

Table 2.1.5.3. P-star projections. Short term specifications for OFL and ABC recommendations. Additional P-star projections may be requested by the SSC once the ABC control rule is applied.

| Basis | Value | Years to Project | $\mathrm{P}^{*}$ applies to |
| :---: | :---: | :---: | :---: |
| $\mathrm{P}^{*}$ | $50 \%$ | Interim +5 | Probability of <br> overfishing |
| $\mathrm{P}^{*}$ | $35 \%$ | Interim +5 | Probability of <br> overfishing |
| Exploitation | $\mathrm{F}_{\mathrm{MSY}}$ | Interim +5 | NA |
| Exploitation | $75 \%$ of $\mathrm{F}_{\mathrm{MSY}}$ | Interim +5 | NA |

Table 2.1.5.4. Quota Calculation Details
If the stock is managed by quota, please provide the following information

| Current Acceptable Biological Catch <br> (ABC) and Total Annual Catch Level <br> (ACL) Value for Red Porgy | $328,000 \mathrm{lbs}$ ww |
| :--- | :---: |
| Commercial ACL for Red Porgy | $164,000 \mathrm{lbs}$ ww |
| Recreational ACL for Red Porgy | $164,000 \mathrm{lbs}$ ww |
| Next Scheduled Quota Change | N/A |
| Annual or averaged quota? | annual |
| If averaged, number of years to average | N/A |
| Does the quota include bycatch/discard? | No |

How is the quota calculated - conditioned upon exploitation or average landings? The ACL is set equal to the ABC , which comes directly from the assessment projections. The sector allocations were set by the Council at $50 \%$ commercial and $50 \%$ recreational. These allocations were chosen because they were closest to the status quo at the time allocations were being discussed for Red Porgy (average landings 1999-2003 were 49\% commercial:51\% recreational).

Does the quota include bycatch/discard estimates? If so, what is the source of the bycatch/discard values? What are the bycatch/discard allowances?
The quota does not explicitly include estimates of discards in it. However, the projections assume a certain number of dead discards will occur when the quota is met and that the total F associated with both the landings and discards will not result in overfishing.

## Are there additional details of which the analysts should be aware to properly determine quotas for this stock?

### 2.2 SAFMC Management and Regulatory Timeline

The following tables provide a timeline of federal management actions by fishery.

2.2.2 South Atlantic Red Porgy Federal Recreational Regulatory History

| Year | Quota (lbs) | ACL (lbs) | $\begin{aligned} & \text { Days } \\ & \text { Dopen } \end{aligned}$ | fishing season | reason for closure | season start date (first day implemented) | season end date (last day effective) | Size limit (in TL) | size limit start date | size limit end date | Retention Limit (\# fish) | Retention Limit Start Date | Retention Limit End Date |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 ${ }^{\text {A }}$ | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 31-Aug | 31-Dec | NA | NA | NA |
| 1984 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 1-Jan | 31-Dec | NA | NA | NA |
| 1985 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 1-Jan | 31-Dec | NA | NA | NA |
| 1986 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 1 -Jan | 31-Dec | NA | NA | NA |
| 1987 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 1-Jan | 31-Dec | NA | NA | NA |
| 1988 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 1-Jan | 31-Dec | NA | NA | NA |
| 1989 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 1-Jan | 31-Dec | NA | NA | NA |
| 1990 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 1-Jan | 31-Dec | NA | NA | NA |
| 1991 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | NA | 1-Jan | 31-Dec | NA | NA | NA |
| $1992{ }^{\text {B }}$ | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | $12^{\text {b }}$ | 1-Jan | 31-Dec | NA | NA | NA |
| 1993 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 12 | 1-Jan | 31-Dec | NA | NA | NA |
| 1994 | NA | NA | 365 | open | NA | 1-Jan | 31--Dec | 12 | 1-Jan | 31-Dec | NA | NA | NA |
| 1995 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 12 | 1-Jan | 31-Dec | NA | NA | NA |
| 1996 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 12 | 1-Jan | 31-Dec | NA | NA | NA |
| 1997 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 12 | 1-Jan | 31-Dec | NA | NA | NA |
| 1998 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 12 | 1-Jan | 31-Dec | NA | NA | NA |
| $1999{ }^{\text {C. D D }}$ | NA | NA | 250 | open | NA | 1-Jan | 23 -eb | 12 | 1-Jan | 23-Feb | NA | 1-Jan | 23-Feb |
|  |  |  |  |  |  | 24-Feb | 7-Sep | $14^{\text {c }}$ | 24-Feb | 7-Sep | $5^{\text {c }}$ | 24-Feb | 7-Sep |
|  |  |  | 115 | closed | emergency rule ${ }^{\circ}$ | 8 -Sep | 31-Dec |  |  |  |  |  |  |
| 2000 D. E | NA | NA | 239 | closed | emergency rule | 1-Jan | 27-Aug |  |  |  |  |  |  |
|  | NA | NA | 126 | open | NA | 28-Aug | 21-Sep | 14 | 28-Aug | 21-Sep | 5 | 28-Aug | 21-Sep |
|  |  |  |  | open | NA | 22-Sep | 31-Dec | 14 | 22-Sep | 31-Dec | $1{ }^{\text {E }}$ | 22-Sep | 31-Dec |
| 2001 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec | 1 | 1-Jan | 31-Dec |
| 2002 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec | 1 | 1-Jan | 31-Dec |
| 2003 | NA | NA | 365 | open | NA | 1-Jan | 31--Dec | 14 | 1-Jan | 31-Dec | 1 | 1-Jan | 31-Dec |
| 2004 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1 -Jan | 31-Dec |  | 1-Jan | 31-Dec |
| 2005 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec |  | 1-Jan | 31-Dec |
| 2006 F | NA | NA | 365 | open | NA | 1-Jan | 22-Oct | 14 | 1-Jan | 22-Oct | 1 | 1-Jan | 22-Oct |
|  |  |  |  |  |  | 23-Oct | 31-Dec | 14 | 23-Oct | 31-Dec | $3^{\text {F }}$ | 23-0ct | 31-Dec |
| 2007 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec | 3 | 1-Jan | 31-Dec |
| 2008 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec | 3 | 1-Jan | 31-Dec |
| 2009 | NA | NA | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1 -Jan | 31-Dec | 3 | 1-Jan | 31-Dec |
| $2010^{6}$ | NA | NA | 365 | open | NA | 1-Jan | 14-Feb | 14 | 1-Jan | 14-Feb | 3 | 1-Jan | 14-Feb |
|  | $190,050 \mathrm{gw}{ }^{\text {G }}$ | NA |  |  |  | 15-Feb | 31-Dec | 14 | 15-Feb | 31-Dec | 3 | 15-Feb | 31-Dec |
| 2011 H | $190,050 \mathrm{gw}$ | NA | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec | 3 | 1-Jan | 31-Dec |
| $2012{ }^{1}$ | 190,050 gw | NA | 365 | open | NA | 1-Jan | $15-\mathrm{Apr}$ | 14 | 1-Jan | 15-Apr | 3 | 1-Jan | $15-\mathrm{Apr}$ |
|  | NA | 197,652 ww ${ }^{\text {1 }}$ |  |  |  | 16-Apr | 31-Dec | 14 | 16-Apr | 31-Dec | 3 | 16-Apr | 31-Dec |
| $2013{ }^{\text {J }}$ | NA | 197,652 ww | 365 | open | NA | 1-Jan | 4-Sep | 14 | 1-Jan | 4-Sep | 3 | 1-Jan | 4-Sep |
|  |  | $153,000 \mathrm{ww}^{\text {J }}$ |  |  |  | 5-Sep | 31-Dec | 14 | 5-Sep | 31-Dec | 3 | 5-Sep | 31-Dec |
| 2014 | NA | $154,500 \mathrm{ww}$ | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec | 3 | 1-Jan | 31-Dec |
| 2015 | NA | $164,000 \mathrm{ww}$ | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec | 3 | 1-Jan | 31-Dec |
| 2016 | NA | $164,000 \mathrm{ww}$ | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | 31-Dec | 3 | 1 -Jan | 31-Dec |
| 2017 | NA | $164,000 \mathrm{ww}$ | 365 | open | NA | 1-Jan | 31-Dec | 14 | 1-Jan | $31-\mathrm{Dec}$ | 3 | 1-Jan | $31-\mathrm{Dec}$ |

A: Original SAFMC FMP effective 8/31/1983-included the 4 " trawl mesh size regulation
B: Amendment 4 (effective date $1 / 1 / 92$ ) included establishment of recreational 12 in $T L$ size limit and rebuilding timeframe $<=10$ years (year $1=1991$
C: Amendment 9 (effective 2/24/99) included increase in recreational minimum size limit to 14 in TL and established recreational bag limit of 5 fish per person per day
D: Emergency Rule prohibited harvest and possession of red porgy from 9/8/1999 through $8 / 28 / 2000$
. Ament 12 (efiective $9 / 22 / 2000$ ) included rebuilding timeframe=18 years ( $1999=$ year 1 ); no sale of red porgy during Jan-Apri;) 4 fish recreational bag limit
F: Amendment 13 C (effective $10 / 23 / 06$ ) retained recreational 14 in TL size limit and increased bag limit to 3 per person per day or per trip, whichever is more restrictive
G: Amendment 15 B (effective $2 / 15 / 2010$ ) prohibited the sale of snapper grouper species harvested or pose
by vessels with a Federal charter vessel/headboat permit for South Atlantic snapper grouper regardless of where harvested.
Established allocations for red porgy ( $50 \%$ commercial $\& 50 \%$ recreational). Recreational quota $=190,050$ los gutted weigh.
H: Amedment 17 A (effective $3 / 3 / 2011$ ) required use of circle hooks when fishing for sg species with hook-and-line gear and natural baits north of 28 degrees N Latitude
I: Comprehensive ACL Amendment (effective 4/16/2012) established quota as ACL
J: Regulatory Amendment 18 (effective $9 / 5 / 13$ ) revised MSY, OY, ABC and ACLs
J: Regulatory Amend
gw $=$ gutted weight
$w w=w h o l e ~ w e i g h t ~$

### 2.2.1 Closures Due to Meeting Commercial Quota or Commercial/Recreational ACL

Commercial closure - 12/02/2013 - exceeded commercial ACL

## 2.3 . State Regulatory History

### 2.3.1 North Carolina:

There are currently no North Carolina state-specific regulations for red porgy. North Carolina has complemented federal regulations, including quota and/or annual catch limit closures, for all snapper grouper species via proclamation authority since January 1991, when rule 15A NCAC 03M . 0506 was first implemented:

## 15A NCAC 03M . 0506 SNAPPER-GROUPER

The Fisheries Director may, by proclamation, until September 1, 1991, impose any or all of the following restrictions in the fishery for species of the snapper-grouper complex listed in the South Atlantic Fishery Management Council Fishery Management Plan for the Snapper-Grouper Fishery of the South Atlantic Region:
(1) Specify size;
(2) Specify seasons;
(3) Specify areas;
(4) Specify quantity;
(5) Specify means/methods; and
(6) Require submission of statistical and biological data

History Note: Statutory Authority G.S. 113-134; 113-182; 113-221; 143B-289.4. Eff. January 1, 1991.

The rule was modified slightly to remove the phrase "until September 1, 1991" effective September 1, 1991. The first proclamation (FF-19-94) pertaining to red porgy was issued under the authority of this rule effective July 1, 1994 and established a 12-inch total length minimum size limit (both sectors).

Rule 15A NCAC 03M . 0506 remained unchanged until March 1, 1996 when species-specific regulations for all snapper grouper species were added to the proclamation authority contained in the rule. Specific to red porgy, the rule was amended to include the minimum size limit initially established in FF-19-94:

## 15A NCAC 03M . 0506 SNAPPER-GROUPER

(n) It is unlawful to possess red porgy (pink or silver snapper) less than 12 inches total length.

History Note: Statutory Authority G.S. 113-134; 113-182; 113-221; 143B-289.4. Eff. January 1, 1991. Amended eff. March 1, 1996; September 1, 1991.

In addition to the above change, rule 15A NCAC 03 M .0512 was implemented effective March 1 , 1996 and provided supplementary proclamation authority to the Fisheries Director to modify any existing size and harvest limits for species subject to interstate and federal management:

## 15A NCAC 03M . 0152 COMPLIANCE WITH FISHERY MANAGEMENT PLANS

In order to comply with management requirements incorporated in Federal Fishery Management Council Management Plans or Atlantic States Marine Fisheries Commission Management Plans, the Fisheries Director may, by proclamation, suspend the minimum size and harvest limits established by the Marine Fisheries Commission, and implement different minimum size and harvest limits. Proclamations issued under this Section shall be subject to approval, cancellation, or modification by the Marine Fisheries Commission at its next regularly scheduled meeting or an emergency meeting held pursuant to G.S. 113-221(e1).
History Note: Authority G.S. 113-134; 113-182; 143B-289.4; Eff. March 1, 1996.

Proclamation FF-11-99 was issued effective March 1, 1999 and established a prohibition on the sale and purchase of red porgy during the months of March and April to complement the federal spawning closure. This was subsequently incorporated into modifications to rule 15A NCAC 03M . 0506 that became effective in May 1999; these included additional restrictions on harvest of red porgy and movement of regulations to a different sub-item within the rule:

## 15A NCAC 03M . 0506 SNAPPER-GROUPER

(o) Red Porgy:
(1) It is unlawful to possess red porgy (pink or silver snapper) less than 14 inchestotal length.
(2) It is unlawful to possess more than five red porgy per person per day without a valid Federal Commercial Snapper-Grouper permit.
(3) It is unlawful to possess more than five red porgy per person per day during the months of March and April.
(4) It is unlawful to sell or purchase red porgy taken from waters under the jurisdiction of North Carolina or the South Atlantic Fishery Management Council during the months of March and April.

History Note: Statutory Authority G.S. 113-134; 113-182; 113-221; 143B-289.4. Eff. January 1, 1991. Amended eff. March 1, 1996; September 1, 1991. Temporary Amendment Eff.

December 23, 1996; Amended Eff. August 1, 1998; April 1, 1997; Temporary Amendment Eff.
May 24, 1999.

Proclamation FF-20-99 was issued effective September 15, 1999 which prohibited all commercial and recreational harvest and possession, complementing the federal emergency closure of the fishery.

On August 29, 2000 rule 15A NCAC 03M . 0506 was amended to reflect the reopening of the commercial and recreational fisheries and additional changes in red porgy harvest seasons and possession limits:

## 15A NCAC 03M . 0506 SNAPPER-GROUPER

(o) Red Porgy (Pagrus pagrus):
(1) It is unlawful to possess red porgy (pink or silver snapper) less than 14 inchestotal length.
(2) It is unlawful to possess more than one red porgy per person per day without a valid Federal Commercial Snapper-Grouper permit.
(3) It is unlawful to sell or offer for sale red porgy from January 1 through April 30.
(4) It is unlawful to land more than 50 pounds of red porgy from May 1 through December 31 in a commercial fishing operation.

History Note: Statutory Authority G.S. 113-134; 113-182; 113-221; 143B-289.4. Eff. January 1, 1991. Amended eff. March 1, 1996; September 1, 1991. Temporary Amendment Eff.

December 23, 1996; Amended Eff. August 1, 1998; April 1, 1997; Temporary Amendment Eff.
August 29, 2000; January 1, 2000; May 24, 1999.

No further modifications to rule 15A NCAC 03M . 0506 pertaining to red porgy were implemented. In 2002, North Carolina adopted its Inter-Jurisdictional Fishery Management Plan (IJ FMP), which incorporates all Atlantic States Marine Fisheries Commission and council- managed species by reference and adopts all federal regulations as minimum standards for management, as appropriate. In 2007, the statutorily-mandated five-year review of the IJ FMP began, with final adoption of the updated plan in 2008. Changes to the FMP included removal of all species-specific regulations from rule 15A NCAC 03M . 0506 effective October 1, 2008, and proclamation authority to implement changes for all species under federal or interstate management was moved to rule 15 A NCAC 03 M . 0512 .

Because the 2007/2008 review of the IJ FMP occurred during the time when additional changes in federal management of red porgy were implemented, several proclamations were issued in 2007 to suspend the relevant portions of rule 15A NCAC 03M . 0506 and issue compatible regulations to reflect changes in commercial and recreational possession limits. (Because of requirements that continuing rule suspensions require N.C. Marine Fisheries Commission review and approval, issuance of multiple proclamations was required):

- FF-19-2007, FF-38-2007, FF-46-2007, FF-55-2007, FF-60-2007, FF-11-2008, FF-55-2008, FF-632008: Suspended relevant portions of 15A NCAC 03M . 0506
- FF-20-2007, FF-39-2007, FF-42-2007, FF-47-2007, FF-56-2007, FF-60-2007, FF-10-2008, FF-542008, FF-64-2008: Implemented a three-fish/person daily possession limit for persons without a federal commercial snapper grouper permit; implemented a commercial trip limit of 120 red porgy (effective 2/26/2007 via FF-20-2007; maintained through subsequently dated proclamations)

Once the changes to rules 15 A NCAC 03 M .0506 and 03 M .0512 described above were implemented, proclamation FF-66-2008 was issued effective October 1, 2008 and contained all relevant commercial and recreational regulations for all snapper grouper species. The portion of the proclamation specific to red porgy is excerpted as follows:
VIII. Red Porgy
A. It is unlawful to possess red porgy less than 14 inches total length.
B. It is unlawful to possess more than three red porgies per person per day without avalid Federal Commercial Snapper-Grouper permit.
C. It is unlawful to sell or offer for sale or purchase red porgy from January 1 through April 30. D. It is unlawful for a vessel with a valid Federal Commercial Snapper-Grouper permit to possess or land more than 120 individual red porgy per vessel per trip from May 1 through December 31.
Because there have been no additional modifications to federal management of red porgy, the above regulations have been maintained in subsequent proclamations since 2008. Proclamation FF-66-2009 added the prohibition on sale of fish harvested under the recreational bag limit without a federal commercial snapper grouper permit (as per Amendment 15B) to the general regulations for the entire fishery. Future proclamations modified the construction of the regulations slightly to clarify commercial vs. recreational restrictions:

## Red Porgy

A. For recreational purposes:

1. It is unlawful to possess red porgy less than 14 inches total length.
2. It is unlawful to possess more than three red porgies per person per day without avalid Federal Commercial Snapper-Grouper permit.
B. For commercial purposes:
3. It is unlawful to possess red porgy less than 14 inches total length.
4. It is unlawful to sell or offer for sale or purchase red porgy from January 1 through April 30.
5. It is unlawful for a vessel with a valid Federal Commercial South Atlantic SnapperGrouper permit to possess or land more than 120 individual red porgy per vessel pertrip from May 1 through December 31.

An information update to the IJ FMP was completed and approved in November 2015 and contained no additional modifications to rules 15A NCAC 03M . 0506 and 15A NCAC 03M .0512. The only procedural modifications that have occurred are starting in 2013, proclamations establishing the size limits, possession limits and seasons for the upcoming calendar year ("seasonopening" proclamations) have been issued in December of the preceding year; and beginning in 2015, commercial and recreational regulations have been moved into separate proclamations for ease of use by the public. The most current Snapper Grouper proclamations, as well as previous versions from 2001 onward, can be found online using this link: http://portal.ncdenr.org $/ \mathrm{web} / \mathrm{mf} /$ proclamations. Proclamations issued prior to 2001 are contained in hard copy archives.

Tables 1 and 2 contain a summary of recreational and commercial regulations, respectively. Because many snapper grouper proclamations are issued throughout the year to complement federal management measures, only those proclamations that were issued specific to red porgy in any one year are listed.

The current versions of rules 15A NCAC 03 M .0506 and 15A NCAC 03 M .0512 are below:

## 15A NCAC 03M . 0506 SNAPPER-GROUPER COMPLEX

(a) In the Atlantic Ocean, it is unlawful for an individual fishing under a Recreational Commercial Gear License with seines, shrimp trawls, pots, trotlines or gill nets to take any species of the Snapper-Grouper complex.
(b) The species of the snapper-grouper complex listed in the South Atlantic Fishery Management Council Fishery Management Plan for the Snapper-Grouper Fishery of the South Atlantic Region are hereby incorporated by reference and copies are available via theFederal Register posted on the Internet at www.safmc.net and at the Division of Marine Fisheries, P.O. Box 769, Morehead City, North Carolina 28557 at no cost.
History Note: Authority G.S. 113-134; 113-182; 113-221; 143B-289.52;
Eff. January 1, 1991;
Amended Eff. April 1, 1997; March 1, 1996; September 1, 1991;
Temporary Amendment Eff. December 23, 1996;
Amended Eff. August 1, 1998; April 1, 1997;
Temporary Amendment Eff. January 1, 2002; August 29, 2000; January 1, 2000; May 24, 1999;
Amended Eff. October 1, 2008; May 1, 2004; July 1, 2003; April 1, 2003; August 1, 2002.

## 15A NCAC 03M . 0512 COMPLIANCE WITH FISHERY MANAGEMENT PLANS

(a) In order to comply with management requirements incorporated in Federal Fishery Management Council Management Plans or Atlantic States Marine Fisheries Commission Management Plans or to implement state management measures, the Fisheries Director may, by proclamation, take any or all of the following actions for species listed in the Interjurisdictional Fisheries Management Plan:
(1) Specify size;
(2) Specify seasons;
(3) Specify areas;
(4) Specify quantity;
(5) Specify means and methods; and
(6) Require submission of statistical and biological data.
(b) Proclamations issued under this Rule shall be subject to approval, cancellation, or modification by the Marine Fisheries Commission at its next regularly scheduled meeting or an emergency meeting held pursuant to G.S. 113-221.1.
History Note: Authority G.S. 113-134; 113-182; 113-221; 113-221.1; 143B-289.4;
Eff. March 1, 1996;
Amended Eff. October 1, 2008.

Table 2.3.1.1. North Carolina recreational red porgy regulations in state waters 1991-2018. (TL $=$ total length)

| Year | Season | Min. Size (TL) | Daily Possession Limit | Regulation(s) |
| :---: | :---: | :---: | :---: | :---: |
| 1991 | Year-round | n/a | n/a | 15A NCAC 03M . 0506 |
| 1992 | Year-round | n/a | $\mathrm{n} / \mathrm{a}$ | 15A NCAC 03M . 0506 |
| 1993 | Year-round | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | 15A NCAC 03M . 0506 |
| 1994 | Year-round | 12 inches | n/a | 15A NCAC 03M .0506/FF-19-94 (eff. 7/1/1994) |
| 1995 | Year-round | 12 inches | n/a | 15A NCAC 03M .0506/FF-19-94 |
| 1996 | Year-round | 12 inches | n/a | 15A NCAC 03M .0506/03M .0512/FF-19-94 |
| 1997 | Year-round | 12 inches | n/a | 15A NCAC 03M . $0506 / 03 \mathrm{M} .0512$ |
| 1998 | Year-round | 12 inches | n/a | 15A NCAC 03M .0506/03M .0512 |
| 1999* | Closed 9/15/1999 | 12/14 inches | 5 fish/person | 15A NCAC 03M .0506/03M .0512/FF-11-99, FF-20-99 |
| 2000** | Year-round | 14 inches | 1 fish/person | 15A NCAC 03M .0506/03M .0512 |
| 2001 | Year-round | 14 inches | 1 fish/person | 15A NCAC 03M .0506/03M . 0512 |
| 2002 | Year-round | 14 inches | 1 fish/person | 15A NCAC 03M .0506/03M . 0512 |
| 2003 | Year-round | 14 inches | 1 fish/person | 15A NCAC 03M .0506/03M . 0512 |
| 2004 | Year-round | 14 inches | 1 fish/person | 15A NCAC 03M .0506/03M .0512 |
| 2005 | Year-round | 14 inches | 1 fish/person | 15A NCAC 03M .0506/03M . 0512 |
| 2006 | Year-round | 14 inches | 1 fish/person | 15A NCAC 03M .0506/03M . 0512 |
| 2007*** | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-19-2007, FF-20-2007 |
| 2008 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-10-2008, FF-112008, FF-66-2008 |
| 2009 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-48-2009, FF-66-2009 |
| 2010 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-60-2010 |
| 2011 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-19-2011 |
| 2012 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-10-2012 |
| 2013 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-5-2013 |
| 2014 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-76-2013 |
| 2015 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-94-2014 |
| 2016 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-71-2015 |
| 2017 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-68-2016 |
| 2018 | Year-round | 14 inches | 3 fish/person | 15A NCAC 03M .0506/03M .0512/FF-57-2017(revised) |

*FF-11-99 established March/April spawning closure (effective 3/1/1999); minimum size limit increase and possession limit established in changes to 15A
NCAC 03M . 0506 (effective 5/24/1999); federal emergency closure complemented via FF-20-99 (effective 9/15/1999)
**Possession limit change effective August 29, 2000
***Possession limit change effective February 26, 2007

Table 2.3.1.2. North Carolina commercial red porgy regulations in state waters 1991-2018. (TL = total length)

| Year | Season | Min. Size (TL) | Trip/Possession Limit | Regulation(s) |
| :---: | :---: | :---: | :---: | :---: |
| 1991 | Year-round | n/a | n/a | 15A NCAC 03M . 0506 |
| 1992 | Year-round | n/a | n/a | 15A NCAC 03M . 0506 |
| 1993 | Year-round | n/a | n/a | 15A NCAC 03M . 0506 |
| 1994 | Year-round | 12 inches | $\mathrm{n} / \mathrm{a}$ | 15A NCAC 03M .0506/FF-19-94 (eff. 7/1/1994) |
| 1995 | Year-round | 12 inches | n/a | 15A NCAC 03M .0506/FF-19-94 |
| 1996 | Year-round | 12 inches | n/a | 15A NCAC 03M .0506/03M .0512/FF-19-94 |
| 1997 | Year-round | 12 inches | $\mathrm{n} / \mathrm{a}$ | 15A NCAC 03M . $0506 / 03 \mathrm{M} .0512$ |
| 1998 | Year-round | 12 inches | n/a | 15A NCAC 03M .0506/03M .0512 |
| 1999* | Closed March/April; fishery closure 9/15/1999 | 12/14 inches |  | 15A NCAC 03M .0506/03M .0512/FF-11-99, FF-20-99 |
| 2000** | Closed January -April | 14 inches | 50 pounds | 15A NCAC 03M .0506/03M .0512 |
| 2001 | Closed January -April | 14 inches | 50 pounds | 15A NCAC 03M .0506/03M .0512 |
| 2002 | Closed January -April | 14 inches | 50 pounds | 15A NCAC 03M .0506/03M .0512 |
| 2003 | Closed January -April | 14 inches | 50 pounds | 15A NCAC 03M .0506/03M .0512 |
| 2004 | Closed January -April | 14 inches | 50 pounds | 15A NCAC 03M .0506/03M .0512 |
| 2005 | Closed January -April | 14 inches | 50 pounds | 15A NCAC 03M .0506/03M .0512 |
| 2006 | Closed January -April | 14 inches | 50 pounds | 15A NCAC 03M .0506/03M .0512 |
| 2007*** | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-19-2007, FF-20-2007 |
| 2008 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-10-2008, FF-112008, FF-66-2008 |
| 2009 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-48-2009, FF-66-2009 |
| 2010 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-60-2010 |
| 2011 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-19-2011 |
| 2012 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-10-2012 |
| 2013^ | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-5-2013, FF-64-2013 |
| 2014 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-76-2013 |
| 2015 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-93-2014 |
| 2016 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-70-2015 |
| 2017 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-67-2016 |
| 2018 | Closed January -April | 14 inches | 120 fish | 15A NCAC 03M .0506/03M .0512/FF-58-2017 |

*FF-11-99 established March/April spawning closure (effective 3/1/1999); minimum size limit increase and possession limit established in changes to 15A
NCAC 03M . 0506 (effective 5/24/1999); federal emergency closure complemented via FF-20-99 (effective 9/15/1999)
**Effective August 29, 2000
***Effective February 26, 2007
${ }^{\wedge}$ Commercial closure due to annual catch limit being met (effective December 2, 2013)

### 2.3.2 South Carolina:

1992: SC Code of Laws Section 50-17-510(C) adopted the federal minimum size limits automatically for all species managed under the Fishery Conservation and Management Act (PL94265); and Section 50-17-510(F) adopted the federal catch and possession limits for a number of listed species managed under the Fishery Conservation and Management Act (PL94265) as the Law of the State of SC, with all managed species of porgy specifically mentioned.

2001: SC Marine-related Laws reorganized under SC Code of Laws Title 50 Chapter 5.
SC Code of Laws Section 50-5-2730 reads - "Unless otherwise provided by law, any regulations promulgated by the federal government under the Fishery Conservation and Management Act (PL94-265) or the Atlantic Tuna Conservation Act (PL 94-70) which establishes seasons, fishing periods, gear restrictions, sales restrictions, or bag, catch, size, or possession limits on fish are declared to be the law of this State and apply statewide including in state waters." As such, SC red porgy-related regulation is pulled directly from the federal regulations as promulgated under Magnuson. No changes have been made to this approach in covering red porgy since the Chapter 5 rewrite.

### 2.3.3 Georgia:

In Georgia current regulations for Red Porgy are 3 fish per person, 14 inch TL, open all year (GA DNR Reg, 391-2-4-. 04 (3)(n)).

### 2.3.4 Florida:

Atlantic Red Porgy Regulation History

| Year | $\frac{\text { Minimum Size }}{\underline{\text { Limit }}}$ | Recreational Daily Harvest Limits | Commercial Daily Harvest Limits | Regulation Changes | $\begin{gathered} \frac{\text { Rule Change }}{} \\ \frac{\text { Effective }}{\text { Date }} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | None | None | None |  |  |
| 1981 | None | None | None |  |  |
| 1982 | None | None | None |  |  |
| 1983 | None | None | None |  |  |
| 1984 | None | None | None |  |  |
| 1985 | None | None | None |  |  |
| 1986 | None | None | None |  |  |
| 1987 | None | 2 fish or 250 pounds per person, whichever is greater | None |  |  |
| 1988 | None | 2 fish or 250 pounds per person, whichever is greater | None |  |  |


| 1989 | None | 2 fish or 100 pounds per person, whichever is greater | None |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | None | 2 fish or 100 pounds per person, whichever is greater | None |  |  |
| 1991 | None | 2 fish or 100 pounds per person, whichever is greater | None |  |  |
| 1992 | None | 2 fish or 100 pounds per person, whichever is greater | None |  |  |
| 1993 | None | 2 fish or 100 pounds per person, whichever is greater | None |  |  |
| 1994 | 12 inches TL | 2 fish or 100 <br> pounds per person, whichever is greater | None | Established a minimum size limit of 12 inches for red porgy in Atlantic state waters. <br> Modified rule language to provide the same definitions of Gulf of Mexico and Atlantic Ocean regions. | $\begin{gathered} \text { March 1, } \\ 1994 \end{gathered}$ |
| 1995 | 12 inches TL | 2 fish or 100 pounds per person, whichever is greater | None |  |  |
| 1996 | 12 inches TL | 2 fish or 100 pounds per person, whichever is greater | None |  |  |


| 1997 | 12 inches TL | 2 fish or 100 pounds per person, whichever is greater | None |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 14 inches TL | 5 fish per person | 5 fish per person | Increased the minimum size limit for Atlantic red porgy to 14 inches TL. <br> Established a recreational bag limit for red porgy of 5 fish per person in Atlantic state waters. Allowed a two-day possession limit for red porgy for persons aboard charter and headboats on trips exceeding 24 hours provided the vessel is equipped with a permanent berth for each passenger aboard, and each passenger has a receipt verifying the trip length. Prohibited the harvest and sale of Atlantic red porgy in excess of the bag limit. <br> Prohibited the sale of Atlantic red porgy in March and April. Established that if commercial harvest of red porgy is closed in adjacent federal waters, commercial harvest will close in state waters five days after the federal closure date and remain closed until federal waters reopen. Required that all reef fish species managed in Florida, including red porgy, be landed in whole condition. Designated allowable gear for all reef fish species, including red porgy, as hook-and-line, black sea bass traps, and spearing (does not include powerheads, bangsticks, or handheld devices employing an explosive charge). | $\begin{aligned} & \text { December } \\ & 31,1998 \end{aligned}$ |


| $\begin{aligned} & 1998 \\ & \text { cont. } \end{aligned}$ | 14 inches TL | 5 fish per person | 5 fish per person | Retention is limited to the recreational bag and possession limits when red porgy is harvested as incidental bycatch with gear that is not allowed. <br> Designated red porgy as a "restricted species." This means harvesters must possess a Florida Saltwater Products License and Restricted Species Endorsement, as well as a federal snapper grouper commercial permit, to exceed the recreational bag limit and sell reef fish. | $\begin{aligned} & \text { December 31, } \\ & 1998 \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 14 inches TL | 5 fish per person | 5 fish per person | Closed Atlantic state waters to the recreational harvest of red porgy through March 5, 2000. | $\begin{gathered} \text { October 22, } \\ 1999 \end{gathered}$ |
| 2000 | Closed | Closed | Closed | (1) Eliminated the 5-day commercial closure extension. <br> (2) Prohibited all harvest of red porgy from Atlantic state waters. | (1) January <br> 1, 2000 <br> (2) March 6, 2000 |
| 2001 | 14 inches TL | 1 fish per person | $\begin{aligned} & 50 \text { lbs. per } \\ & \text { vessel } \end{aligned}$ | Allowed a one-fish per person daily recreational bag limit and a 50-pound commercial vessel limit for Atlantic red porgy. <br> Established a minimum size limit of 14 inches total length for Atlantic red porgy. <br> Prohibited commercial harvest and sale of Atlantic red porgy January through April. Permitted persons harvesting other species for commercial purposes during the closure to harvest and possess the recreational bag limit of red porgy. | $\begin{gathered} \text { March 1, } \\ 2001 \end{gathered}$ |


| 2002 | 14 inches TL | 1 fish per person | 50 lbs. per vessel |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 14 inches TL | 1 fish per person | 50 lbs. per vessel |  |  |
| 2004 | 14 inches TL | 1 fish per person | 50 lbs. per vessel |  |  |
| 2005 | 14 inches TL | 1 fish per person | 50 lbs . per vessel |  |  |
| 2006 | 14 inches TL | 1 fish per person | 50 lbs. per vessel | Provided that, for purposes of determining the legal size of reef fish species, "total length" means the straight-line distance from the most forward point of the head with the mouth closed, to the farthest tip of the tail with the tail compressed or squeezed, while the fish is lying on its side. | July 1, 2006 |
| 2007 | 14 inches TL | 3 fish per person | 50 lbs. per vessel | Increased the daily recreational bag limit for Atlantic red porgy to 3 fish per person. | July 1, 2007 |
| 2008 | 14 inches TL | 3 fish per person | $\begin{gathered} 50 \text { lbs. per } \\ \text { vessel } \end{gathered}$ |  |  |
| 2009 | 14 inches TL | 3 fish per person | $\begin{gathered} 50 \text { lbs. per } \\ \text { vessel } \end{gathered}$ |  |  |
| 2010 | 14 inches TL | 3 fish per person | 50 lbs. per vessel | Required dehooking tools to be aboard commercial and recreational vessels for anglers to use as needed to remove hooks from Atlantic reef fish. | $\begin{aligned} & \text { January 19, } \\ & 2010 \end{aligned}$ |
| 2011 | 14 inches TL | 3 fish per person | 50 lbs. per vessel |  |  |
| 2012 | 14 inches TL | 3 fish per person | 50 lbs. per vessel | Removed red porgy from the exception allowing a two-day possession limit for reef fish statewide for persons aboard charter <br> and headboats on trips exceeding 24 hours. | July 1, 2012 |
| 2013 | 14 inches TL | 3 fish per person | $\begin{gathered} 50 \mathrm{lbs} \text { per } \\ \text { vessel } \end{gathered}$ |  |  |
| 2014 | 14 inches TL | 3 fish per person | 50 lbs . per vessel |  |  |


| 2015 | 14 inches TL | 3 fish per <br> person | 50 lbs. per <br> vessel |  |  |
| :---: | :---: | :---: | :---: | :--- | :--- |
| 2016 | 14 inches TL | 3 fish per <br> person | 50 lbs. per <br> vessel |  |  |
| 2017 | 14 inches TL | 3 fish per <br> person | 50 lbs. per <br> vessel |  |  |
| 2018 | 14 inches TL | 3 fish per <br> person | 50 lbs. per <br> vessel |  |  |

## References

None provided.

## 3. Assessment History

An early stock assessment of South Atlantic red porgy (Vaughan et al. 2001), conducted before the SEDAR process existed used age-aggregated and age-structured production models to determine status of the stock and the fishery. This assessment found the stock to be overfished $\left(\mathrm{SSB}_{\mathbf{2 0 0 0}} \operatorname{MSST}=0.13-\right.$ 0.25 ) but not undergoing overfishing ( $F_{\mathbf{2 0 0 0}} / F_{\mathbf{M S Y}}=0.34-0.44$ ).

The first SEDAR stock assessment of red porgy was a benchmark assessment which used the fully agestructured Beaufort Assessment Model (BAM) to model the population from 1972-2001 (SEDAR 01; SEDAR 2002). As of 2001, the stock was overfished $\left(\mathrm{SSB}_{\mathbf{2 0 0 1}} / \mathrm{MSST}=0.55 ; \mathrm{SSB}_{\mathbf{2 0 0 1}} / \mathrm{SSB}_{\mathbf{M s Y}}=0.43\right)$, but overfishing was not occurring ( $F_{\mathbf{2 0 0 1}} / F_{\mathbf{M s Y}}=0.45$; SEDAR 2002). Following the benchmark, there have been two update assessments (SEDAR 2006; 2012) prior to the current standard assessment, both of which have also used the BAM as the primary model. Assessment model timelines for the update assessments are as follows: SEDAR 1, 2006 Update assessment (1972-2004; SEDAR 2006) and SEDAR 1, 2012 Update assessment (1972-2011; SEDAR 2012). The SEDAR 1, 2006 Update assessment found that the red porgy stock was not rebuilt $\left(\mathrm{SSB}_{\mathbf{2 0 0 5}} / \mathrm{SSB}_{\mathbf{M S Y}}=0.66\right)$, but was not undergoing overfishing $\left(F_{\mathbf{2 0 0 4}} / F_{\mathbf{M s Y}}=0.45\right)$. The SEDAR 1, 2012 Update assessment also found that the red porgy stock was not rebuilt $\left(\mathrm{SSB}_{\mathbf{2 0 1 1}} / \mathrm{SSB}_{\mathbf{M S Y}}=0.47\right)$ but was not undergoing overfishing $\left(F_{\mathbf{2 0 0 9 - 2 0 1 1}} / F_{\mathbf{M S Y}}=0.64\right)$.

Input values of constant $M$ were the same for the three previous red porgy assessments (terminal years: 2001, 2004, 2011; $M: 0.225,0.225,0.225)$. Steepness was estimated in all assessments and has decreased between the 2006 and 2012 updates ( $h: 0.48,0.50,0.41$ ). Estimates of $F_{\mathbf{M s Y}}$ have remained in a similar range ( $F_{\text {MsY }}: 0.19,0.20,0.17$ ). Estimates of MSY have fluctuated [MSY ( 1000 lb ): 826, 626, 834] and estimates of SSB $_{\mathbf{M S Y}}$ have generally increased over the course of the three previous SEDAR assessments [ $\mathrm{SSB}_{\mathbf{M S Y}}$, (mt): 3050, 3236, 3933].

## References

SEDAR, 2002. SEDAR 1: Stock Assessment of South Atlantic Red Porgy.
SEDAR, 2006. SEDAR 1 Update Assessment: Stock Assessment of Red Porgy off the Southeastern United States.
SEDAR, 2012. Stock Assessment of Red Porgy off the Southeastern United States: SEDAR Update Assessment.
Vaughan, D. S., E. H. Williams, and M. H. Prager. 2001. Updated status of red porgy off southeastern United States. Unpublished manuscript dated November 7, 2001. NOAA Center for Coastal Fisheries and Habitat Research, Beaufort, NC 28516. Prepared for the South Atlantic Fishery Management Council. One Southpark Circle, Suite 306. Charleston, SC 29407. .

## 4. Regional Maps



Figure 4.1: South Atlantic Fishery Management Council and EEZ boundaries.

## 5. Abbreviations

| APAIS | Access Point Angler Intercept Survey |
| :--- | :--- |
| ABC | Allowable Biological Catch |
| ACCSP | Atlantic Coastal Cooperative Statistics Program |
| ADMB | AD Model Builder software program |
| ALS | Accumulated Landings System; SEFSC fisheries data collection program |
| AMRD | Alabama Marine Resources Division |
| ASMFC | Atlantic States Marine Fisheries Commission |
| ASPIC | a stock production model incorporating covariates |
| ASPM | age-structured production model |
| B | stock biomass level |
| BAM | Beaufort Assessment Model |
| BMSY | value of B capable of producing MSY on a continuing basis |
| CFMC | Caribbean Fishery Management Council |
| CIE | Center for Independent Experts |
| CPUE | catch per unit of effort |
| EEZ | exclusive economic zone |
| F | fishing mortality (instantaneous) |
| FMSY | fishing mortality to produce MSY under equilibrium conditions |
| FOY | fishing mortality rate to produce Optimum Yield under equilibrium |
| FXX\% SPR | fishing mortality rate that will result in retaining XX\% of the maximum spawning production |
| under equilibrium conditions |  |


| LDWF | Louisiana Department of Wildlife and Fisheries <br> M |
| :--- | :--- |
| natural mortality (instantaneous) |  |
| MAFMC | Mid-Atlantic Fishery Management Council |
| MARMAP | Marine Resources Monitoring, Assessment, and Prediction |
| MDMR | Mississippi Department of Marine Resources <br> maximum fishing mortality threshold, a value of F above which overfishing is deemed to be <br> mFMT |
| occurring |  |
| MRFSS | Marine Recreational Fisheries Statistics Survey; combines a telephone survey of households to <br> estimate number of trips with creel surveys to estimate catch and effort per trip |
| MRIP | Marine Recreational Information Program |
| MSST | minimum stock size threshold, a value of B below which the stock is deemed to be overfished <br> MSY |
| maximum sustainable yield |  |



## SEDAR

## Southeast Data, Assessment, and Review

## SEDAR 60

# South Atlantic Red Porgy 

## Section II: Assessment Report

April 2020
SEDAR
4055 Faber Place Drive, Suite 201 North Charleston, SC 29405

## Contents

1 Introduction ..... 8
1.1 Executive Summary ..... 8
1.2 Workshop Time and Place ..... 8
1.3 Terms of Reference ..... 9
1.4 List of Participants ..... 10
1.5 Document List ..... 12
1.6 Statements Addressing Each Term of Reference ..... 14
2 Data Review and Update ..... 16
2.1 Data Review ..... 16
2.2 Data Update ..... 17
2.2.1 Life History ..... 17
2.2.2 Life History Tables and Figures ..... 19
2.2.3 Commercial Landings and Discards ..... 22
2.2.4 Commercial Landings and Discards Tables and Figures ..... 25
2.2.5 Recreational Fisheries ..... 29
2.2.6 Recreational Fisheries Tables ..... 31
2.2.7 Indices of Abundance ..... 33
2.2.8 Length Compositions ..... 33
2.2.9 Age Compositions ..... 34
2.2.10 Discard Mortality ..... 34
2.2.11 Discard Mortality Tables ..... 36
3 Stock Assessment Methods ..... 38
3.1 Overview ..... 38
3.2 Data Sources ..... 38
3.3 Model Configuration and Equations ..... 39
3.4 Parameters Estimated ..... 42
3.5 Per Recruit and Equilibrium Analyses ..... 42
3.6 Benchmark/Reference Point Methods ..... 43
3.7 Comparison to Previous Assessments ..... 44
3.8 Uncertainty and Measures of Precision ..... 45
3.8.1 Bootstrapping of Observed Data ..... 45
3.8.2 Monte Carlo Sampling ..... 46
3.9 Projection Methods ..... 46
3.9.1 Initialization of Projections ..... 46
3.9.2 Uncertainty of Projections ..... 47
3.9.3 Projection Scenarios ..... 47
4 Stock Assessment Results ..... 48
4.1 Measures of Overall Model Fit ..... 48
4.2 Parameter Estimates ..... 48
4.3 Stock Abundance and Recruitment ..... 48
4.4 Total and Spawning Biomass ..... 48
4.5 Selectivity ..... 48
4.6 Fishing Mortality, Landings, and Discards ..... 49
4.7 Spawner-Recruitment Parameters ..... 49
4.8 Per Recruit and Equilibrium Analyses ..... 49
4.9 Benchmarks / Reference Points ..... 49
4.10 Status of the Stock and Fishery ..... 50
4.11 Comparison to Previous Assessments ..... 50
4.12 Sensitivity Analyses ..... 51
4.13 Retrospective Analyses ..... 51
4.14 Projections ..... 52
5 Discussion ..... 52
5.1 Comments on Assessment Results ..... 52
5.2 Comments on Projections ..... 53
6 Research Recommendations ..... 53
7 References ..... 54
8 Tables ..... 58
9 Figures ..... 84
Appendices ..... 137
A Abbreviations and Symbols ..... 137
B ADMB Parameter Estimates ..... 138

## List of Tables

1 Point estimates of natural mortality ( $M$ ) ..... 19
2 Age specific estimates of natural mortality ( $M$ ) ..... 20
3 Annual commercial landing totals of Red Porgy ..... 25
4 Yearly calculated total discards of Red Porgy from vertical line vessels ..... 26
5 Comparison of Red Porgy discard rates by year ..... 27
6 Red Porgy recreational landings ..... 31
7 Red Porgy recreational discards ..... 32
8 Red porgy commercial hook-and-line total discard mortality estimates ..... 36
9 Red porgy recreational hook-and-line total discard mortality estimates ..... 37
10 Observed time series of landings and discards ..... 59
11 Observed time series of CVs used in ensemble modeling (MCB) for landings and discards. ..... 60
12 Sample sizes of length and age compositions (numbers of trips). ..... 61
13 Sample sizes of length and age compositions (numbers of fish). ..... 62
14 Observed time series of indices of abundance. ..... 63
15 Life history at age. ..... 64
16 Estimated total abundance at age (1000 fish) ..... 65
17 Estimated total abundance at age (mt) ..... 66
18 Estimated total abundance at age (1000 lb) ..... 67
19 Estimated time series of status indicators, fishing mortality, biomass, and sex ratio. ..... 68
20 Selectivities by survey or fleet ..... 69
21 Estimated time series of fully selected fishing mortality rates by fleet. ..... 70
22 Estimated instantaneous fishing mortality rate. ..... 71
23 Estimated total landings at age in numbers (1000 fish) ..... 72
24 Estimated total landings at age in whole weight (1000 lb) ..... 73
25 Estimated time series of landings in numbers (1000 fish) by fleet. ..... 74
26 Estimated time series of landings in weight (1000 lb) by fleet. ..... 75
27 Estimated time series of dead discards in numbers (1000 fish) by fleet. ..... 76
28 Estimated time series of dead discards in weight (1000 lb) by fleet. ..... 77
29 Estimated status indicators and benchmarks ..... 78
30 Results from sensitivity runs. ..... 79
31 Projection results for $F=F_{\mathrm{P}_{50 \%}^{*}}$ ..... 80
32 Projection results for $F=F_{\mathrm{MSY}}$ ..... 81
33 Projection results for $F=0.75 F_{\mathrm{MSY}}$ ..... 82
34 Projection results for $F=0$ ..... 83
35 Abbreviations and symbols ..... 137
36 Parameter estimates from the BAM base run ..... 138

## List of Figures

1 Population growth model of Red Porgy ..... 21
2 Frequency plot of Red Porgy commercial discard rates by year ..... 28
3 Length, female maturity, and reproductive output at age. ..... 85
4 Observed and estimated annual age and length compositions by fleet ..... 86
5 Observed and estimated landings: commercial handline ..... 92
6 Observed and estimated landings: commercial trawl ..... 93
7 Observed and estimated landings: recreational MRIP ..... 94
8 Observed and estimated landings: recreational headboat ..... 95
9 Observed and estimated discards: commercial handline ..... 96
10 Observed and estimated discards: recreational MRIP ..... 97
11 Observed and estimated discards: recreational headboat ..... 98
12 Observed and estimated index of abundance: recreational headboat ..... 99
13 Observed and estimated index of abundance: Chevron trap/video survey ..... 100
14 Estimated annual abundance (numbers) at age ..... 101
15 Estimated annual biomass (weight) at age ..... 102
16 Estimated recruitment time series ..... 103
17 Estimated total and spawning stock biomass time series ..... 104
18 Selectivity by fleet: SERFS Chevron trap/video index ..... 105
19 Selectivity by fleet: commercial landings ..... 106
20 Selectivity by fleet: recreational landings ..... 107
21 Selectivity by fleet: discards ..... 108
22 Average selectivity from the terminal assessment year ..... 109
23 Estimated fully selected fishing mortality rates by fleet ..... 110
24 Estimated landings in numbers by fleet ..... 111
25 Estimated landings in weight by fleet ..... 112
26 Estimated discards in numbers by fleet ..... 113
27 Estimated discards in weight by fleet ..... 114
28 Spawner recruit curve ..... 115
29 Probability densities of spawner-recruit quantities ..... 116
30 Yield per recruit and spawning potential ratio at $F$ ..... 117
31 Equilibrium removals and spawning stock at $F$ ..... 118
32 Probability densities of MSY-related benchmarks from MCB analysis ..... 119
33 Estimated time series of SSB and $F$ relative to benchmarks ..... 120
34 Probability densities of terminal status estimates from MCB analysis ..... 121
35 Phase plot of terminal status estimates from MCB analysis ..... 122
36 Age structure relative to the equilibrium expected at $F_{\mathrm{MSY}}$. ..... 123
37 Sensitivity to natural mortality: sensitivity runs S1-S2 ..... 124
38 Sensitivity to steepness: sensitivity runs S3-S4 ..... 125
39 Sensitivity to R0: sensitivity run S5 ..... 126
40 Sensitivity to Florida Snapper Trap Index and Age Comps: sensitivity run S6 ..... 127
41 Sensitivity to time-varying female maturity: sensitivity run S7 ..... 128
42 Sensitivity to upweighting the headboat index: sensitivity runs S8-S9 ..... 129
43 Sensitivity to 2016 MRIP landings and discards: sensitivity run S10 ..... 130
44 Sensitivity status phase plots: S1-S10 ..... 131
45 Retrospective plots ..... 132
46 Projections with fishing mortality rate at fixed $F$ that provides $P^{*}=0.50$ ..... 133
47 Projections with fishing mortality rate fixed at $F=F_{\text {MSY }}$ ..... 134
48 Projections with fishing mortality rate fixed at $F=75 \% F_{\text {MSY }}$ ..... 135
49 Projections with fishing mortality rate fixed at $F=0$ ..... 136

## 1 Introduction

### 1.1 Executive Summary

This standard assessment evaluated the stock of Red Porgy (Pagrus pagrus) off the southeastern United States. The primary objectives of this assessment were to update the 2002 SEDAR-1 benchmark, and 2006 and 2012 update assessments of Red Porgy and to conduct fresh stock projections. Data compilation and assessment methods were guided by methods used in previous Red Porgy assessments. The benchmark assessment included data from 1972-2001, the 2006 update contained data through 2004, the 2012 update included data through 2011 and this assessment contained data through 2017. This assessment was conducted by the Southeast Fisheries Science Center in cooperation with regional data providers.

Available data on this stock included indices of abundance, landings, discards, and samples of annual length compositions and age compositions from fishery-dependent and fishery-independent sources. Two indices of abundance were developed during the SEDAR process and fitted by the model: one from the NMFS headboat survey and one from the fishery-independent SouthEast Reef Fish Survey (SERFS: MARMAP, SEAMAP-SA, and SEFIS) program, combined chevron trap and video data. Landings data were available from all recreational and commercial fleets.

The model used in all previous assessments of this stock-and updated here-was the Beaufort Assessment Model (BAM), a statistical catch-age formulation. A base run of BAM was configured to provide estimates of key management quantities, such as stock and fishery status. Uncertainty in estimates from the base run was evaluated through a mixed Monte Carlo/Bootstrap (MCB) procedure.

Results suggest that spawning stock biomass has decreased considerably since the terminal year of the previous assessment (2011) and the 2017 value of $\mathrm{SSB}\left(\mathrm{SSB}_{2017}=780 \mathrm{klb}\right)$ was below the minimum stock size threshold (MSST $=2249 \mathrm{klb}$ ) using the Council's definition of MSST as $(1-M) \mathrm{SSB}_{\mathrm{MSY}}$ and assuming a natural mortality rate of $M=0.22$. This resulted in a terminal stock status estimate of $\mathrm{SSB}_{2017} / \mathrm{MSST}=0.347$ and rebuild status estimate of $\mathrm{SSB}_{2017} / \mathrm{SSB}_{\mathrm{MSY}}=0.27$. Though fishing mortality $F$ was generally below or near $F_{\mathrm{MSY}}=0.18$ between 2009 and 2015 , it was substantially higher and above $F_{\text {MSY }}$ in 2016 and 2017 , with the terminal $F$-status estimate $F_{2015-2017} / F_{\mathrm{MSY}}=1.73$. Recruitment has generally been declining throughout the time series, and has been below the recruitment level corresponding to MSY ( $R_{\text {MSY }}$ ) for most of the past three decades. Thus, this assessment indicates that the stock is overfished, and undergoing overfishing.

The MCB analysis indicates that these estimates of stock and fishery status are robust, with little uncertainty in the conclusions. Of all MCB runs, $100 \%$ were in qualitative agreement that the stock is overfished ( $\mathrm{SSB}_{2017} / \mathrm{MSST}<1.0$ ), and $98.2 \%$ that the stock is experiencing overfishing $\left(F_{2015-2017} / F_{\mathrm{MSY}} \geq 1.0\right)$.

The estimated trends from this standard assessment are similar to those from the SEDAR 1, 2002 Benchmark, and the 2006 and 2012 updates. However, this assessment did show some differences from previous assessments, which was not surprising, given modifications made to both the data and model (described throughout the report).

### 1.2 Workshop Time and Place

The SEDAR 60 South Atlantic Red Porgy Assessment took place over a series of webinars held January 29, 2018; March 25, 2019; November 15, 2019; January 22, 2020; February 6, 2020; and February 28, 2020 and an in-person workshop held December 10-12, 2019 at Beaufort NC.

### 1.3 Terms of Reference

1. Prepare a standard assessment, based on the approved 2012 SEDAR 1 South Atlantic Red Porgy Update assessment with data through 2017. Provide commercial and recreational landings and discards in pounds and numbers.
2. Evaluate and document the following specific changes in input data or deviations from the update model. (List below each topic or new dataset that will be considered in this assessment.)

- Consider including the SERFS video index
- Incorporate the latest BAM model configurations and updates to data calculation methodologies, detailing the changes made and the impacts of those changes, between the 2012 SEDAR 1 South Atlantic Red Porgy Update assessment model and the proposed SEDAR 60 model.
- Re-consider use of age and length composition data.

3. Document any changes or corrections made to the model and input datasets and provide updated input data tables. Fully document and describe the impacts (on population parameters and management benchmarks) of any changes to the model structure, methods, application or fitting procedures made between this assessment and the 2012 SEDAR 1 South Atlantic Red Porgy Update assessment.
4. Update model parameter estimates and their variances, model uncertainties, and estimates of stock status and management benchmarks. Compare population parameter trends and management benchmarks estimated in this assessment with values from the previous assessment, and comment on the impacts of changes in data, assumptions or assessment methods on estimated population conditions and benchmarks.
5. Provide stock projections, including a pdf (probability density function) for biological reference point estimates and yield separated for landings and discards reported in pounds and numbers. Projection results are required through 2024, with projected fishing level changes beginning in late 2019. The panel shall provide guidance on appropriate assumptions to address harvest and mortality levels in the interim years between the assessment terminal year (2017) and the first year of management (2019). Projection criteria:

- To determine OFL: (1) $\mathrm{P}^{*}=50 \%$; (2) $F_{\mathrm{MSY}}$
- To evaluate the existing rebuilding plan: base on fixed exploitation at $75 \% F_{\text {MSY }}$. In addition to reporting yield and stock status as described above, for this projection also report the probability that $\mathrm{SSB}>$ $\mathrm{SSB}_{\mathrm{MSY}}$.

6. Review, evaluate, and report on the status and progress of all research recommendations listed in the last assessment, peer review reports, and SSC report concerning this stock.
7. Develop a stock assessment update report to address these TORS and fully document the input data, methods, and results of the stock assessment update.

### 1.4 List of Participants

## Appointee

## ANALYTICAL TEAM

Nikolai Klibansky
Rob Cheshire
Kyle Shertzer
Erik Williams

PANELISTS

| Nate Bacheler | Data provider |
| :--- | :--- |
| Wally Bubley | Data Provider |
| Scott Crosson* | SSC |
| Julie DeFilippi-Simpson* | Data Provider |
| Joe Evans | Data Provider |
| Kelly Fitzpatrick | Data Provider |
| Vivian Matter* | Data provider |
| Kevin McCarthy* | Data provider |
| Jennifer Potts | Data provider |
| Marcel Reichert* | SSC |
| Fred Scharf | SSC |
| George Sedberry | SSC |
| Amanda Tong | Data provider |
| Beth Wrege* | Data provider |
| Dave Wyanski | Data Provider |

APPOINTED OBSERVERS

| Jack Cox* | Fisherman | NC; SG AP |
| :--- | :--- | :--- |
| Kenny Fex* | Fisherman | NC |
| Bobby Freeman* | Fisherman | NC; SG AP |
| Ben Hartig* | Fisherman | FL |

## APPOINTED COUNCIL MEMBERS

Tim Griner

STAFF
Myra Brouwer*
Julia Byrd*
Mike Errigo
Cierra Graham
Kathleen Howington
Jeff Pulver

Council member

| Council lead | SAFMC |
| :--- | :--- |
| Citizen Science Coordinator | SAFMC |
| Fishery Biologist | SAFMC |
| Administrator | SAFMC |
| Coordinator | SEDAR |
| Fishery Biologist | SERO |

Appointees marked with an * were unable to attend the in-person workshop

| Appointee | Function | Affiliation |
| :--- | :--- | :--- |
| OTHER |  |  |
| Roger Brothers | Observer | SEFSC Beaufort |
| Matt Damiano | Observer | NC State |
| Eric Fitzpatrick | Observer | SEFSC Beaufort |
| Dalton Knight | Observer | SEFSC Beaufort |
| Stephen Long | Observer | SCDNR |
| Stephanie Martinez | Observer |  |
| Tracy McCulloch | Observer | SEFSC Beaufort |
| Andy Ostrowski | Observer | SEFSC Beaufort |
| Cassidy Peterson | Observer | SEFSC Beaufort |
| Walter Rogers | Observer | SEFSC Beaufort |
| Mclean Seward | Observer | NCDMF |
| Tracy Smart | Observer | SCDNR |
| Kevin Spanik | Observer | SCDNR |
| NON-PANEL DATA PROVIDERS | SEFSC Miami |  |
| Larry Beerkircher | Data Provider | NC DMF |
| Alan Bianchi | Data Provider | SEFSC Beaufort |
| Ken Brennan | Data Provider | FL FWCC |
| Steve Brown | Data Provider | GA DNR |
| Julie Califf | Data Provider | NCDMF |
| Andrew Cathey | Data Provider | SC DNR |
| Amy Dukes | Data Provider | SC DNR |
| Eric Hiltz | Data Provider | FL FWCC |
| Dominique Lazarre | Data Provider | SC DNR |
| Kayla Rudnay | Data Provider | NC DMF |
| Beverly Sauls | Data Provider | Data Provider |

Appointees marked with an * were unable to attend the in-person workshop

### 1.5 Document List

| Document number | Title | Authors |
| :---: | :---: | :---: |
|  | Documents Prepared for SEDAR 60 |  |
| SEDAR60-WP01 | Red Porgy Fishery-Independent Index of Abundance in US South Atlantic Waters Based on a Chevron Trap Survey (1990-2017) | Bubley and Smart 2019 |
| SEDAR60-WP02 | Update of Red Porgy, Pagrus pagrus, Reproductive Life History from the MARMAP/SERFS program. | Wyanski et al. 2019 |
| SEDAR60-WP03 | Changes to NMFS age readings of U.S. South Atlantic Red Porgy (Pagrus pagrus) | Potts et al. 2018 |
| SEDAR60-WP04 | South Atlantic Red Porgy Commercial Hook-and-Line Discard Mortality Estimates Based on Observer Data | Pulver 2018 |
| SEDAR60-WP05 | Red Porgy Edge Analysis Memo | Bubley et al. 2018 |
| SEDAR60-WP06 | Commercial landings - Not Received |  |
| SEDAR60-WP07 | Standardized video counts of Southeast U.S. Atlantic red porgy (Pagrus pagrus) from the Southeast Reef Fish Survey | Cheshire and Bacheler 2018 |
| SEDAR60-WP08 | Red Porgy Length Frequency Distributions from At-Sea Headboat and Charter Observer Surveys in the South Atlantic, 2005 to 2017 | Lazarre et al. 2019 |
| SEDAR60-WP09 | Using Historical Data to Assign a Calendar Age to Red Porgy Otoliths without an Edge Type Assigned - revised May 3, 2019 | Bubley et al. 2019a |
|  | Reference Documents |  |
| SEDAR60-RD01 | 2012 SEDAR 1 South Atlantic Red Porgy Update Assessment Report | SEDAR 2012 |
| SEDAR60-RD02 | 2006 SEDAR 1 South Atlantic Red Porgy Update Assessment Report | SEDAR 2006 |
| SEDAR60-RD03 | SEDAR 1 Stock Assessment Report: South Atlantic Red Porgy | SEDAR 2002 |
| SEDAR60-RD04 | List of documents and working papers for SEDAR 1 (South Atlantic Red Porgy) - most documents available on the SEDAR website. | SEDAR 2002 |
| SEDAR60-RD05 | Southeast Reef Fish Survey Video Index Development Workshop | Bacheler Carmichael 2014 |
| SEDAR60-RD06 | Overview of sampling gears and standard protocols used by the Southeast Reef Fish Survey and its partners | Smart et al. 2014 |
| SEDAR60-RD07 | Technical documentation of the Beaufort Assessment Model (BAM) | Williams and Shertzer 2015 |
| SEDAR60-RD08 | Assessing barotrauma among angled red snapper (Pagrus auratus) and the utility of release methods | Butcher et al. 2012 |
| SEDAR60-RD09 | Survival estimates for demersal reef fishes released by anglers | Collins 1996 |


| Document number | Title | Authors |
| :---: | :---: | :---: |
| SEDAR60-RD10 | Age validation, movements and growth rates of tagged gag (Mycteroperca microlepis), Black Sea Bass (Centropristis striata) and Red Porgy (Pagrus pagrus) | Collins et al. 1996 |
| SEDAR60-RD11 | Excerpt from October 2017 SAFMC SSC Minutes (pages 37-66 where SEDAR activities were discussed) | SAFMC SSC |
| SEDAR60-RD12 | Commercial catch composition with discard and immediate release mortality proportions off the southeastern coast of the United States | Stephen and Harris 2010 |
| SEDAR60-RD13 | Release mortality of undersized fish from the snappergrouper complex off the North Carolina coast | Overton et al. 2008 |
| SEDAR60-RD14 | Discard composition and release fate in the snapper and grouper commercial hook-and-line fishery in North Carolina, USA | Rudershausen et al. 2007 |
| SEDAR60-RD15 | Snapper Grouper Advisory Panel Red Porgy Fishery Performance Report: April 2018 | SAMFC Snapper <br> Grouper AP  |
| SEDAR60-RD16 | Survival estimates for demersal reef fishes released by anglers | Collins 1996 |
| SEDAR60-RD17 | SEDAR1-DW6 - 2002 Tuesday Life History Group Discussion | SEDAR 01 Life History Group |
| SEDAR60-RD18 | 2016 MRIP Red Porgy Data Point Discussion Compiled by Mike Errigo | SSC MRIP Workshop August 2019 |
| SEDAR60-RD19 | Evaluating the Efficacy of Descender Devices in Increasing the Survival of Deepwater Groupers Using Telemetry | Brendan J. Runde and Jeffrey A. Buckel |

### 1.6 Statements Addressing Each Term of Reference

Note: Original ToRs are in normal font. Statements addressing ToRs are in italics and preceded by a dash ( - ).

1. Prepare a standard assessment, based on the approved 2012 SEDAR 1 South Atlantic Red Porgy Update assessment with data through 2017. Provide commercial and recreational landings and discards in pounds and numbers.

- This report documents the preparation of a standard assessment, based on the approved SEDAR 1, 2012 Update assessment (SEDAR 2012) with data through 2017. Observed time series of landings and discards are presented in Table 10, with associated CVs in Table 11. Estimated time series of landings are presented in numbers (Tables 25) and pounds (Table 26). Estimated time series of discards are presented in numbers (Tables 27) and pounds (Table 28).

2. Evaluate and document the following specific changes in input data or deviations from the update model. (List below each topic or new dataset that will be considered in this assessment.)

- Consider including the SERFS video index
- The SEDAR 60 panel agreed that the SERFS video index should be included in the current assessment by combining it with the SERFS chevron trap index to produce the SERFS chevron trap/video index. This index is included in the SEDAR 60 base model (Figure 13; Table 14).
- Incorporate the latest BAM model configurations and updates to data calculation methodologies, detailing the changes made and the impacts of those changes, between the 2012 SEDAR 1 South Atlantic Red Porgy Update assessment model and the proposed SEDAR 60 model.
- The latest BAM model configurations and updates to data calculation methodologies have been considered and included the the SEDAR 60 base model.
- Re-consider use of age and length composition data.
- Use of age and length composition data sets has been reconsidered. Age composition data for recent years have been added, and in most cases the same age composition data sets were used as in the SEDAR 1, 2012 Update, while most length composition data is no longer included in the model (Figure 4).

3. Document any changes or corrections made to the model and input datasets and provide updated input data tables. Fully document and describe the impacts (on population parameters and management benchmarks) of any changes to the model structure, methods, application or fitting procedures made between this assessment and the 2012 SEDAR 1 South Atlantic Red Porgy Update assessment.

- Changes made to the model and input datasets are documented throughout this report. Direct comparisons between the SEDAR 60 and SEDAR 12012 update models are described in in $\S 3.7$ and 4.11

4. Update model parameter estimates and their variances, model uncertainties, and estimates of stock status and management benchmarks. Compare population parameter trends and management benchmarks estimated in this assessment with values from the previous assessment, and comment on the impacts of changes in data, assumptions or assessment methods on estimated population conditions and benchmarks.

- Estimates of all model paramaters are presented in Appendix B. Estimates of stock status and management benchmarks are presented in Table 29. Direct comparisons between the SEDAR 60 and SEDAR 12012 update models are described in in $\S 3.7$ and 4.11

5. Provide stock projections, including a pdf (probability density function) for biological reference point estimates and yield separated for landings and discards reported in pounds and numbers. Projection results are required through 2024, with projected fishing level changes beginning in late 2019. The panel shall provide guidance on appropriate assumptions to address harvest and mortality levels in the interim years between the assessment terminal year (2017) and the first year of management (2019). Projection criteria:

- To determine OFL: (1) $\mathrm{P}^{*}=50 \%$; (2) $F_{\mathrm{MSY}}$
- To evaluate the existing rebuilding plan: base on fixed exploitation at $75 \% F_{\text {MSY }}$. In addition to reporting yield and stock status as described above, for this projection also report the probability that $\mathrm{SSB}>$ $\mathrm{SSB}_{\mathrm{MSY}}$.
- Projection results are described in §4.14. Relevant figures and tables are cited therein.

6. Review, evaluate, and report on the status and progress of all research recommendations listed in the last assessment, peer review reports, and SSC report concerning this stock.

- No research recommendations were made in the last assessment report (SEDAR 2012). I am not aware of research recommendations made in peer review reports or the SSC report associated with the last assessment.

7. Develop a stock assessment update report to address these TORS and fully document the input data, methods, and results of the stock assessment update.

- This SEDAR 60 Standard Assessment Report satisfies this ToR.


## 2 Data Review and Update

The benchmark assessment for Red Porgy, SEDAR-1, considered data from 1972-2001 (SEDAR 2002). An update to SEDAR-1 was completed in 2006 and considered data from 1972-2004 (SEDAR 2006). In another update, completed in 2012, the terminal year was extended to 2011 (SEDAR 2012). In the current SEDAR 60 assessment data up through 2017 were considered. For most data sources, the data were simply updated with the additional years of data (2012-2017) using the same methods as in the prior assessments. However, for some sources, it was necessary to update data prior to 2012 as well. The input data for this assessment are described below, with focus on the data that required modification beyond just the addition of years.

### 2.1 Data Review

In this standard assessment, the Beaufort assessment model (BAM) was fitted to many of the same data sources as in SEDAR-1 and the 2006 and 2012 updates.

- Landings: commercial handline, commercial trawl, headboat, general recreational (MRIP)
- Discards: commercial handline, headboat, general recreational (MRIP)
- Indices of abundance: SERFS chevron trap/video (formerly MARMAP chevron trap), headboat
- Length compositions of landings: commercial trawl
- Age compositions of surveys or landings: SERFS chevron trap (formerly MARMAP chevron trap), commercial handline, headboat

Contrasts to data used in the 2012 update assessment include:

- Commercial trap landings have now been combined with commercial handline landings
- The MARMAP Florida trap index and associated composition data are no longer included in the model.
- The MARMAP Chevron trap index was modified by the addition of video data since 2011 and is now known as the "SERFS chevron trap/video index"
- All sources of length composition data fitted to in the previous assessment have been excluded from the current assessment except for commercial trawl
- Commercial trawl length composition data has been pooled into a single year, due to small sample sizes.
- The limited data for age-0 fish were not included, as the current model starts at age-1
- Ages are now in calendar age compared with increment ages used in previous Red Porgy assessments

In addition to data fitted by the model, prior assessments utilized life-history information that was treated as input, much of which has been updated in SEDAR 60. The same length-weight and male maturity at age relationships used in the 2012 update were used here. Estimates of proportion male at age were updated for SEDAR 60 with more recent data. Whereas prior assessments of Red Porgy included time blocks for female maturity at age, in SEDAR 60 a single time-invariant vector of female maturity at age was developed including the most recent data. Also new to SEDAR 60, age-varying estimates of natural mortality rate have been developed included, whereas prior assessments treated it as constant across ages. Discard mortality rates have also been reevaluated and modified for this assessment. In SEDAR 1, 2012 Update, empirical estimates of growth model parameters were used as starting values in the assessment model, and the parameters were estimated in the assessment. By constrast, in SEDAR 60, updated estimates of growth model parameters were fixed in the base model.

### 2.2 Data Update

### 2.2.1 Life History

All of the life history inputs have been updated since the SEDAR 1, 2012 Update assessment with additional data from 2011-2017 (SEDAR 2012) and include changes to age data for the full time series of the assessment, reproductive parameters and natural mortality. The primary change was to the age data. NMFS Beaufort Laboratory staff conducted an age validation study, refining the methodology for ageing Red Porgy, specifically regarding first annulus. Since SEDAR 1 (2002), all age data were recorded as increment, or annuli count, with no additional adjustment. Age samples processed and read since the SEDAR 1, 2012 Update have edge types included. As a result of the age validation study, NMFS Beaufort re-read all their old samples and included edge types (Potts et al. 2018). Thus, NMFS Beaufort was able to assign calendar ages to all samples. SCDNR did not have funding and time to re-age their historic samples, originally using whole otoliths, for this SEDAR, but in a comparison of 265 samples that were sectioned, read and compared to whole otolith readings, SCDNR determined that the original ages were consistent with the ages from the sections. It was agreed that SCDNR did not need to re-age the historic samples at this time, but would in the future, provided available funding and time. One issue with those samples was the readings did not include edge types. SCDNR provided a proxy method to assign edge types to samples and then calculated calendar age using this method (Bubley and Smart 2019). The SEDAR 60 panel accepted this method as an appropriate way to convert increment counts to calendar age, and as a result the current age data set for SEDAR 60 is based on calendar age. Updated analyses of life history parameters that rely on the new age data were completed. A population growth model, reproductive parameters and natural mortality were all updated.

Changes to the population growth model used in the SEDAR 1, 2012 Update include the use of fractional, or biological, age and inclusion of the correction for the minimum size-limit bias on the size-at-age distribution of the fishery-dependent samples following McGarvey and Fowler (2002). The data for the population growth model included 42,434 samples, spanning years from 1979 through 2017, collected by MARMAP and the SERFS and from the commercial and recreational fisheries. The calendar age for each fish was converted to a fractional age based on the month of capture and month of peak spawning (February), using the following formula:

$$
\begin{equation*}
A_{f}=A_{c}+\left[\left(M_{c}-M_{s}\right) / 12\right] \tag{1}
\end{equation*}
$$

where, $A_{f}=$ Fractional, or biological, age; $A_{c}=$ Calendar, or cohort, age; $M_{c}=$ month of capture, and $M_{s}=$ month of peak spawning.

A minimum size limit was assigned to each sample from the fishery based on the management history. The parameter values for the von Bertalanffy growth model ( $\pm$ standard error) are $L_{\infty}=422.6$ ( 1.25 ; TL, mm), $k=0.30$ ( 0.004 ), and $t_{0}=-1.47$ (0.036, years; Figure 1).

The SEDAR 60 Data/Assessment panel had a robust discussion of natural mortality (M), which included maximum age of the population, single point estimate of $M$ and equations, age-varying $M$ and whether to scale the age-varying $M$ to the point estimate. With the updated age data for this assessment, the max age of Red Porgy in the US South Atlantic has increased from 19 years to 25 years. Consideration was given to what max age we should use for calculation of a point estimate of M . The staff engaged in aging Red Porgy have confidence in the oldest ages because of an age validation study and the repeatability of reading of the samples from the old fish across time and multiple readers. Though only 12 fish in the entire age data set were age- 20 or older, they occur in every fishery sector and the fishery-independent survey. Those fish were collected in the most recent years, 2011-2017 and have birth years of 1988-1997 with 5 of 12 in 1993. The fish were from a time of heavy exploitation and the maximum observed age may actually be an underestimate of the true max age. A recommendation was made to use a range
of max age of 20 to 30 years with the mid-point of the range for the base run of the assessment model. The panel had most confidence in using point estimates of $M$ calculated from equations using max age rather than those using von Bertalanffy growth parameters, because of the confidence of the experts assigning the ages to each sample. The panel considered estimates generated from Hoenig's (1983) original equation, as used in SEDAR01, and Then et al. (2014). During discussion of what equation to use for the point estimate of M, the estimates calculated from the Then et al. (2014) age based equation ( 0.26 , max age $=25$ ) were considerably higher than those using the original Hoenig (1983) equation ( 0.17 , max age $=25$ ). Table 1 shows the range of $M$ for ages 20,25 and 30 . Annual total mortality ( Z ) from the age composition data were calculated for comparison to the point estimates of M . In some years, Z was less than $M$ of 0.32 as calculated from Then et al. (2014) for max age of 20 . The panel recommended using $M=0.22$, the average of the values in Table 1, as the point estimate for the base run of the assessment model.

To be consistent with recent stock assessments, the panel proposed to use an age-varying $M$ based on Charnov et al. (2013), but whether to scale the estimates to the point estimate based on the fully recruited ages was discussed by the panel. An age-varying $M$ has the advantage of recognizing that the smallest fish are subjected to a higher rate of natural mortality than larger fish. The age specific estimates of $M$ saturated around 0.33 at age- 14 , but is still higher than Z for some years. For this reason, the panel recommended to scale the age specific estimates of $M$ to the point estimate using age-3+ as the fully recruited ages (Table 2). The fully recruited age was determined from the age composition from the fishery-independent chevron trap data set. The scaling provided equivalent cumulative survival across ages as would be achieved with the age-invariant point estimate of $M$.

Reproductive biology parameters were updated with additional data collected between 2012 and 2016, which included female and male maturity at age with the new calendar ages. Some members of the panel expressed concern at the SEDAR 60 Red Porgy in-person Workshop (December 2019) about using period-specific estimates of life history parameters in the model for only female maturity, as was done in the 2006 and 2012 assessments. It is possible that other parameters such as growth and sex ratio also exhibit the plasticity seen in female maturity, which could affect these maturity estimates. In addition, a panel member asked if the period-specific maturity ogives are statistically different.

To address the question of statistical significance, female maturity data from MARMAP and SERFS sampling were grouped into three periods (1979-1987, 1990-2002, and 2003-2016), with the latter two periods representing data from chevron traps. Maturity ogives for the three periods were compared using a Probit analysis with the logistic distribution function. The results showed that the proportion of mature females at calendar age decreased significantly $(P<0.001)$ between the early and middle periods and then increased significantly $(P<0.001)$ between the middle and latter periods, with the differences in maturity ogives for the early and latter periods not being statistically significant ( $P=0.067$; Table 7 in addendum of Wyanski et al. 2019).

Although there is statistical evidence for the use of period-specific maturity ogives, the consensus of the workshop panel was to shift to an overall (1979-2016) maturity ogive in the model until temporal trends in other life history parameters can be investigated. Parameter estimates for the overall ogive are presented in the updates of Tables 4 and 5 (addendum of Wyanski et al. 2019).

### 2.2.2 Life History Tables and Figures

Table 1. Point estimates of natural mortality ( $M$ ) based on equations using maximum age in the population from Then et al. (2014) and Hoenig (1983) for South Atlantic Red Porgy.

| Maximum Age (years) | Then et al. (2014) | Hoenig (1983) |
| :---: | :---: | :---: |
| 20 | 0.32 | 0.21 |
| 25 | 0.26 | 0.17 |
| 30 | 0.22 | 0.14 |

Table 2. Age specific estimates of natural mortality (M) calculated from the Charnov et al. (2013) equation and scaled to the cumulative survival of fish ages $3+$ from the recommended point estimate of $M=0.22$.

| Age | M |
| :---: | :---: |
| 1 | 0.455 |
| 2 | 0.355 |
| 3 | 0.302 |
| 4 | 0.271 |
| 5 | 0.251 |
| 6 | 0.238 |
| 7 | 0.229 |
| 8 | 0.223 |
| 9 | 0.218 |
| 10 | 0.215 |
| 11 | 0.212 |
| 12 | 0.211 |
| 13 | 0.209 |
| 14 | 0.209 |

Figure 1. Population growth model of Red Porgy ( $n=42,434 ; T L \mathrm{~mm}$ ) including size-limit correction on fisherydependent samples.


### 2.2.3 Commercial Landings and Discards

Red Porgy commercial landings were compiled for years 1972-2017 with a U.S. Atlantic Coast stock boundary from the Virginia southern border down through to the southern tip of Florida: North Carolina, South Carolina, Georgia, east coast of Florida. Landings totals from 1972-2001 were left unchanged from SEDAR01. Years 2002-2017 were updated. Combined annual landings are provided in Table 3.

Direct sources for the landings included the Atlantic Coastal Cooperative Statistics Program (ACCSP), the North Carolina Division of Marine Fisheries (NCDMF), and the South Carolina Department of Natural Resources (SCDNR).

Statistics on commercial landings (1972 to present) for all species on the Atlantic coast are maintained in the ACCSP Data Warehouse. The Data Warehouse is an online database of fisheries dependent data provided by the ACCSP state and federal partners. The Data Warehouse was queried for all Red Porgy landings (annual summaries by gear category) from 1972 - 2017 from North Carolina through Florida (ACCSP 2019). Commercial landings in pounds (whole weights using state specific conversion factors) were provided. All landings were then aggregated into year and state summaries. Georgia (GADNR) and Florida Fish and Wildlife (FWC) staff examined ACCSP landings and compared them to state held versions. It was determined that ACCSP landings were a match and would be used in place of state provided data for the entire time series.

The NCDMF provided North Carolina's landings data from 1972 - 2017. This data set was a collective grouping of historical data collection by the NMFS/NCDMF Cooperative Statistics Program, its predecessors, and the NC Trip Ticket Program. Data continuity and accuracy dramatically increased over time. From $1994-2017$ landings data collection was provided by the NC Trip Ticket Program and considered the most consistent and inclusive portion of the NC dataset. Landings were reported in both whole and gutted conditions. The landings reported in gutted weight, were converted to whole with a state conversion factor of 1.25 per pound. Whole weight records were directly supplied without conversion.

In 1972, South Carolina began collecting landings data from coastal dealers in cooperation with federal agents. Mandatory monthly landings reports on forms supplied by the Department are required from all licensed wholesale dealers in South Carolina. Until fall of 2003, those monthly reports were summaries collecting species, pounds landed, disposition (gutted or whole) and market category, gear type and area fished; since September 2003, landings have been reported by a mandatory trip ticket system collecting landings by species, disposition and market category, pounds landed, ex-vessel prices with associated effort data to include gear type and amount, time fished, area fished, vessel and fisherman information.

SCDNR provided landings data for Red Porgy from 1972 - 2017. Data from 1978 - 2003 were collected in monthly totals through collaborative efforts by SCDNR and the NMFS Cooperative Statistics Program and data collected from 2004-2017 were more comprehensive, as SCDNR instituted a mandatory Trip Ticket Program in late 2003. Landed weights were collected as both gutted and whole. Annual Catch Limits are categorized as "landed weight" since both categories are present in the fishery. All gutted weights were converted to whole weight using the state conversion factor.

## Commercial gear

Initial commercial fleets for Red Porgy consisted of trap, trawl, and handline, which included hook and line, diving, spear, other, and unreported gears. Upon reviewing gear selectivity and the small scale of the trap landings, the trap fleet was combined with the handline fleet. The trawl fleet was determined to generally use gear that targets smaller fish and therefore was kept as a separate fleet.

## Commercial discards

Two approaches were investigated for the calculation of Red Porgy discards from the commercial handline fishery. Both methods calculate total discards as discard rate*total effort of the fishery. The first technique (continuity method) followed the methods used in the SEDAR 1, 2006 Updateassessment by modeling discard rates. The second technique followed the methods recommended in SEDAR 32 and subsequent South Atlantic SEDAR assessments (standard method) where discard rates were directly calculated from discard logbook data. Total effort data were available from commercial logbook data for both methods. Although the results of both analyses were available for the assessment, the standard method was recommended as the preferred method for commercial discard calculation.

Red Porgy discard calculation used data reported by fishers between January 1, 2002 and December 31, 2017 in the US South Atlantic (south of US1 in the Florida Keys to $37^{\circ} \mathrm{N}$ ) from vertical line (handline and electric/hydraulic gears) trips. Approximately $98 \%$ of reported Red Porgy discards were from vessels fishing vertical line gear. Data filtering followed the methods recommended during SEDARs 32 and 41 (McCarthy 2013; 2015). Effort data were also filtered to exclude trips landing only mackerel because the SEDAR 32 and 41 panels noted that for trips targeting mackerel only, the likelihood of catching species other than mackerel was extremely low. To avoid removing mixed effort trips, however, only trips with $100 \%$ mackerel landings were excluded from the analysis.

A final data filter designed to address possible underreporting of commercial discards was included following the recommendation of the SEDARs 32 and 41 commercial work groups. The percentage of discard reports returned with "no discards" from vertical line trips has increased from 33 to 73 percent in the US South Atlantic over the period 2002 - 2017. The data were filtered to remove records from vessels that never reported discards of any species during a year. Following the SEDAR 32 and 41 commercial working groups' recommendations, data from vessels that reported many more trips than the fleet average before a discard was reported (the mean number of trips prior to the first trip with reported discards plus two standard deviations above that mean) were excluded.

Yearly discard rates of vertical line vessels were calculated as the mean rate (discards per hook hour fished) during the years 2002 - 2017. Discard rates were calculated separately for open and closed Red Porgy seasons. Yearly total effort (hook hours, available from commercial logbook reports) of all trips, by season (open/closed), was multiplied by the yearly season specific mean discard rate to calculate total discards of Red Porgy by vertical line vessels:

Calculated discards per region $=$ yearly mean Red Porgy discard rate per season $\times$ total effort per season
where total effort per season indicates total effort post filtering.
For years prior to 2002 (the first year of discard data), the mean discard rate, by season, for the years 2002 - 2006 was used to calculate discards for the years 1993 - 2001 when only effort data were available.

$$
\begin{equation*}
\text { Calculated discards per region }=02-06 \text { mean Red Porgy discard rate per season } \times \text { total effort per season } \tag{3}
\end{equation*}
$$

where total effort per season indicates total effort post filtering.
Total discards are provided in Table 4 for combined seasons in number ( $1,000 \mathrm{~s}$ ) of fish. The very high number of estimated discards in 2002 was due to a number of trips with much higher discard rates than those reported from trips in other years (Table 5 and Figure 2). Year 2002 was determined to be the best data available and was not adjusted or weighted differently.

The discard calculations rely on self-reported discard and effort data. Perhaps the most important source of error in the commercial discard calculations was misreporting and nonreporting of discards, both of Red Porgy and other
species. An effort was made to minimize that potential error by removing data from vessels that never reported discards of any species during a year or reported many more trips than the fleet average before a discard was reported. Although such clear instances of discard non-reporting were identified and excluded, other cases of non-reporting and misreporting have not been quantified. The degree to which continued non or misreporting may have affected the discard calculations is unknown. The discard totals provided may represent a minimum estimate of the number of Red Porgy discarded from the commercial vertical line fishery.

### 2.2.4 Commercial Landings and Discards Tables and Figures

Table 3. Annual commercial landing totals of Red Porgy from all fisheries reported in 1000 pounds and metric tons.

| Year | Landings (lb) | Landings (mt) |
| :---: | :---: | :---: |
| 1972 | 32.84 | 14.90 |
| 1973 | 27.60 | 12.52 |
| 1974 | 108.35 | 49.14 |
| 1975 | 198.90 | 90.22 |
| 1976 | 250.96 | 113.83 |
| 1977 | 437.10 | 198.26 |
| 1978 | 726.39 | 329.48 |
| 1979 | 1066.67 | 483.83 |
| 1980 | 1233.80 | 559.64 |
| 1981 | 1571.19 | 712.68 |
| 1982 | 1606.09 | 728.51 |
| 1983 | 1295.81 | 587.77 |
| 1984 | 1124.96 | 510.27 |
| 1985 | 863.58 | 391.71 |
| 1986 | 921.25 | 417.87 |
| 1987 | 787.13 | 357.03 |
| 1988 | 893.05 | 405.08 |
| 1989 | 924.36 | 419.28 |
| 1990 | 1138.59 | 516.45 |
| 1991 | 832.44 | 377.59 |
| 1992 | 516.53 | 234.30 |
| 1993 | 470.08 | 213.22 |
| 1994 | 436.36 | 197.93 |
| 1995 | 432.07 | 195.98 |
| 1996 | 429.61 | 194.87 |
| 1997 | 425.70 | 193.09 |
| 1998 | 317.99 | 144.24 |
| 1999 | 105.14 | 47.69 |
| 2000 | 26.21 | 11.89 |
| 2001 | 66.17 | 30.02 |
| 2002 | 58.17 | 26.39 |
| 2003 | 50.37 | 22.85 |
| 2004 | 49.68 | 22.54 |
| 2005 | 48.66 | 22.07 |
| 2006 | 83.81 | 38.02 |
| 2007 | 144.29 | 65.45 |
| 2008 | 171.96 | 78.00 |
| 2009 | 164.53 | 74.63 |
| 2010 | 158.83 | 72.04 |
| 2011 | 202.83 | 92.00 |
| 2012 | 162.26 | 73.60 |
| 2013 | 171.46 | 77.77 |
| 2014 | 158.15 | 71.74 |
| 2015 | 154.82 | 70.22 |
| 2016 | 127.44 | 57.81 |
| 2017 | 129.81 | 58.88 |
|  |  |  |
|  |  |  |

Table 4. Yearly calculated total discards of Red Porgy from vertical line vessels, seasons (open/closed) combined, using SEDAR 32 methods. Discards are reported as number of fish in 1,000s.

| Year | Total discards (1000 fish) |
| :---: | :---: |
| 1993 | 78.26 |
| 1994 | 96.75 |
| 1995 | 101.04 |
| 1996 | 100.07 |
| 1997 | 102.78 |
| 1998 | 78.52 |
| 1999 | 79.91 |
| 2000 | 87.67 |
| 2001 | 81.61 |
| 2002 | 250.94 |
| 2003 | 45.92 |
| 2004 | 39.10 |
| 2005 | 25.03 |
| 2006 | 40.23 |
| 2007 | 25.33 |
| 2008 | 40.18 |
| 2009 | 33.59 |
| 2010 | 21.02 |
| 2011 | 11.94 |
| 2012 | 27.83 |
| 2013 | 26.37 |
| 2014 | 28.11 |
| 2015 | 30.37 |
| 2016 | 16.77 |
| 2017 | 18.43 |

Table 5. Comparison of Red Porgy commercial discard rates (DR; number of fish discarded per hook hour fished) by year (2002-2017). The discard rate mean and standard deviation (SD) were calculated for each year. The bound (B) of the mean was calculated as two times the SD. A high rate for each year was calculated by adding the $B$ and the mean discard rate. The number of trips with discard rates greater than the yearly high rate are provided in the $N G_{\text {yearly }}$ column ( $N G=$ Number Greater). The number of trips with discard rates greater than the total mean high rate (0.638) are provided in the $N G_{\text {mean }}$ column. All trips were used for the calculations including trips with a discard rate of zero.

| Year | Mean Discard Rate | SD Discard Rate | $\mathrm{B}\left(2^{*} \mathrm{SD}\right)$ | High Rate (B+Mean) | $N G_{\text {yearly }}$ | $N G_{\text {mean }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 0.208 | 1.447 | 2.894 | 3.102 | 17 | 58 |
| 2003 | 0.043 | 0.361 | 0.721 | 0.764 | 22 | 25 |
| 2004 | 0.044 | 0.291 | 0.582 | 0.626 | 21 | 21 |
| 2005 | 0.030 | 0.142 | 0.285 | 0.315 | 35 | 11 |
| 2006 | 0.041 | 0.185 | 0.369 | 0.410 | 21 | 12 |
| 2007 | 0.030 | 0.233 | 0.466 | 0.497 | 30 | 22 |
| 2008 | 0.046 | 0.464 | 0.928 | 0.973 | 21 | 54 |
| 2009 | 0.036 | 0.171 | 0.341 | 0.377 | 61 | 27 |
| 2010 | 0.027 | 0.185 | 0.369 | 0.396 | 54 | 31 |
| 2011 | 0.016 | 0.123 | 0.246 | 0.262 | 62 | 11 |
| 2012 | 0.042 | 0.282 | 0.564 | 0.607 | 51 | 46 |
| 2013 | 0.036 | 0.188 | 0.377 | 0.413 | 92 | 53 |
| 2014 | 0.036 | 0.179 | 0.359 | 0.395 | 116 | 58 |
| 2015 | 0.042 | 0.222 | 0.445 | 0.487 | 82 | 64 |
| 2016 | 0.023 | 0.137 | 0.274 | 0.297 | 83 | 27 |
| 2017 | 0.028 | 0.133 | 0.266 | 0.295 | 72 | 32 |
| Total Mean | 0.046 | 0.296 | 0.593 | 0.638 | 53 | 35 |

Figure 2. Frequency plot of Red Porgy commercial discard rates (number of fish discarded per hook hour fished) by year $(2002-2017)$ with calculated high rate (3.102, as defined in Table 5) for the year 2002 (blue dashed line). All trips with a discard rate of zero were removed for these plots. Panels have different $x$-axis and $y$-axis ranges.


### 2.2.5 Recreational Fisheries

The primary recreational modes of fishing for South Atlantic Red Porgy are private, charter, and headboat. Estimates of the catch of Red Porgy come from a combination of results from two surveys: (1) the Marine Recreational Information Program (MRIP), formerly the Marine Recreational Fisheries Statistics Survey (MRFSS), conducted by NMFS; and (2) the Southeast Region Headboat Survey (SRHS) conducted by NMFS, Southeast Fisheries Science Center Beaufort Laboratory in North Carolina. The MRIP survey is sampling-based, whereas the SRHS is a census of headboats using logbooks. The two surveys together provide estimates of catch in numbers, estimates of effort, length and weight samples, and catch-effort observations for recreational fishing.

## MRIP transition

The Marine Recreational Information Program completed a three year transition in 2018 (NOAA Fisheries 2018). Estimates of fishing effort for the private and shore modes are now obtained from a Fishing Effort Survey conducted via mail, which uses angler license and registration information to identify and contact anglers as well as supplemental data from the U.S. Postal Service that includes nearly all U.S. households. Effort estimates for charter and party boats are still obtained from the For-Hire Telephone Survey and are not affected by the new Fishing Effort Survey. Previously, estimates of private and shore fishing effort came from the legacy Coastal Household Telephone Survey, which used random-digit dialing of homes in coastal counties to contact anglers. Concerns over low response rates, the gatekeeper effect (i.e., speaking to someone other than the angler), the tendency to ignore unknown callers, and coverage limited to only coastal counties in the Coastal Household Telephone Survey were motivation for the new survey, which is considered to provide more accurate estimates of trips. By design, the Fishing Effort Survey is reaching more anglers, getting into the right hands, providing a higher response rate, and extracting more information from anglers with an improved survey questionnaire. Benchmarking of the Fishing Effort Survey alongside the Coastal Household Telephone Survey for three years allowed for apples-to-apples comparisons between data from the two different surveys and the creation of a peer-reviewed calibration model. The calibration model was peer reviewed by reviewers appointed by the Center for Independent Experts (see Rago et al. 2017). Additional details can be found at: https://www.fisheries.noaa.gov/event/fishing-effort-survey-calibration-model-peer-review.

The MRIP transition also accounted for the 2013 design change in the Access Point Angler Intercept Survey (Foster et al. 2018). Improved survey procedures were incorporated that better account for all types of completed trips and remove potential sources of bias from the survey design. For example, the new sampling design provides more complete coverage of angler fishing trips ending throughout the day and night, whereas the old design often missed nighttime trips or off-peak daytime trips. In addition, conversion factors were developed to account for any consistent effects of the redesign on catch rate estimates produced by the Access Point Angler Intercept Survey. The new Access Point Angler Intercept Survey design uses a sample weight adjustment method and is more statistically sound because it more strictly adheres to formal probability sampling protocols. The Access Point Angler Intercept Survey calibration model developed by MRIP and the statistical approach proposed for the conversion of catch estimates by MRIP were peer reviewed by reviewers appointed by the Center for Independent Experts. Additional details can be found at: https://www.fisheries.noaa.gov/event/access-point-angler-intercept-survey-calibration-workshop.

## Charter calibration

The MRIP transition resulted in the release of new recreational catch estimates for all species and all modes, including charter mode estimates. As a result, the SEFSC conducted a calibration analysis using the newly released data to correct for this change from the Coastal Household Telephone Survey to the For-Hire Telephone Survey (Dettloff and Matter 2019). The analysis uses a statistically sound, consistent methodology to provide improved calibrations for estimating For-Hire Telephone Survey charterboat effort and landings with associated uncertainties from Coastal Household Telephone Survey estimates. Additional details are provided in Dettloff and Matter (2019).

## Recreational Fisheries

Recreational landings in number were aggregated into two separate recreational fleets, headboat and general recreational (charterboat and private boat) and used in the SEDAR 60 assessment model. The headboat (1981-2017) and general recreational (1981-2017) landings and discards were updated based on data from the SRHS and from MRIP. Recreational discards in numbers of Red Porgy by the headboat, charter, and private modes were used in the assessment model. MRIP estimates of live released fish (B2) for charter, private, and headboat (1981 - 1985 only) were adjusted in the same manner as landings (i.e., discussed above) and did include Monroe County. Self-reported discards have been reported in the SRHS logbook since 2004 and were validated using the At-Sea Observer Program. As a result, headboat discards from 2004-present were derived directly from the SRHS. Headboat discards were recalculated for the entire time series, as it is a model-based approach. The accepted SEDAR Best Practice method MRIP Charter:SRHS discard ratio was recommended as a proxy to estimate Red Porgy discards from headboats for years prior to 2004 (Fisheries Ecosytems Branch, National Marine Fisheries Service, Southeast Fisheries Science Center, Beaufort, NC 2017). Recreational landings and discards, as provided, are shown in Tables 6 and 7, respectively.

### 2.2.6 Recreational Fisheries Tables

Table 6. Red Porgy landings in numbers ( $n$ ) and pounds (lb) from the recreational fishery (1981-2017).

| Year | Headboat (n) | Charter boat (n) | Private boat (n) | Headboat (lb) | Charter boat (lb) | Private boat (lb) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 168,286 | 45,223 | 10,951 | 325,458 | 63,116 | 15,718 |
| 1982 | 272,883 | 46,577 | 8,319 | 431,938 | 72,166 | 13,294 |
| 1983 | 155,738 | 18,805 | 2,477 | 261,450 | 33,223 | 4,277 |
| 1984 | 129,970 | 252,002 | 6,262 | 217,036 | 344,189 | 8,773 |
| 1985 | 176,576 | 12,955 | 182,719 | 260,381 | 16,844 | 247,131 |
| 1986 | 161,041 | 16,874 | 19,429 | 222,088 | 25,630 | 29,737 |
| 1987 | 173,568 | 43,912 | 26,689 | 220,476 | 61,827 | 41,170 |
| 1988 | 168,556 | 27,736 | 180,805 | 215,534 | 39,527 | 219,920 |
| 1989 | 146,488 | 103,168 | 35,326 | 165,050 | 172,263 | 46,149 |
| 1990 | 104,762 | 51,150 | 56,275 | 125,265 | 85,643 | 71,210 |
| 1991 | 129,879 | 19,251 | 41,281 | 140,820 | 37,506 | 59,328 |
| 1992 | 85,893 | 54,466 | 104,152 | 109,858 | 90,178 | 120,823 |
| 1993 | 81,695 | 36,656 | 17,925 | 101,027 | 50,286 | 24,889 |
| 1994 | 70,390 | 30,206 | 40,111 | 87,572 | 43,566 | 50,320 |
| 1995 | 70,713 | 39,061 | 5,880 | 93,032 | 54,357 | 8,186 |
| 1996 | 64,907 | 23,117 | 42,886 | 82,218 | 33,122 | 64,357 |
| 1997 | 53,865 | 12,536 | 8,211 | 75,298 | 17,123 | 13,189 |
| 1998 | 53,878 | 22,011 | 9,089 | 69,262 | 22,354 | 10,242 |
| 1999 | 31,954 | 14,939 | 13,444 | 48,657 | 21,241 | 20,537 |
| 2000 | 8,036 | 1,041 | 9,516 | 13,906 | 2,095 | 19,149 |
| 2001 | 28,862 | 29,483 | 10,887 | 46,308 | 56,791 | 20,566 |
| 2002 | 20,925 | 50,742 | 3,873 | 33,341 | 127,190 | 15,896 |
| 2003 | 20,174 | 30,273 | 24,491 | 34,743 | 66,532 | 57,078 |
| 2004 | 23,461 | 24,962 | 10,484 | 49,575 | 49,309 | 49,056 |

Table 7. Red Porgy discards in numbers ( $n$ ) from the recreational fishery $(1981-2017)$ released alive or dead.

| Year | Headboat (alive) | Headboat (dead) | Charter boat (alive) | Private boat (alive) |
| :---: | :---: | :---: | :---: | :---: |
| 1981 | 2,823 |  | 1,263 |  |
| 1982 | 1,593 |  | 487 | 1,483 |
| 1983 | 3,621 |  | 705 |  |
| 1984 | 1,511 |  | 1,492 |  |
| 1985 | 4,542 |  | 541 | 14,520 |
| 1986 |  |  |  | 547 |
| 1987 |  |  |  | 22,635 |
| 1988 |  |  |  | 671 |
| 1989 | 632 |  | 105 | 20,966 |
| 1990 |  |  |  |  |
| 1991 | 825 |  | 203 |  |
| 1992 | 8,594 |  | 7,737 | 5,288 |
| 1993 | 11,600 |  | 8,158 |  |
| 1994 | 2,038 |  | 1,542 | 1,042 |
| 1995 | 17,501 |  | 12,748 | 9,452 |
| 1996 | 2,144 |  | 946 | 6,067 |
| 1997 |  |  |  | 1,759 |
| 1998 |  |  |  | 6,500 |
| 1999 | 81,463 |  | 41,167 | 19,474 |
| 2000 | 32,147 |  | 6,422 | 16,201 |
| 2001 | 56,794 |  | 93,330 | 24,397 |
| 2002 | 9,291 |  | 53,758 | 8,153 |
| 2003 | 10,235 |  | 27,629 | 69,304 |
| 2004 | 61,341 | 2,397 | 53,893 | 37,882 |
| 2005 | 18,216 | 560 | 27,862 | 23,630 |
| 2006 | 42,338 | 957 | 5,718 | 2,494 |
| 2007 | 42,069 | 503 | 19,956 | 32,131 |
| 2008 | 26,784 | 1,038 | 20,534 | 92,078 |
| 2009 | 14,531 | 237 | 1,694 | 9,326 |
| 2010 | 12,827 | 93 | 2,326 | 20,891 |
| 2011 | 14,795 | 299 | 2,678 | 19,265 |
| 2012 | 16,488 | 417 | 5,693 | 3,036 |
| 2013 | 13,908 |  | 3,579 | 10,859 |
| 2014 | 17,844 |  | 3,942 | 31,638 |
| 2015 | 18,782 |  | 36,203 | 30,930 |
| 2016 | 15,457 |  | 1,023 | 272,938 |
| 2017 | 11,202 |  | 1,492 | 34,376 |

### 2.2.7 Indices of Abundance

The 2012 SEDAR-01 update assessment of Red Porgy included three indices of abundance: one derived from the headboat fleet (1973-1998), one from MARMAP sampling with chevron traps (1990-2011), and one from MARMAP sampling with Florida traps ( 1983 - 1987). The headboat index was standardized using a delta-GLM approach. Neither MARMAP index was standardized, however a sensitivity run of the 2012 assessment model was conducted using a standardized version of the chevron trap index.

For this SEDAR-60 assessment, the headboat index was left intact at values from the previous 2012 assessment. The index was not reevaluated with more recent years for reasons stated in the 2012 update report, primarily that harvest regulations since 1999 have likely compromised fishery dependent catch per unit effort as a meaningful measure of abundance. Data from the chevron trap survey were updated through 2017 and include sampling from SERFS (MARMAP, SEAMAP-SA, and SEFIS). The chevron trap index was standardized using a zero-inflated negative binomial model (Bubley and Smart 2019). Additionally, video sampling from SERFS was included, spanning 2011 - 2017. The video index was also standardized using a zero-inflated negative binomial model (Cheshire and Bacheler 2018). The two indices from SERFS gears (chevron traps and video) were combined using the method of Conn (2010), as has been done in several recent SEDAR assessments. The MARMAP Florida trap index was excluded from this assessment, for the following reasons: 1) the index was not standardized but rather was simple nominal catch per effort, as developed during SEDAR-01, 2) the 5 -year time-series was relatively short, 3) geographic coverage, depth coverage, and sampling intensity were less extensive than for other indices, 4) the index occurred during a time period that already contained what is believed to be a reliable index (headboat), and 5) the index was not informative for the assessment model (as indicated by model runs with and without the index).

### 2.2.8 Length Compositions

Length compositions in total length (TL) for all data sources were developed in 1-cm bins over the entire size range. These were later pooled at the tails to a range $12-72 \mathrm{~cm}$ (labeled at bin center). All fishery-dependent length compositions were weighted by regional landings defined by sample size for each fleet. A 30 fish minimum sample size for each region and year was used to prevent spikes in the compositions when length sampling was disproportionally small relative to landings.

The commercial handline lengths were weighted by the regional landings. The regions were defined by sample size as NC, SC, and combined GA and FL. For many state and year combinations, the commercial lengths were collected in centimeters fork length. Problems with missing or heaped bins can result from applying a conversion developed to data collected on a different scale ( mm vs. cm ). A random tenth decimal was added to these values prior to conversion to total length. Some states collected lengths in half or quarter centimeter bins in early years. For these values a similar approach was used to distribute lengths in half-centimeter increments to the adjacent bins. The commercial trap length data was minimal. Commercial trap lengths were very similar to commercial handline for the few years with adequate trap samples for comparison. The trap lengths were removed from the model based on the decision to combine handline and trap fleets. The few annual commercial trawl length compositions used in previous Red Porgy assessments were discovered to be from very few trips each year and were pooled across years.

Headboat and general recreational length compositions were developed and weighted regionally. However, the general recreational lengths were not retained due to minimal sample sizes. The headboat lengths were weighted by the regional landings at the same spatial strata as commercial data.

Red Porgy length compositions were provided for the chevron trap time series 1990-2017. All Red Porgy collected in chevron traps for monitoring purposes were enumerated and measured to produce length compositions, with FL
measurements converted to TL based on a meristic conversion as needed (Bubley et al. (2019b); TL=$\frac{(F L+3.4449)}{0.8744}$, $\left.r^{2}=0.997, n=25,789\right)$. There were two time periods regarding length measurements, with measurements being in fork length (FL) to the whole centimeter from 1990 to 2011 and maximum total length (TL) from 2012 to present. The length compositions from fishery-independent chevron traps were developed and discussed at the data workshop. An issue with the compositions was discussed but no action was needed because they were not recommended for use based on other criteria. The length compositions from the Florida snapper trap used in the previous assessments were no longer needed based on the decision to exclude the Florida snapper trap index.

Including both length and age compositions from the same fleet can result in overweighting of composition data, and recent SEDAR assessments have removed length composition data when sufficient age composition data are available (e.g., SEDAR 41, SEDAR 55, SEDAR 56, and SEDAR 58). Age compositions were not fit well when length compositions were included. The SEDAR 60 panel recommended excluding all length composition data with the exception of commercial trawl where no ages were available. This pooled composition was recommended to inform trawl selectivity only.

### 2.2.9 Age Compositions

Fishery-dependent age compositions were weighted by the region- and fleet-specific length compositions to address potential disproportionate sampling among regions and bias in selection of fish to be aged (see Sustainable Fisheries Branch - NMFS 2017; Fisheries Ecosytems Branch, National Marine Fisheries Service, Southeast Fisheries Science Center, Beaufort, NC 2017, for methods). Annual region-fleet sampling with fewer than 10 fish were excluded to limit problems with up-weighting small samples.

Red Porgy age compositions were provided for the chevron trap time series 1990-2017. Red Porgy age compositions had to account for differing sub-sampling routines and length measurements during the survey. There were two time periods for life history subsampling routines, with 1990 to 2007 having nonrandom subsampling based on tallies within length bins and 2008-present having random subsampling or no subsampling. Age compositions in the early time period had to be corrected using methods developed for Black Sea Bass during SEDAR 25 (Ballenger et al. 2011), while those for the most recent time period were summarized and did not have to be corrected. Because age compositions from earlier time periods required a correction utilizing length compositions, FL measurements ( 1990 - 2011 ) were converted to TL based on the meristic conversion above, while TL measurements ( 2012 - present) were unchanged. Age compositions for SEDAR 60 differed from the SEDAR 1, 2012 Updatedue to differences in ageing structure, ageing methodology, and the use of calendar age instead of increment count (Bubley et al. 2019a).

### 2.2.10 Discard Mortality

Discard mortality estimates were proposed for the commercial hook-and-line and recreational sectors. Logistic models based on observer data from the Gulf of Mexico were used to estimate a range of immediate mortality rates for each sector (Pulver 2018). In addition to immediate mortality, delayed mortality rates were estimated from literature and depredation mortality rates were proposed at the SEDAR 60 workshop. Depredation mortality rates from 5 to $10 \%$ were estimated for both sectors based on panelist input at the workshop. The total commercial hook-andline mortality estimate ranged from 45 to $64 \%$ with a proposed midpoint value of $53 \%$ (Table 8). The commercial immediate mortality estimates are the weighted logistic predictions from the SEFSC logbook with the midpoint value $(25 \%)$ assuming the majority of Red Porgy are being vented. The delayed mortality lower bounds of $26 \%$ is based on Rudershausen et al. (2007) and the upper bounds of $35 \%$ is based on 24 -hour cage survival at $46-54 \mathrm{~m}$ from Collins (1996). The total recreational hook-and-line mortality estimate ranged from 27 to $53 \%$ with a proposed midpoint value of $41 \%$ (Table 9). The recreational immediate mortality estimates are the weighted logistic prediction from
the SRHS eLog data with the midpoint value assuming $50 \%$ of Red Porgy are being vented. The delayed mortality lower bounds of $8 \%$ is based on 24 -hour cage survival at 36 m from Collins (1996) and the upper bounds of $26 \%$ is based on Rudershausen et al. (2007).

### 2.2.11 Discard Mortality Tables

Table 8. Red porgy commercial hook-and-line total discard mortality estimates (\%) based on a range of immediate, delayed, and depredation mortality values.

| Immediate | Delayed | Depredation | Total Discard |
| :---: | :---: | :---: | :---: |
| 20 | 26 | 5.0 | 45 |
| 25 | 30 | 7.5 | 53 |
| 35 | 35 | 10.0 | 64 |

Table 9. Red porgy recreational hook-and-line total discard mortality estimates (\%) based on a range of immediate, delayed, and depredation mortality values.

| Immediate | Delayed | Depredation | Total Discard |
| :---: | :---: | :---: | :---: |
| 16 | 8 | 5.0 | 27 |
| 22 | 17 | 7.5 | 41 |
| 28 | 26 | 10.0 | 53 |

## 3 Stock Assessment Methods

This assessment updates the primary model applied during the SEDAR 1, 2002 Benchmark, the SEDAR 1, 2006 Update and the SEDAR 1, 2012 Update for Red Porgy off the southeast United States. The methods are reviewed below, and any changes since the SEDAR 1, 2012 Update are emphasized.

### 3.1 Overview

The primary model in this assessment was the Beaufort assessment model (BAM), which applies a statistical catchage formulation. The model was implemented with the AD Model Builder software (Fournier et al. 2012). In essence, the model simulates a population forward in time while including fishing processes (Quinn and Deriso 1999; Shertzer et al. 2008). Quantities to be estimated are systematically varied until characteristics of the simulated populations match available data on the real population. Statistical catch-age models share many attributes with ADAPT-style tuned and untuned VPAs.

The method of forward projection has a long history in fishery models. It was introduced by Pella and Tomlinson (1969) for fitting production models and then, among many applications, used by Fournier and Archibald (1982), by Deriso et al. (1985) in their CAGEAN model, and by Methot (1989; 2009) in his Stock Synthesis model. The catch-age model of this assessment is similar in structure to the CAGEAN and Stock Synthesis models. Versions of this assessment model have been used in previous SEDAR assessments of reef fishes in the U.S. South Atlantic, such as Vermilion Snapper, Black Sea Bass, Golden Tilefish, Snowy Grouper, Gag Grouper, Greater Amberjack, Spanish Mackerel, Red Grouper, and Red Snapper, as well as in previous SEDAR assessments of Red Porgy (SEDAR 2002; 2006; 2012).

### 3.2 Data Sources

The catch-age model included data from four fleets that caught Red Porgy in southeastern U.S. waters: commercial hook-and-line (handline), commercial trawl, general recreational, and recreational headboat. The model was fitted to data on annual landings (in whole weight for commercial fleets and in numbers for recreational fleets), annual discard mortalities (in numbers for commercial handline and recreational fleets; Table 10). Data providers also provided CVs associated with landings and discards (Table 11), though these were only used to generate bootstrap data sets during the ensemble model analysis. The model was also fitted to annual length compositions of commercial trawl landings, annual age compositions of commercial handline and recreational headboat landings and SERFS Chevron trap catches. Samples sizes associated with composition data are provided in numbers of trips (Table 12) and numbers of fish (Table 13). The model was also fitted to one fishery dependent (Southeast Regional Headboat Survey) and one fishery independent (SERFS Chevron trap/video) index of abundance (Table 14). Data used in the model are tabulated in $\S 2$ of this report.

The general recreational fleet was sampled from 1981-2007 by the Marine Recreational Fishery Statistics Survey (MRFSS) and by the Marine Recreational Information Program (MRIP) since 2008. For years from 1972-1980, and as in SEDAR-1 and the 2006 and 2012 updates, landings values were assumed to be equal to the average landings from 1981-1990.

Data on annual discard mortalities, as fitted by the model, were computed by multiplying total discards (tabulated in $\S 2$ ) by the fleet-specific release mortality rates of 0.53 for the commercial handline fleet and 0.41 for the headboat and general recreational fleet (Pulver 2018).

### 3.3 Model Configuration and Equations

Model structure and equations of the BAM are detailed in Williams and Shertzer (2015). The assessment time period for this assessment was 1972-2017. A general description of the assessment model follows.

Stock dynamics In the assessment model, new biomass was acquired through growth and recruitment, while abundance of existing cohorts experienced exponential decay from fishing and natural mortality. The population was assumed closed to immigration and emigration. The model included age classes $1-14^{+}$, where the oldest age class $14^{+}$allowed for the accumulation of fish (i.e., plus group).

Initialization Initial (1972) abundance at age was estimated in the model as follows. First, the equilibrium age structure was computed for ages $1-14$ based on natural and fishing mortality ( $F_{\text {init }}$ ), where $F_{\text {init }}$ was assumed equal to the geometric mean of estimated F for the period 1972-1974. Second, lognormal deviations around that equilibrium age structure were estimated. The deviations were lightly penalized, such that the initial abundance of each age could vary from equilibrium if suggested by early composition data, but remain estimable if data were uninformative. Given the initial abundance of ages 2-14, initial (1972) abundance of age-1 fish was computed using the same methods as for recruits in other years (described below).

Natural mortality rate The natural mortality rate $(M)$ was assumed constant over time, but decreasing with age. The form of $M$ as a function of age was based on Charnov et al. (2013), a change from the SEDAR 1, 2012 Update which assumed natural mortality was constant across ages. The Charnov et al. (2013) approach inversely relates the natural mortality at age to somatic growth. As in previous SEDAR assessments, the age-dependent estimates of $M_{a}$ were rescaled to provide the same fraction of fish surviving from age 3 through the oldest observed age ( 25 yr ) as would occur with constant $M=0.22$. The constant value of $M$ was determined at the SEDAR 60 Workshop panel, as the average of six values, calculated from all combinations of three estimates of maximum age $\left(t_{\max }=20,25,30\right)$ and two methods of calculating $M$ as a function of $t_{\max }$ (Hoenig 1983; Then et al. 2014). This set of values was also used to develop a truncated normal distribution for the MCB analysis defined by the mean and standard deviation $(s=0.063)$ of these values truncated to a range of $0.14-0.32$.

Growth Mean length ( mm ) at age of the population (total length, TL) was modeled with the von Bertalanffy equation, and weight at age (whole weight, WW) was modeled as a function of total length (Table 15, Figure 3). Parameters of the relationship between TL and WW were specified by the SEDAR-1 DW and were treated as fixed input to the assessment model $\left(W W=(2.7 e-08) T L^{2.894}\right)$. Parameters of the von Bertalanffy equation relating TL and age $\left(T L=L_{\infty}\left(1-e^{-K\left(a-t_{0}\right)}\right)\right)$ were estimated external to the assessment model during the SEDAR 60 process, and input into the model as fixed values where $a=a g e+0.5=$ age at midyear, $L_{\infty}=422.6, K=0.3$, and $t_{0}=-1.47$. For fitting length composition data, the distribution of size at age was assumed normal with CV estimated external to the assessment model during the SEDAR 60 process $(\widehat{\mathrm{CV}}=0.136)$.

Spawning stock Spawning biomass was modeled as the biomass of mature female and male fish as in prior assessments of Red Porgy. Additionally, for protogynous fish like Red Porgy, computing spawning potential as a function of mature fish biomass has been shown to better account for the contribution of males when estimating biological reference points (Brooks et al. 2008). The sex ratio at age was assumed constant over time and estimated from fish captured in the MARMAP fishery independent monitoring program. This program also supplied the proportion of mature females at age (Table 15). Spawning biomass was computed at the approximate time of peak spawning in each year (February $1^{\text {st }}$; spawn_time_frac $=0.167$; Klibansky and Scharf 2013; MARMAP unpublished data).

Recruitment Expected recruitment of age-1 fish was predicted from spawning stock (biomass of mature fish) using the Beverton-Holt spawner-recruit model. As in the previous assessment, annual variation in recruitment was assumed to occur with lognormal deviations starting in 1975, when composition data could provide information
on year-class strength. In years prior, recruitment followed the Beverton-Holt model precisely, similar to an agestructured production model. Recruitment deviations in the last two years of the model were lightly constrained, penalizing extreme values, since the model has less information to inform recruitment deviations at the end of the time series.

Landings Time series of landings from four fleets were modeled (Table 10): commercial handline (1972-2017), commercial trawl (1972-1988), headboat (1972-2017), and general recreational (1972-2017). A zero value in trawl landings in 1974 was replaced with the smallest non-zero value in the time series ( 675 lb , reported for 1972). This has almost no effect on model results and was done largely for convenience. Landings were modeled with the Baranov catch equation (Baranov 1918) and were fitted in either weight or numbers, depending on how the data were collected [whole weight (mt) for commercial fleets and 1000 fish for recreational fleets)].

Discards In 1992 a 12 " size-limit went into effect for Red Porgy, and 1999 the size limit was increased to 14". Discard mortality data were available for commercial handline (1999-2017), recreational headboat (2001-2017), and general recreational fleets (1981-2017). The model estimated discards for all those years, and also estimated discards during the 12 " size-limit for commercial handline and recreational headboat (1992-1998), and during the beginning of the 14 " size-limit for recreational headboat (1999-2000; Table 10). During these periods, discards weren't available but were likely to have occurred. In years without observed discards, predicted discards were generated in the assessment model, by applying the fleet-specific geometric mean discard F from years with data. A zero value in general recreational discards in 1990 was replaced with the smallest non-zero value in the time series ( 200 fish, reported for 1991). This has almost no effect on model results and was done largely for convenience. As with landings, discard mortalities (in units of 1000 fish) were modeled with the Baranov catch equation (Baranov 1918), which required estimates of discard selectivities (described below) and release mortality rates. New discard mortality rate estimates were developed for this SEDAR 60 assessment. Fleet-specific release mortality rates were 0.53 for the commercial handline fleet and 0.41 for the headboat and general recreational fleets (Pulver 2018)

Fishing For each time series of landings and discard mortalities, the assessment model estimated a separate full fishing mortality rate $(F)$. Age-specific rates were then computed as the product of full $F$ and selectivity at age. Apical $F$ was computed as the maximum of $F$ at age summed across fleets.

Selectivities In all cases, selectivity at age was estimated using a two-parameter logistic model. This parametric approach reduces the number of estimated parameters and imposes theoretical structure on selectivity. Age and size composition data are critical for estimating selectivity functions.

Selectivity of each fishery was generally fixed within each period of size-limit regulations, but was permitted to vary among periods. With the exception of the commercial trawl fishery, all fisheries experienced three periods of size-limit regulations (no limit prior to 1992, 12 " limit during 1992-1998, 14 " limit 1999-2017). Ideally, a model would have sufficient age composition data from each fishery over time to estimate selectivities in each period of regulations. That was not the case here, and thus additional assumptions were applied to define selectivities, as follows.

Logistic selectivity functions were estimated for the commercial handline fleet informed by age composition data during two regulatory periods (1972-1998, 1999-2017). No age composition data were available during 1972-1991 for commercial handline to estimate a separate selectivity for this period. A logistic selectivity function was estimated for the commercial trawl fleet during the first regulatory period based on length composition data pooled across years of available data (1977, 1979, 1984, 1986-1988). A logistic selectivity function was estimated for the headboat fleet during each regulatory time period and informed by age composition data. Following previous assessments, the selectivity of the general recreational fleet was set equal to the headboat fleet during each regulatory period. The SERFS chevron trap/video selectivity was estimated to be logistic based solely on age composition data from chevron trap catches (no age or length data is collected directly from videos), departing from the use of dome-shaped
selectivity in previous assessments of Red Porgy. The change was made based on examination of the age composition data and the consensus of panel members that the chevron traps do not exclude large Red Porgy.

Similar to the methods in previous assessments of Red Porgy, discard selectivities of the commercial handline and recreational fleets were informed by the selectivities of the landings. For each fleet, the discard selectivity at each age was assumed to be the maximum landing selectivity at age across the all regulatory time periods. Since the selectivity of landings were identical for the headboat and general recreational fleets, the discard selectivities were also identical.

In this assessment, no selectivity parameters were fixed, but normal prior distributions were applied to slope parameters during estimation. Priors were relatively light $(C V=1.0)$, only loosely guiding the estimation of these slope parameters.

Indices of abundance The model was fitted to one fishery dependent index of abundance (headboat 1973-1998) and one fishery independent index of abundance (SERFS Chevron trap/video 1990-2017; Table 14). Predicted indices were computed from numbers at age at the beginning of the year.

Catchability In the BAM, catchability scales indices of relative abundance to the estimated vulnerable population at large. As in prior assessments, catchability coefficients of both indices (fishery independent and fishery dependent) were assumed constant. Thus, the fishery dependent index (headboat fleet) was not assumed to have a technologically induced trend in catchability as has been hypothesized in some SEDAR assessments (SEDAR Procedural Guidance 2009).

Biological reference points Biological reference points (benchmarks) were calculated based on maximum sustainable yield (MSY) estimates from the Beverton-Holt spawner-recruit model with bias correction (expected values in arithmetic space). Computed benchmarks included MSY, fishing mortality rate at MSY ( $F_{\text {MSY }}$ ), and spawning stock at MSY ( $\mathrm{SSB}_{\mathrm{MSY}}$ ). In this assessment, spawning stock measures the biomass of all mature fish (both sexes) in the population. These benchmarks are conditional on the estimated selectivity functions and the relative contributions of each fleet's fishing mortality. The selectivity pattern used here was the effort-weighted selectivities at age, with effort from each fishery (including discard mortalities) estimated as the full $F$ averaged over the last three years of the assessment.

Fitting criterion The fitting criterion was a likelihood approach in which observed landings and discards were fit closely, and observed composition data and abundance indices were fit to the degree that they were compatible. Landings, discards, and index data were fit using lognormal likelihoods. Length and age composition data were fit using the Dirichlet-multinomial distribution, with sample size represented by the annual number of trips (Table 14), adjusted by an estimated variance inflation factor. The previous assessment fit composition data using multinomial likelihoods, and many SEDAR assessments since then have applied a robust version of the multinomial likelihood, as recommended by Francis (2011). More recent work has questioned use of the multinomial distribution in stock assessment models (Francis 2014), and of the alternative distributions, two appear most promising, the Dirichletmultinomial and logistic-normal (Francis 2017; Thorson et al. 2017). Both are self-weighting and therefore iterative re-weighting (e.g. Francis 2011) is unnecessary, and both better account for intra-haul correlations (i.e., fish caught in the same set are more alike in length or age than fish caught in a different set). The Dirichlet-multinomial allows for observed zeros (the logistic-normal does not), and has recently been implemented in Stock Synthesis (Methot and Wetzel 2013).

The model includes the capability for each component of the likelihood to be weighted by user-supplied values. When applied to landings and indices, these weights modify the effect of the input CVs. In this application to Red Porgy, CVs of landings (in arithmetic space) were assumed equal to 0.05 to achieve a close fit to these data while allowing some imprecision. In practice, the small CVs are a matter of computational convenience, as they help achieve a close
fit to the landings, while avoiding having to solve the Baranov equation iteratively (which is complex when there are multiple fisheries). In contrast to the previous assessment of Red Porgy, weights of likelihood components were not varied during model development, and were all equal in the base model.

Configuration of base run The base run was configured as described above. However, the base run configuration was not considered to represent all uncertainty. Sensitivity analyses, retrospective analyses, and ensemble modeling was conducted to better characterize the uncertainty in base run point estimates.

Sensitivity analyses Sensitivity of results to some key model inputs and assumptions was examined through sensitivity analyses. Sensitivity runs were chosen to investigate issues that arose specifically with SEDAR 60. These model runs vary from the base run as follows.

- S1: Low value of natural mortality $(M=0.14)$
- S2: High value of natural mortality $(M=0.32)$
- S3: Low value of (fixed) steepness $(h=0.25)$
- S4: High value of (fixed) steepness $(h=0.51)$
- S5: Low value of (fixed) $R_{0}\left(\log \left(R_{0}\right)=13.9\right)$
- S6: Include MARMAP Florida Snapper Trap Index and age composition data
- S7: Include female maturity at age as a time-varying vector
- S8-S9: Upweight headboat index: $2 \times, 3 \times$
- S10: Replace 2016 MRIP landings and discards values, with average of values from 2015 and 2017

Retrospective analyses Retrospective analyses were run by reducing the terminal year of the model from 2017 to 2011-2016, thereby trimming all time series accordingly, and rerunning the assessment model. This analysis facilitates investigation of patterns in model results, particularly terminal status estimates, that may occur when recent data are excluded.

### 3.4 Parameters Estimated

The model estimated deviations in the initial age structure (13 parameters), average fishing mortality rates (7 parameters) and annual fishing mortality rates (228 parameters) for each fleet, selectivity parameters (12 parameters), Dirichlet-multinomial variance inflation factors (4 parameters), a catchability coefficient associated with each index (2 parameters), steepness of the stock-recruit relationship and initial mean recruitment (2 parameters), variance of the recruitment deviations (1 parameter), and annual recruitment deviations (43 parameters).

### 3.5 Per Recruit and Equilibrium Analyses

Yield per recruit and spawning potential ratio were computed as functions of $F$, as were equilibrium landings and spawning biomass. Equilibrium landings and discards were also computed as functions of biomass $B$, which itself is a function of $F$. As in computation of MSY-related benchmarks (described in §3.6), per recruit and equilibrium analyses applied the most recent selectivity patterns averaged across fleets, weighted by each fleet's $F$ from the last three years (2015-2017) of the assessment.

### 3.6 Benchmark/Reference Point Methods

In this assessment of Red Porgy, the quantities $F_{\mathrm{MSY}}, \mathrm{SSB}_{\mathrm{MSY}}, B_{\mathrm{MSY}}$, and MSY were estimated by the method of Shepherd (1982). In that method, the point of maximum yield is calculated from the spawner-recruit curve and parameters describing growth, natural mortality, maturity, and selectivity. The value of $F_{\text {MSY }}$ is the $F$ that maximizes equilibrium landings.

On average, expected recruitment is higher than that estimated directly from the spawner-recruit curve, because of lognormal deviation in recruitment. Thus, in this assessment, the method of benchmark estimation accounted for lognormal deviation by including a bias correction in equilibrium recruitment. The bias correction ( $\varsigma$ ) was computed from the variance $\left(\sigma_{R}^{2}\right)$ of recruitment deviation in $\log$ space: $\varsigma=\exp \left(\sigma_{R}^{2} / 2\right)$. Then, equilibrium recruitment $\left(R_{e q}\right)$ associated with any $F$ is,

$$
\begin{equation*}
R_{e q}=\frac{R_{0}\left[\varsigma 0.8 h \Phi_{F}-0.2(1-h)\right]}{(h-0.2) \Phi_{F}} \tag{4}
\end{equation*}
$$

where $R_{0}$ is virgin recruitment, $h$ is steepness, and $\Phi_{F}$ is spawning potential ratio given growth, maturity, and total mortality at age (including natural, fishing, and discard mortality rates). The $R_{e q}$ and mortality schedule imply an equilibrium age structure and an average sustainable yield (ASY). The estimate of $F_{\text {MSY }}$ is the $F$ giving the highest ASY (excluding discards), and the estimate of MSY is that ASY. The estimate of $\mathrm{SSB}_{\mathrm{MSY}}$ follows from the corresponding equilibrium age structure, as does the estimate of discard mortalities $\left(D_{\text {MSY }}\right)$, here separated from ASY (and consequently, MSY).

Estimates of MSY and related benchmarks are conditional on selectivity pattern. The selectivity pattern used here was an average of terminal-year selectivities from each fleet, where each fleet-specific selectivity was weighted in proportion to its corresponding estimate of $F$ averaged over the last three years (2015-2017) of the assessment. If the selectivities or relative fishing mortalities among fleets were to change, so would the estimates of MSY and related benchmarks.

The maximum fishing mortality threshold (MFMT) is defined by the SAFMC as $F_{\text {MSY }}$, and the minimum stock size threshold (MSST) as MSST $=(1-M)$ SSB $_{\text {MSY }}$ (Restrepo et al. 1998), with constant $M$ here equated to 0.22 . Overfishing is defined as $F>$ MFMT and overfished as SSB $<$ MSST. Current status of the stock is represented by SSB in the last assessment year (2017), and current status of the fishery is represented by the geometric mean of $F$ from the last three years (2015-2017).

In addition to the MSY-related benchmarks, the assessment considered proxies based on per recruit analyses (e.g., $F_{40 \%}$ ). The values of $F_{X \%}$ are defined as those $F$ s corresponding to $\mathrm{X} \%$ spawning potential ratio, i.e., spawners (spawning biomass) per recruit relative to that at the unfished level. These quantities may serve as proxies for $F_{\text {MSY }}$, if the spawner-recruit relationship cannot be estimated reliably. Mace (1994) recommended $F_{40 \%}$ as a proxy; however, later studies have found that $F_{40 \%}$ is too high of a fishing rate across many life-history strategies (Williams and Shertzer 2003; Brooks et al. 2009) and can lead to undesirably low levels of biomass and recruitment (Clark 2002).

### 3.7 Comparison to Previous Assessments

This SEDAR 60 standard assessment builds upon the SEDAR 1, 2012 Update with an additional 6 years of data, substantial improvements to the structure of the Beaufort Assessment Model, and several changes to the configuration of the model, generally simplifying the data structure. The only new data source included in SEDAR 60 was the SERFS video index, which showed a similar trend as the SERFS chevron trap, and was combined with it.

Changes to the life history information used in the model included:

1. Updated estimates of constant natural mortality based on new estimates of maximum age
2. Included age-varying natural mortality, following current SEDAR standards for most assessments
3. Updated most estimates of life history parameters, including more recent data
4. Treated female maturity-at-age as constant over time
5. Time of spawning changed from default January $1^{\text {st }}$ to a value of February $1^{\text {st }}$ based on empirical data
6. Much more uncertainty in $M$ incorporated into MCB analysis in SEDAR $60(0.14-0.32)$ than in the 2012 update (0.20-0.25)

Changes to model configuration include:

1. The youngest age modeled is age-1 (there were very few age- 0 fish in the age composition data)
2. Initialization of numbers at age in 1972 was done using a method used in SEDAR (2017), where equilibrium age structure is computed and deviations at age were estimated. In contrast, in the SEDAR 1, 2012 Update assessment, the model started in 1958 assuming the population was at $90 \%$ of virgin biomass, and estimated recruitment deviations for these 14 early years to inform age structure in 1972.
3. Growth model parameters are fixed (i.e. not estimated) within the model
4. Selectivity of commercial handline included only two time blocks, since there was not age composition data available to inform selectivity in the earliest time block
5. Selectivity of SERFS chevron trap was changed from dome-shaped to flat-topped (logistic)
6. Length and age compositions were fit using Dirichlet multinomial likelihoods, compared with multinomial likelihoods used in the SEDAR 1, 2012 Update
7. Data sources being fitted were not re-weighted by user-supplied weights. In the SEDAR 1, 2012 Update assessment, data weights were treated as inputs and varied across data sources.

Changes in data structure include:

1. A zero value in trawl landings in 1974 was replaced with the smallest non-zero value in the time series ( 675 lb , as reported for 1972). This has almost no effect on model results and was done largely for convenience. The SEDAR 1, 2012 Update assessment made several changes to the model to allow the model to run properly with a zero landings value.
2. Combined the relatively small amount ( $3 \%$ of total commercial landings) of commercial trap landings with commercial handline landings
3. Excluded most length composition information data, which conflicted with age composition data
4. Pooled annual commercial trawl length composition data into a single composition
5. Excluded MARMAP Florida snapper trap index and corresponding age and length composition data

### 3.8 Uncertainty and Measures of Precision

For the base run of the catch-age model (BAM), uncertainty in results and precision of estimates was computed thoroughly through an ensemble modeling approach (Scott et al. 2016) using a mixed Monte Carlo and bootstrap framework (Efron and Tibshirani 1993; Manly 1997). Monte Carlo and bootstrap methods are often used to characterize uncertainty in ecological studies, and the mixed approach has been applied successfully in stock assessment (Restrepo et al. 1992; Legault et al. 2001; SEDAR4 2004; SEDAR19 2009; SEDAR24 2010). The approach is among those recommended for use in SEDAR assessments (SEDAR Procedural Guidance 2010).

The approach translates uncertainty in model input into uncertainty in model output, by fitting the model many times with different values of "observed" data and key input parameters. A chief advantage of the approach is that the results describe a range of possible outcomes, so that uncertainty is characterized more thoroughly than it could be by any single fit or small set of sensitivity runs. A minor disadvantage of the approach is that computation times can be long, though current parallel computing techniques largely mitigate those demands [i.e. computing results many times (e.g. $40 \times$ ) as fast as a single processor].

In this assessment, the BAM was re-fit in $n=4000$ trials that differed from the original inputs by bootstrapping on data sources, and by Monte Carlo sampling of several key input parameters. Of the 4000 trials, 3350 were ultimately retained in the uncertainty analysis. The remaining runs were discarded because of poor model convergence or because values of $R_{0}$ were in the extreme tails of the distribution among all runs (lower and upper $0.5 \%$ ).

The MCB analysis should be interpreted as providing an approximation to the uncertainty associated with each output. The results are approximate for two related reasons. First, not all combinations of Monte Carlo parameter inputs are equally likely, as biological parameters might be correlated. Second, all runs are given equal weight in the results, yet some might provide better fits to data than others.

### 3.8.1 Bootstrapping of Observed Data

To include uncertainty in time series of observed landings, discards, and indices of abundance, multiplicative lognormal errors were applied through a parametric bootstrap. To implement this approach in the MCB trials, random variables $\left(x_{s, y}\right)$ were drawn for each year $y$ of time series $s$ from a normal distribution with mean 0 and variance $\sigma_{s, y}^{2}$ [that is, $x_{s, y} \sim N\left(0, \sigma_{s, y}^{2}\right)$ ]. Annual observations were then perturbed from their original values $\left(\hat{O}_{s, y}\right)$,

$$
\begin{equation*}
O_{s, y}=\hat{O}_{s, y}\left[\exp \left(x_{s, y}-\sigma_{s, y}^{2} / 2\right)\right] \tag{5}
\end{equation*}
$$

The term $\sigma_{s, y}^{2} / 2$ is a bias correction that centers the multiplicative error on the value of 1.0. Standard deviations in $\log$ space were computed from CVs in arithmetic space, $\sigma_{s, y}=\sqrt{\log \left(1.0+C V_{s, y}^{2}\right)}$. The CVs used to generate bootstrap data sets of landings and discards were supplied by the data providers (Table 11). Note that these values are different and generally higher than the CVs used to estimate landings and discards when fitting the assessment
model (i.e. 0.05 for all years and fleets). The CVs used to generate bootstrap data sets of indices of abundance were the same as those used when fitting the assessment model (Table 14).

Uncertainty in age and length compositions were included by drawing new distributions for each year of each data source, following a multinomial sampling process. Ages (or lengths) of individual fish (Table 13) were drawn at random with replacement using the cell probabilities of the original data. For each year of each data source, the number of fish sampled was the same as in the original data (Table 14).

### 3.8.2 Monte Carlo Sampling

In each successive fit of the model, several parameters were fixed (i.e., not estimated) at values drawn at random from distributions described below.

Natural mortality The point estimate of natural mortality ( $M=0.22$ ) was provided by the SEDAR 60 Workshop Panel with some uncertainty. To carry forward this source of uncertainty, Monte Carlo sampling was used to generate deviations from the point estimate. A new $M$ value was drawn for each MCB trial from a truncated normal distribution (described above) defined by the mean of 0.22 and standard deviation of 0.063 , and truncated to a range of $0.14-0.32$. In each run of the ensemble, a drawn value of constant $M$ was then used to rescale natural mortality at age, as described for the base model above.

Discard mortalities Similarly, discard mortalities $\delta$ were subjected to Monte Carlo variation as follows. New values for commercial handline were drawn for each MCB trial from a uniform distribution (range [0.45, 0.64]), and new values for recreational fleets (headboat and general recreational) were drawn from a uniform distribution (range [0.27, 0.53]).

### 3.9 Projection Methods

Projections were run to determine the overfishing limit (OFL) and evaluate the existing rebuilding plan as requested in the TORs. The structure of the projection model was the same as that of the assessment model, and parameter estimates were those from the assessment. Any time-varying quantities, such as selectivity, were fixed to the most recent values of the assessment period. A single selectivity curve was applied to calculate landings computed by averaging selectivities across fleets using geometric mean Fs from the last three years of the assessment period, similar to computation of MSY benchmarks (§3.6).

Expected values of SSB (time of peak spawning), $F$, recruits, and landings were represented by deterministic projections using parameter estimates from the base run. These projections were built on the estimated spawner-recruit relationship with bias correction, and were thus consistent with estimated benchmarks in the sense that long-term fishing at $F_{\text {MSY }}$ would yield MSY from a stock size at $\mathrm{SSB}_{\mathrm{MSY}}$. Uncertainty in future time series was quantified through stochastic projections that extended the ensemble model fits of the stock assessment model.

### 3.9.1 Initialization of Projections

Although the terminal year of the assessment is 2017, the assessment model computes abundance at age $\left(N_{a}\right)$ at the start of 2018. For projections, those estimates were used to initialize $N_{a}$. However, the assessment has no information to inform the strength of 2018 recruitment, and thus it computes 2018 recruits $\left(N_{1}\right)$ as the expected value, that is, without deviation from the spawner-recruit curve, and corrected to be unbiased in arithmetic space. In the stochastic projections, lognormal stochasticity was applied to these abundances after adjusting them to be
unbiased in $\log$ space, with variability based on the estimate of $\sigma_{R}$. Thus, the initial abundance in year one (2018) of projections included this variability in $N_{1}$. The deterministic projections were not adjusted in this manner, because deterministic recruitment follows Beverton-Holt expectation.

Fishing rates that define the projections were assumed to start in 2021. Because the assessment period ended in 2017, the projections required an interim period (2018-2020). Fishing mortality during this interim period was set at $F_{\text {current }}=0.31$.

### 3.9.2 Uncertainty of Projections

To characterize uncertainty in future stock dynamics, stochasticity was included in replicate projections, each an extension of a single assessment fit from the ensemble. Thus, projections carried forward uncertainties in natural mortality and discard mortality, as well as in estimated quantities such as spawner-recruit parameters ( $R_{0}$ and $\sigma_{R}$, selectivity curves, and in initial (start of 2018) abundance at age.

Initial and subsequent recruitment values were generated with stochasticity using a Monte Carlo procedure, in which the estimated recruitment of each model within the ensemble is used to compute mean annual recruitment values $\left(\bar{R}_{y}\right)$. Variability is added to the mean values by choosing multiplicative deviations at random from a lognormal distribution,

$$
\begin{equation*}
R_{y}=\bar{R}_{y} \exp \left(\epsilon_{y}\right) \tag{6}
\end{equation*}
$$

Here $\epsilon_{y}$ is drawn from a normal distribution with mean 0 and standard deviation $\sigma_{R}$, where $\sigma_{R}$ is the standard deviation from the relevant ensemble model component.

The procedure generated 20,000 replicate projections of models within the ensemble drawn at random (with replacement). In cases where the same model run was drawn, projections would still differ as a result of stochasticity in projected recruitment streams. Central tendencies were represented by the deterministic projections of the base run, as well as by medians of the stochastic projections. Precision of projections was represented graphically by the $5^{t h}$ and $95^{t h}$ percentiles of the replicate projections.

### 3.9.3 Projection Scenarios

Projections were run to determine the overfishing limit (OFL) and evaluate the existing rebuilding plan as requested in the TORs. In the projections, management started in 2021, the earliest year possible at the time of writing. Projections were made out to 2026 or 2032 , Scenarios 1 and 2 were considered to determine the OFL and scenarios 3 and 4 were considered to evaluate the existing rebuilding plan:

- Scenario 1: $F=F_{\mathrm{P}_{50 \%}^{*}}$ from 2021 to 2026 , and with $F=F_{\text {current }}$ from 2018 to 2020.
- Scenario 2: $F=F_{\text {MSY }}$ from 2021 to 2026, with $F=F_{\text {current }}$ from 2018 to 2020.
- Scenario 3: $F=75 \% F_{\text {MSY }}$ from 2021 to 2026 , with $F=F_{\text {current }}$ from 2018 to 2020.
- Scenario 4: $F=0$ from 2021 to 2032 , with $F=F_{\text {current }}$ from 2018 to 2020.


## 4 Stock Assessment Results

### 4.1 Measures of Overall Model Fit

The Beaufort assessment model (BAM) generally fit well to the available data. Predicted age compositions from each fishery were reasonably close to observed data in most years, as were predicted length compositions for the commercial trawl fleet (Figure 4). The model was configured to fit observed commercial and recreational landings closely (Figures 5, 6, 7, 8), as well as observed discards (Figures 9, 10, 11). Fits to indices of abundance captured the general trends but not all annual fluctuations (Figures 12, 13). The model tended to overestimate the headboat index from 1990-1998, years which overlapped with the SERFS chevron trap/video index.

### 4.2 Parameter Estimates

Estimates of all parameters from the catch-age model are shown in Appendix B. Estimates of management quantities and some key parameters, such as those of the spawner-recruit model, are reported in sections below.

### 4.3 Stock Abundance and Recruitment

Estimated abundance at age showed truncation of the older ages beginning in the 1980s (Figure 14; Table 16). The model predicted a large recruitment event in 1977 driving high abundance for several subsequent years. Total abundance was high during the 1970s peaking in 1977 and declining to a low in 2000. Abundance increased modestly during the 2000s to a peak in 2011, but has been declining from 2012 to the end of the assessment period. Total estimated abundance was at its lowest values at the end of the time series. Annual number of recruits is shown in Table 16 (age-1 column) and in Figure 16. In the most recent decade, the strongest year class (age-1 fish) was predicted to have occurred in 2010. The SEDAR 1, 2012 Update assessment report (SEDAR 2012) noted below average recruitment during the last five years (2007-2011). In the current assessment, predicted recruitment values during the last six years (2013-2018) are among the lowest for the entire time series.

### 4.4 Total and Spawning Biomass

Estimated biomass at age followed a similar pattern as abundance at age (Figure 15; Tables 17, 18). Total biomass and spawning biomass showed similar trends-general decline from the late 1970s to 2000, followed by a gradual recovery through 2011 , followed by a rapid decline to the end of the time series (Figure 17; Table 19).

### 4.5 Selectivity

Selectivity of the SERFS chevron trap/video survey is shown in (Figure 18), selectivities of landings from commercial and recreational fleets are shown in Figures 19 and 20. In the most recent years, full selection occurred near age5 to age-6, depending on the fleet. Selectivities of discard mortalities were a function of logistic shaped landings selectivities (Figures 21).

Average selectivities of landings and of discard mortalities were computed from $F$-weighted selectivities in the most recent period of regulations (Figure 22). These average selectivities were used to compute point estimates of benchmarks. All selectivities from the most recent period, including average selectivities, are tabulated in Table 20.

### 4.6 Fishing Mortality, Landings, and Discards

The estimated fishing mortality rates $(F)$ showed an increasing trend from the early 1970s to peak levels in 1990. Subsequently, $F$ declined to lower levels since 2000, with two high values at the end of the time series largely attributed to the general recreational fleet (Figure 23). The commercial handline fleet had been the largest contributor to total $F$ during much of the 1980 s and 1990 s, but since 2000 large and sometimes predominant proportions of total fishing mortality have been attributed to recreational fleets (Table 21).

The overall pattern in landings over time is similar to the pattern in $F$, though the decrease in total landings since 2000 compared with earlier landings, is more substantial than the corresponding decrease in $F$. Landings have been low in the past two decades, but so has abundance, so the $F$ is still comparatively higher. A majority of estimated landings during the 1980s and 1990s were from the commercial sector, but since the early 2000s, larger proportions of Red Porgy landings have come from the recreational sector in many years (Figures 24, 25; Tables 25, 26). Estimated discard mortalities occurred on a much smaller scale than landings ( $5 \%$ of removals by numbers). Both the commercial and recreational sectors contribute substantially to discards. Dead discards have been highly variable with a notable peak in 2002 largely attributed to the commercial handline fleet and peaks in 2008 and 2016 largely attributed to the general recreational fleet (Figures 26 and 27; Tables 27, 28).

### 4.7 Spawner-Recruitment Parameters

The estimated Beverton-Holt spawner-recruit curve is shown in Figure 28, along with the effect of density dependence on recruitment, depicted graphically by recruits per spawner as a function of spawners (spawning biomass). Values of recruitment-related parameters were as follows: steepness $\hat{h}=0.38$, unfished age- 1 recruitment $\widehat{R}_{0}=3,430,000$, unfished spawning biomass (mt) per recruit $\phi_{0}=0.00174$, and standard deviation of recruitment residuals in log space $\widehat{\sigma}_{R}=0.45$ (which resulted in bias correction of $\varsigma=1.11$ ). Uncertainty in these quantities was estimated through the Monte Carlo/bootstrap (MCB) analysis (Figure 29).

### 4.8 Per Recruit and Equilibrium Analyses

Yield per recruit and spawning potential ratio were computed as functions of $F$ (Figure 30). As in computation of MSY-related benchmarks, per recruit analyses applied the most recent selectivity patterns averaged across fisheries, weighted by $F$ from the last three years $(2015-2017)$. The $F$ that provides $40 \% \mathrm{SPR}$ is $F_{40 \%}=0.6,30 \%$ is $F_{30 \%}=1.21$, and $20 \%$ is $F_{20 \%}=3.12$.

As in per recruit analyses, equilibrium landings and spawning biomass were computed as functions of $F$ (Figure 31). By definition, the $F$ that maximizes equilibrium landings is $F_{\text {MSY }}$, and the corresponding landings and spawning biomass are MSY and $\mathrm{SSB}_{\text {MSY }}$.

### 4.9 Benchmarks / Reference Points

As described in §3.6, biological reference points (benchmarks) were derived analytically assuming equilibrium dynamics, corresponding to the expected spawner-recruit curve (Figure 28). Reference points estimated were $F_{\text {MSY }}$, MSY, $B_{\mathrm{MSY}}$ and $\mathrm{SSB}_{\mathrm{MSY}}$. Based on $F_{\mathrm{MSY}}$, three possible values of $F$ at optimum yield (OY) were considered$F_{\mathrm{OY}}=65 \% F_{\mathrm{MSY}}, F_{\mathrm{OY}}=75 \% F_{\mathrm{MSY}}$, and $F_{\mathrm{OY}}=85 \% F_{\mathrm{MSY}}$ - and for each, the corresponding yield was computed. Estimates of benchmarks are summarized in Table 29. Standard errors of benchmarks were approximated as those from Monte Carlo/bootstrap analysis (§3.8).

Maximum likelihood estimates (base run) of benchmarks, as well as median values from MCB analysis, are summarized in Table 29. Point estimates of MSY-related quantities were $F_{\text {MSY }}=0.18\left(\mathrm{y}^{-1}\right)$, MSY $=531$ (1000 lb), $B_{\mathrm{MSY}}=3605(\mathrm{mt}), \mathrm{MSST}=2249(\mathrm{mt})$, and $\mathrm{SSB}_{\mathrm{MSY}}=2884(\mathrm{mt})$. The estimate of $\mathrm{SSB}_{\mathrm{MSY}}$ is about $48 \%$ of the unfished spawning biomass. Median estimates were $F_{\mathrm{MSY}}=0.18\left(\mathrm{y}^{-1}\right)$, MSY $=538(1000 \mathrm{lb}), B_{\mathrm{MSY}}=3594(\mathrm{mt})$, $\operatorname{MSST}=2261$, and $\mathrm{SSB}_{\mathrm{MSY}}=2903(\mathrm{mt})$. Distributions of these benchmarks from the MCB analysis are shown in Figure 32.

### 4.10 Status of the Stock and Fishery

Estimated time series of stock status ( $\mathrm{SSB} / \mathrm{MSST}$ and $\mathrm{SSB} / \mathrm{SSB}_{\mathrm{MSY}}$ ) showed a rapid decline from favorable stock sizes in the 1970s to low levels in the late 1990s. From 1999 to 2011 stock status was gradually recovering, but has been in decline again since 2012 (Figure 33, Table 19). The increasing trend observed from 1998 to 2011 appears to have been driven largely by decreases in landings, while being hampered by low recruitment.

Current stock status was estimated in the base run to be $\mathrm{SSB}_{2017} / \mathrm{MSST}=0.347$ and $\mathrm{SSB}_{2017} / \mathrm{SSB}_{\mathrm{MSY}}=0.27$ (Table 29), indicating that the stock remains in an overfished state. Results from the MCB analysis suggested that the estimate of SSB relative to $\mathrm{SSB}_{\mathrm{MSY}}$ and the status relative to MSST are robust, and there is little uncertainty in the overfished status (Figures 34, 35). Age structure estimated by the base run during 2017 suggests that the age composition of the population above age- 5 has been recovering toward what is expected at MSY, but abundances at age- 1 to age- 5 are as low as they have ever been (Figure 36). This finding further suggests that reduced fishing mortalities have been promoting recovery of older age classes but below average recruitment has stifled recovery for younger age classes.

The estimated time series of $F / F_{\text {MSY }}$ suggests that overfishing has been occurring throughout most of the assessment period (Table 19), but with some uncertainty demonstrated by the MCB analysis (Figure 33). Current fishery status in the terminal year, with current $F$ represented by the geometric mean from $2015-2017$, was estimated by the base run to be $F_{2015-2017} / F_{\mathrm{MSY}}=1.73$ (Table 29). Thus at the end of the assessment Red Porgy was undergoing overfishing. Results from the MCB analysis show that there is little uncertainty in the status of the fishery (Figures 34, 35). Note that $F_{\text {MSY }}$ is based on average F's from last three years of the assessment and thus it is not the technically correct denominator for all years going back in time. Thus caution should be applied when interpreting F status back in time.

### 4.11 Comparison to Previous Assessments

In 1992, an initial rebuilding plan was put into effect for Red Porgy (SAFMC Amendment 4, 01/01/1992) with a rebuilding time frame of 10 years, beginning in 1991. In 2000, a new rebuilding plan was put into effect (SAFMC Amendment $12,09 / 12 / 2000$ ) with a rebuilding time frame of 18 years, beginning in 1999.

The first SEDAR stock assessment of Red Porgy (SEDAR 01; SEDAR 2002) modeled the population from 19722001. Assessment model timelines for subsequent assessments are as follows: SEDAR 01, 2006 update assessment (1972-2004 SEDAR 2006), SEDAR 01, 2012 update assessment (1972-2011 SEDAR 2012), and the current SEDAR 60 standard assessment (1972-2017).

As of 2001, the stock was overfished $\left(\mathrm{SSB}_{2001} / \mathrm{MSST}=0.55 ; \mathrm{SSB}_{2001} / \mathrm{SSB}_{\mathrm{MSY}}=0.43\right)$, but overfishing was not occurring $\left(F_{2001} / F_{\mathrm{MSY}}=0.45\right.$; SEDAR 2002). Projections from SEDAR-1 found that under the Amendment 12 scenario, the probability of being rebuilt by the terminal years of subsequent assessments (2004, 2011, or 2017) to be $\approx 0 \%, \approx 5 \%$, and $\approx 38 \%$. Terminal status estimates the 2006 and 2012 updates found that the Red Porgy stock
was not rebuilt $\left(\mathrm{SSB}_{2005} / \mathrm{SSB}_{\mathrm{MSY}}=0.66\right.$ and $\left.\mathrm{SSB}_{2011} / \mathrm{SSB}_{\mathrm{MSY}}=0.47\right)$, but was also not undergoing overfishing $\left(F_{2004} / F_{\mathrm{MSY}}=0.45\right.$ and $\left.F_{2009-2011} / F_{\mathrm{MSY}}=0.64\right)$.

Values from the current SEDAR 60 assessment support the stock status designations from all three previous SEDAR assessments, with values in similar ranges for $\mathrm{SSB}_{2001} / \mathrm{SSB}_{\mathrm{MSY}}=0.36, \mathrm{SSB}_{2005} / \mathrm{SSB}_{\mathrm{MSY}}=0.43$, and $\mathrm{SSB}_{2011} / \mathrm{SSB}_{\mathrm{MSY}}=0.51$ (Table 19). However, the current assessement results suggest that overfishing was occurring in 2001 and $2004\left(F_{2001} / F_{\mathrm{MSY}}=1.19, F_{2004} / F_{\mathrm{MSY}}=1.27\right)$ and at the end of SEDAR 1, 2012 Update $\left(F_{2009-2011} / F_{\mathrm{MSY}}=1.05\right)$. In general, time series of $\mathrm{SSB} / \mathrm{SSB}_{\mathrm{MSY}}$ and $F / F_{\mathrm{MSY}}$ produced in SEDAR 60 are similar to time series from previous assessments, where they overlapped. However, compared with SEDAR 1, 2012 Update estimates from SEDAR 60 of biomass and SSB were higher early in the time series (e.g. 1972-1980) and lower from 1990-2000, while full $F$ was higher from 1990-2000. These differences may be due in part to differences in initialization methods and weighting of the headboat index in SEDAR 1, 2012 Update, but are probably mostly due to the revised MRIP estimates, which are higher now.

Input values of constant $M$ have been similar over the four Red Porgy assessments (terminal years: 2001, 2004, 2011, $2017 ; M: 0.225,0.225,0.225,0.22)$. Steepness has been estimated in all assessments and estimates have generally decreased over time $(h: 0.48,0.50,0.41,0.38)$. The contrast in SSB over time is generally considered to be a scenario that is informative of steepness. Estimates of $F_{\text {MSY }}$ have remained in a similar range ( $F_{\text {MSY }}: 0.19,0.20,0.17,0.18$ ). Estimates of MSY and $\mathrm{SSB}_{\text {MSY }}$ are lower in the current assessment [MSY ( 1000 lb ): 826, 626, 834, 531; $\mathrm{SSB}_{\text {MSY }}$, (mt): 3050, 3236, 3933, 2884].

### 4.12 Sensitivity Analyses

Sensitivity runs, described in $\S 3.3$, may be useful for evaluating implications of assumptions in the base assessment model, and for interpreting MCB results in terms of expected effects from input parameters. Time series of $F / F_{\text {MSY }}$, $\mathrm{SSB} / \mathrm{MSST}, B$, and recruitment are plotted to demonstrate sensitivity to natural mortality (Figure 37), the steepness of the stock-recruit relationship (Figure 38), virgin recruitment ( $R_{0}$; Figure 39), inclusion of the MARMAP Florida snapper trap index and composition data (Figure 40), time varying female maturity (Figure 41), upweighting the headboat index (Figure 42), replacing 2016 MRIP landings and discards with the average of adjacent years (Figure 43).

The qualitative results on terminal stock status were the same across all sensitivity runs, supporting the results of the base model that the stock is currently overfished ( $\left.\mathrm{SSB}_{2017}<\mathrm{MSST}\right)$. Mosts runs also supported the result of the base model, that overfishing is occurring (Figure 44, Table 30). The exceptions were runs $S 4$ (high steepness), $S 5$ (low virgin recruitment; $R_{0}$ ), and $S 9$ ( $3 \times$ weight on headboat index). In concert, sensitivity analyses were in general agreement with those of the MCB analysis.

### 4.13 Retrospective Analyses

Retrospective analyses did not suggest any patterns of substantial over- or underestimation in terminal-year estimates of $F / F_{\text {MSY }}$ or recruitment, but terminal values of $\mathrm{SSB} / \mathrm{MSST}$ and biomass $(B)$ consistently underestimated analogous values from the base run (Figure 45).

### 4.14 Projections

Projections results for Red Porgy are shown in Figures 46, 47, 48, and 49, and Tables 31, 32, 33, and 34. Among all scenarios considered, the Red Porgy stock exhibits a range of 0.5 to $6.2 \%$ probability of rebuilding by 2026. Thus under no management prescription, including $F=0$, is the Red Porgy population projected to have a $50 \%$ or greater chance of $\mathrm{SSB}>\mathrm{SSB}_{\mathrm{MSY}}$ by 2026. At $F=0$, the probability that $\mathrm{SSB}>\mathrm{SSB}_{\mathrm{MSY}}$ exceeds $50 \%$ in 2032 . However it is only theoretically possible to achieve $F=0$ owing to discard mortality that will inevitably occur by fisheries targeting other stocks.

## 5 Discussion

### 5.1 Comments on Assessment Results

Estimated benchmarks played a central role in this assessment. Values of MSST and $F_{\text {MSY }}$ were used to gauge the status of the stock and fishery. For rebuilding projections, SSB reaching $\mathrm{SSB}_{\text {MSY }}$ was the criterion that defined a successfully rebuilt stock. Computation of benchmarks was conditional on selectivity. If selectivity patterns change in the future, for example as a result of new size limits or different relative catch allocations among sectors, including discards, estimates of benchmarks would likely change as well.

The base run of the BAM indicated that the stock is not yet rebuilt $\left(\mathrm{SSB}_{2017} / \mathrm{SSB}_{\mathrm{MSY}}=0.27\right)$, and that overfishing is occurring $\left(F_{2015-2017} / F_{\mathrm{MSY}}=1.73\right)$. These results were generally consistent across sensitivity runs and MCB analyses, but with slightly more uncertainty in the overfishing status than in the stock status. Of the sensitivity runs conducted with the BAM, results were least sensitive to inclusion of the MARMAP Florida snapper trap data, smoothing of the 2016 value of MRIP landings and discards, including time varying female maturity, and $2 \times$ upweighting of the headboat index. Results were most sensitive to alternate values of steepness and $R_{0}$, and $3 \times$ upweighting of the headboat index, and somewhat sensitive to alternative values of natural mortality.

Low sensitivity to the MARMAP Florida snapper trap data is not very surprising as this index was a fairly short time series and the model didn't seem to fit it very well in the previous assessment. Lack of sensitivity to the 2016 value of MRIP landings and discards was reassuring since these values are large and have been scrutinized heavily at various steps in the SEDAR process. Counter to expectations, these values were not very influential. Sensitivity to steepness and natural mortality are common in stock assessment. In this assessment, likelihood profiles suggested that it was very unlikely for steepness to be much lower than the base run value, so the sensitivity run fixing steepness at a lower value, which resulted in much poorer SSB and $F$-status seems less likely than the run with higher steepness. The run with natural mortality set at a higher value, which resulted in much better SSB and slightly better $F$-status seems less likely than the run with a lower natural mortality, since estimates of natural mortality tend to get lower over time as older fish are discovered, and are less commonly found to be higher.

Model parameters and biological reference point estimates were similar in precision to other SEDAR assessments, but notably less precise than in the 2012 update. The increased uncertainty in the current assessment is probably due in part to a substantial increase in the uncertainty in natural mortality modeled in SEDAR $60(0.14 \leq M \leq 0.32)$ than in the 2012 update ( $0.20 \leq M \leq 0.25$ ).

South Atlantic Red Porgy is not currently rebuilt, which is consistent with projections made in SEDAR 1, 2002 Benchmark(SEDAR 2002), which estimated the probability of the stock being rebuilt by 2017 was $\approx 0.38$. The recent history of $F$ in the South Atlantic stock of Red Porgy is most similar to projection scenario 4 from the 2012 update (SEDAR 2012), with projections at $F_{2013-2017}=F_{\text {MSY }}=0.17$. The projection from the SEDAR 1,2012 Update estimated $\mathrm{SSB}_{2017} / \mathrm{SSB}_{\mathrm{MSY}}=0.57$ and the probability that $\mathrm{SSB}>\mathrm{SSB}_{\mathrm{MSY}}$ by 2017 to be 0.01. This 2012
projection also predicted that the stock would not be rebuilt by now, but substantially overestimated rebuilding status compared to the SEDAR 60 estimate $\left(\mathrm{SSB}_{2017} / \mathrm{SSB}_{\mathrm{MSY}}=0.27\right)$. The overly optimistic projection may have been partly due to the fact that recruitment has continued to decline since the 2012 update, so observed recruitment values for $2013-2017$ were lower than the average expected from the stock-recruit curve.

### 5.2 Comments on Projections

As usual, projections should be interpreted in light of the model assumptions and key aspects of the data. Some major considerations are the following:

- In general, projections of fish stocks are highly uncertain, particularly in the long term (e.g., beyond 5-10 years).
- Although projections included many major sources of uncertainty, they did not include structural (model) uncertainty. That is, projection results are conditional on one set of functional forms used to describe population dynamics, selectivity, recruitment, etc.
- Fisheries were assumed to continue fishing at their estimated current proportions of total effort, using the estimated current selectivity patterns. Benchmarks (e.g. MSY, $D_{\text {MSY }}$ ) are conditional on the estimated selectivity functions and the relative contributions of each fleet's fishing mortality. Selectivity patterns of landings and discards are different, and therefore projections of landings and discards are not interchangeable. New management regulations that reallocate harvest in a way that alters proportions of F by fleet or selectivity patterns would likely affect projection results.
- The projections assumed that the estimated spawner-recruit relationship applies in the future and that past residuals represent future uncertainty in recruitment. If future recruitment is characterized by runs of large or small year classes, possibly due to environmental or ecological conditions, stock trajectories may be affected.
- Projections apply the Baranov catch equation to relate $F$ and landings using a one-year time step, as in the assessment. The catch equation implicitly assumes that mortality occurs throughout the year. This assumption is violated when seasonal closures are in effect, introducing additional and unquantified uncertainty into the projection results.


## 6 Research Recommendations

- Investigate temporal trends in growth, sex at age, and female maturity at age. In the previous assessments, female maturity at age was estimated for several time blocks and included in the model as a time-varying relationship. During the current assessment process, the basis for modeling only female maturity as timevarying was called into question, given that life history parameters are often linked. The decision was made to use only a single female maturity at age relationship. However the panel judged this to be an important area of future research.


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## 8 Tables

Table 10. Observed time series of landings ( $L$ ) and dead discards ( $D$ ) for commercial handline (cHl), commercial trawl ( $c T w$ ), recreational headboat (rHb), and MRIP (rGe). Commercial landings are in units of 1000 lb whole weight. Recreational landings and all discards are in units of 1000 fish. Discards include fish released dead.

| Year | L.cHl | L.cTw | L.rHb | L.rGe | D.cHl | D.rHb | D.rGe |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 32.17 | 0.67 | 219.90 | 81.54 | . | . | . |
| 1973 | 14.65 | 12.95 | 299.60 | 81.54 | . | . | . |
| 1974 | 108.35 | 0.67 | 219.80 | 81.54 | . | . | . |
| 1975 | 197.74 | 1.16 | 215.50 | 81.54 |  | . |  |
| 1976 | 211.70 | 39.26 | 186.70 | 81.54 |  | . |  |
| 1977 | 288.62 | 148.47 | 243.60 | 81.54 |  | . |  |
| 1978 | 718.95 | 7.44 | 223.70 | 81.54 |  | . |  |
| 1979 | 983.55 | 83.11 | 156.50 | 81.54 |  | . |  |
| 1980 | 940.98 | 292.82 | 168.40 | 81.54 |  | . |  |
| 1981 | 1268.05 | 303.13 | 168.29 | 56.17 |  | . | 0.52 |
| 1982 | 1382.73 | 223.35 | 272.88 | 54.90 |  | . | 0.81 |
| 1983 | 1182.07 | 113.74 | 155.74 | 21.28 |  | . | 0.29 |
| 1984 | 1062.89 | 62.07 | 129.97 | 258.26 |  | . | 0.61 |
| 1985 | 847.75 | 15.83 | 176.58 | 195.67 |  | . | 6.17 |
| 1986 | 906.18 | 15.06 | 161.04 | 36.30 |  | . | 0.23 |
| 1987 | 777.44 | 9.68 | 173.57 | 70.60 | . | . | 9.28 |
| 1988 | 868.35 | 24.71 | 168.56 | 208.54 | . | . | 0.27 |
| 1989 | 924.36 | . | 146.49 | 138.49 | . | . | 8.64 |
| 1990 | 1138.58 | . | 104.76 | 107.43 | . | . | 0.08 |
| 1991 | 832.44 |  | 129.88 | 60.53 | . | . | 0.08 |
| 1992 | 516.53 |  | 85.89 | 158.62 |  | . | 5.34 |
| 1993 | 470.08 |  | 81.69 | 54.58 | . | . | 3.35 |
| 1994 | 436.36 |  | 70.39 | 70.32 | . | . | 1.06 |
| 1995 | 432.07 |  | 70.71 | 44.94 | . | . | 9.10 |
| 1996 | 429.61 |  | 64.91 | 66.00 |  |  | 2.87 |
| 1997 | 425.70 |  | 53.87 | 20.75 |  | . | 0.72 |
| 1998 | 317.98 |  | 53.88 | 31.10 | . | . | 2.66 |
| 1999 | 105.14 | . | 31.95 | 28.38 | 42.35 | . | 24.86 |
| 2000 | 26.21 | . | 8.04 | 10.56 | 46.47 | . | 9.27 |
| 2001 | 66.17 | . | 28.86 | 40.37 | 43.25 | 17.96 | 48.27 |
| 2002 | 58.17 | . | 20.93 | 57.61 | 133.00 | 13.02 | 25.38 |
| 2003 | 50.37 | . | 20.17 | 55.76 | 24.34 | 12.56 | 39.74 |
| 2004 | 49.68 | . | 23.46 | 74.54 | 20.72 | 26.13 | 37.63 |
| 2005 | 48.66 | . | 24.78 | 50.03 | 13.27 | 7.70 | 21.11 |
| 2006 | 83.81 |  | 40.22 | 29.97 | 21.32 | 17.75 | 3.37 |
| 2007 | 144.29 |  | 74.94 | 54.53 | 13.42 | 17.45 | 21.36 |
| 2008 | 171.96 |  | 32.52 | 107.07 | 21.30 | 11.41 | 46.17 |
| 2009 | 164.53 | . | 19.54 | 53.03 | 17.80 | 6.05 | 4.52 |
| 2010 | 158.82 | . | 21.92 | 29.24 | 11.14 | 5.30 | 9.52 |
| 2011 | 202.83 | . | 21.09 | 52.34 | 6.33 | 6.19 | 9.00 |
| 2012 | 162.26 | . | 23.22 | 49.75 | 14.75 | 6.93 | 3.58 |
| 2013 | 171.46 | . | 17.71 | 35.74 | 13.98 | 5.70 | 5.92 |
| 2014 | 158.14 | . | 17.17 | 23.71 | 14.90 | 7.32 | 14.59 |
| 2015 | 154.81 | . | 15.55 | 65.94 | 16.10 | 7.70 | 27.52 |
| 2016 | 127.44 |  | 15.31 | 278.49 | 8.89 | 6.34 | 112.32 |
| 2017 | 129.80 |  | 12.33 | 60.44 | 9.77 | 4.59 | 14.71 |

Table 11. Observed time series of CVs used in ensemble modeling (MCB) associated with landings (L) and discards (D) for commercial handline ( $c H l$ ), commercial trawl ( $c T w$ ), recreational headboat ( $r H b$ ), and MRIP (rGe). These CVs were used to generate bootstrap data sets in the ensemble model analysis only. When fitting the assessment model, CVs of 0.05 were used for estimating landings and discards, in all cases.

| Year | L.cHl | L.cTw | L.rHb | L.rGe | D.cHl | D.rHb | D.rGe |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 0.30 | 0.30 | 0.10 | 0.42 | . | . | . |
| 1973 | 0.30 | 0.30 | 0.10 | 0.42 | . | . |  |
| 1974 | 0.30 | 0.30 | 0.10 | 0.42 | . |  |  |
| 1975 | 0.30 | 0.30 | 0.10 | 0.42 | . | . |  |
| 1976 | 0.30 | 0.30 | 0.10 | 0.42 | . | . |  |
| 1977 | 0.30 | 0.30 | 0.10 | 0.42 | . | . |  |
| 1978 | 0.30 | 0.30 | 0.10 | 0.42 | . | . |  |
| 1979 | 0.30 | 0.30 | 0.10 | 0.42 | . | . |  |
| 1980 | 0.30 | 0.30 | 0.10 | 0.42 | . |  |  |
| 1981 | 0.30 | 0.30 | 0.05 | 0.57 | . | . | 1.00 |
| 1982 | 0.30 | 0.30 | 0.05 | 0.60 | . | . | 0.79 |
| 1983 | 0.30 | 0.30 | 0.05 | 0.52 | . |  | 1.00 |
| 1984 | 0.28 | 0.28 | 0.05 | 0.16 | . |  | 0.68 |
| 1985 | 0.25 | 0.25 | 0.05 | 0.50 | . |  | 0.81 |
| 1986 | 0.23 | 0.23 | 0.05 | 0.51 | . |  | 1.00 |
| 1987 | 0.21 | 0.21 | 0.05 | 0.46 | . |  | 0.69 |
| 1988 | 0.19 | 0.19 | 0.05 | 0.65 | . |  | 1.00 |
| 1989 | 0.16 | . | 0.05 | 0.29 | . |  | 1.00 |
| 1990 | 0.14 | . | 0.05 | 0.32 | . | . | 1.00 |
| 1991 | 0.12 | . | 0.05 | 0.32 | . | . | 1.00 |
| 1992 | 0.10 | . | 0.05 | 0.47 | . | . | 0.42 |
| 1993 | 0.07 | . | 0.05 | 0.29 | . | . | 0.55 |
| 1994 | 0.05 | . | 0.05 | 0.24 | . | . | 0.56 |
| 1995 | 0.05 | . | 0.05 | 0.39 | . | . | 0.57 |
| 1996 | 0.05 | . | 0.05 | 0.59 | . | . | 0.88 |
| 1997 | 0.05 | . | 0.05 | 0.38 | . | . | 1.00 |
| 1998 | 0.05 | . | 0.05 | 0.50 | . |  | 0.71 |
| 1999 | 0.05 | . | 0.05 | 0.30 | 0.10 |  | 0.66 |
| 2000 | 0.05 | . | 0.05 | 0.66 | 0.10 | . | 0.50 |
| 2001 | 0.05 | . | 0.05 | 0.24 | 0.10 | 0.10 | 0.34 |
| 2002 | 0.05 | . | 0.05 | 0.25 | 0.10 | 0.10 | 0.33 |
| 2003 | 0.05 | . | 0.05 | 0.28 | 0.10 | 0.10 | 0.42 |
| 2004 | 0.05 | . | 0.05 | 0.28 | 0.10 | 0.10 | 0.30 |
| 2005 | 0.05 | . | 0.05 | 0.40 | 0.10 | 0.10 | 0.54 |
| 2006 | 0.05 | . | 0.05 | 0.32 | 0.10 | 0.10 | 0.51 |
| 2007 | 0.05 | . | 0.05 | 0.31 | 0.10 | 0.10 | 0.45 |
| 2008 | 0.05 | . | 0.05 | 0.26 | 0.10 | 0.10 | 0.42 |
| 2009 | 0.05 | . | 0.05 | 0.35 | 0.10 | 0.10 | 0.85 |
| 2010 | 0.05 | . | 0.05 | 0.36 | 0.10 | 0.10 | 0.60 |
| 2011 | 0.05 | . | 0.05 | 0.48 | 0.10 | 0.10 | 0.69 |
| 2012 | 0.05 | . | 0.05 | 0.27 | 0.10 | 0.10 | 0.44 |
| 2013 | 0.05 | . | 0.05 | 0.45 | 0.10 | 0.10 | 0.42 |
| 2014 | 0.05 | . | 0.05 | 0.33 | 0.10 | 0.10 | 0.48 |
| 2015 | 0.05 | . | 0.05 | 0.45 | 0.10 | 0.10 | 0.47 |
| 2016 | 0.05 | . | 0.05 | 0.37 | 0.10 | 0.10 | 0.52 |
| 2017 | 0.05 | . | 0.05 | 0.41 | 0.10 | 0.10 | 0.57 |

Table 12. Sample sizes (number of trips) of length compositions (lcomp) or age compositions (acomp) by survey or fleet. Data sources are commercial trawl ( $c T w$ ), commercial handline ( $c \mathrm{Hl}$ ), recreational headboats (rHb), and SERFS chevron trap (sCT).

| Year | lcomp.cTw | acomp.cHl | acomp.rHb | acomp.sCT |
| :---: | :---: | :---: | :---: | :---: |
| 1972 | . | . | - |  |
| 1973 | . | . | . |  |
| 1974 | . | . | . |  |
| 1975 | . | . | . |  |
| 1976 | . | . | . |  |
| 1977 | 15 | . | . |  |
| 1978 | . | . |  |  |
| 1979 | . | . | 10 |  |
| 1980 | . | . | . |  |
| 1981 | . | . | . |  |
| 1982 | . | . | . |  |
| 1983 | . | . | 9 |  |
| 1984 | . | . | 17 |  |
| 1985 | . | . | 13 |  |
| 1986 | . | . | 15 |  |
| 1987 | . | . | 34 |  |
| 1988 | . | . | . |  |
| 1989 | . | . | . |  |
| 1990 | . | . | . | 138 |
| 1991 | . | . | 21 | 122 |
| 1992 | . | . | 7 | 96 |
| 1993 | . | . | . | 106 |
| 1994 | . | . | . | 86 |
| 1995 | . | . | . | 131 |
| 1996 | . | . | . | 207 |
| 1997 | . | 6 | . | 124 |
| 1998 | . | 8 | 58 | 155 |
| 1999 | . | . | . | 101 |
| 2000 | . | 8 | . | 127 |
| 2001 | . | 14 | 3 | 114 |
| 2002 | . | 7 | . | 118 |
| 2003 | . | 7 | . | 102 |
| 2004 | . | 42 | . | 153 |
| 2005 | . | 60 | 23 | 158 |
| 2006 | . | 172 | 25 | 119 |
| 2007 | . | 260 | 64 | 148 |
| 2008 | . | 264 | 26 | 96 |
| 2009 | . | 204 | 24 | 114 |
| 2010 | . | 158 | 21 | 191 |
| 2011 | . | 257 | 29 | 217 |
| 2012 | . | 193 | 23 | 295 |
| 2013 | . | 154 | 58 | 275 |
| 2014 | . | 141 | 38 | 307 |
| 2015 | . | 103 | 26 | 395 |
| 2016 | . | 125 | 45 | 400 |
| 2017 | . | 115 | 16 | 334 |

Table 13. Sample sizes (number of fish) of length compositions (lcomp) or age compositions (acomp) by survey or fleet. Data sources are commercial trawl ( $c T w$ ), commercial handline ( $c \mathrm{Hl}$ ), recreational headboats (rHb), and SERFS chevron trap (sCT).

| Year | lcomp.cTw | acomp.cHl | acomp.rHb | acomp.sCT |
| :---: | :---: | :---: | :---: | :---: |
| 1972 |  | . | . |  |
| 1973 |  | . |  |  |
| 1974 |  | . |  |  |
| 1975 | . | . |  |  |
| 1976 |  | . |  |  |
| 1977 | 2538 | . |  |  |
| 1978 |  | . |  |  |
| 1979 | . | . | 10 |  |
| 1980 |  | . | . |  |
| 1981 |  | . |  |  |
| 1982 | . | . |  |  |
| 1983 |  | . | 19 |  |
| 1984 |  | . | 30 |  |
| 1985 |  | . | 18 |  |
| 1986 |  | . | 28 |  |
| 1987 |  | . | 86 |  |
| 1988 | . | . | . |  |
| 1989 | . | . | . |  |
| 1990 |  | . | . | 953 |
| 1991 | . | . | 54 | 831 |
| 1992 | . | . | 12 | 1111 |
| 1993 |  | . | . | 722 |
| 1994 |  | . | . | 1115 |
| 1995 | . | . | . | 891 |
| 1996 | . | . | . | 1026 |
| 1997 | . | 309 | . | 601 |
| 1998 | . | 37 | 198 | 733 |
| 1999 | . | . | . | 470 |
| 2000 |  | 407 |  | 522 |
| 2001 | . | 307 | 10 | 720 |
| 2002 | . | 37 | . | 581 |
| 2003 | . | 75 | . | 491 |
| 2004 | . | 191 | . | 1084 |
| 2005 | . | 264 | 24 | 1115 |
| 2006 |  | 624 | 25 | 756 |
| 2007 |  | 1015 | 92 | 1154 |
| 2008 |  | 1227 | 26 | 411 |
| 2009 | . | 740 | 34 | 426 |
| 2010 | . | 678 | 29 | 785 |
| 2011 | . | 1070 | 49 | 1032 |
| 2012 |  | 723 | 77 | 1677 |
| 2013 |  | 578 | 188 | 1305 |
| 2014 |  | 670 | 168 | 1836 |
| 2015 |  | 482 | 122 | 1975 |
| 2016 | . | 506 | 128 | 1896 |
| 2017 |  | 456 | 44 | 1583 |

Table 14. Observed indices of abundance and CVs from recreational headboats (rHb) and SERFS chevron trap/video ( $s C T$ ).

| Year | rHb | sCT | cv.rHb | cv.sCT |
| :---: | :---: | :---: | :---: | :---: |
| 1972 | . |  |  |  |
| 1973 | 1.990 |  | 0.177 |  |
| 1974 | 1.994 |  | 0.156 |  |
| 1975 | 1.395 |  | 0.181 |  |
| 1976 | 1.175 |  | 0.134 |  |
| 1977 | 1.986 |  | 0.096 |  |
| 1978 | 2.825 |  | 0.063 |  |
| 1979 | 1.888 |  | 0.088 |  |
| 1980 | 1.905 |  | 0.088 | . |
| 1981 | 1.384 |  | 0.132 |  |
| 1982 | 1.388 |  | 0.137 |  |
| 1983 | 0.677 |  | 0.232 |  |
| 1984 | 0.673 |  | 0.232 |  |
| 1985 | 0.797 |  | 0.188 | . |
| 1986 | 1.055 |  | 0.126 |  |
| 1987 | 0.930 | . | 0.138 |  |
| 1988 | 0.718 |  | 0.188 |  |
| 1989 | 0.753 |  | 0.206 |  |
| 1990 | 0.426 | 0.87 | 0.332 | 0.16 |
| 1991 | 0.386 | 1.38 | 0.348 | 0.16 |
| 1992 | 0.310 | 1.34 | 0.349 | 0.16 |
| 1993 | 0.235 | 0.82 | 0.410 | 0.17 |
| 1994 | 0.237 | 0.96 | 0.406 | 0.16 |
| 1995 | 0.183 | 1.26 | 0.472 | 0.17 |
| 1996 | 0.222 | 0.87 | 0.424 | 0.16 |
| 1997 | 0.275 | 0.66 | 0.415 | 0.18 |
| 1998 | 0.195 | 0.73 | 0.447 | 0.17 |
| 1999 | . | 0.87 |  | 0.17 |
| 2000 |  | 0.81 |  | 0.19 |
| 2001 |  | 1.13 |  | 0.18 |
| 2002 |  | 1.01 |  | 0.19 |
| 2003 |  | 0.80 |  | 0.18 |
| 2004 |  | 1.41 |  | 0.16 |
| 2005 |  | 1.44 |  | 0.16 |
| 2006 |  | 1.00 |  | 0.18 |
| 2007 |  | 1.41 |  | 0.16 |
| 2008 |  | 0.72 |  | 0.19 |
| 2009 |  | 0.62 |  | 0.19 |
| 2010 |  | 1.04 | . | 0.16 |
| 2011 |  | 1.22 |  | 0.13 |
| 2012 |  | 1.21 |  | 0.11 |
| 2013 |  | 0.93 | . | 0.13 |
| 2014 |  | 1.07 |  | 0.11 |
| 2015 |  | 0.89 | . | 0.11 |
| 2016 |  | 0.79 |  | 0.11 |
| 2017 |  | 0.74 | . | 0.12 |

Table 15. Life-history characteristics at age. Variables include total length (TL) in millimeters (mm) and inches (in) and weight (mid-year), and inches (in), the coefficient of variation (CV) of TL, total weight (W) in kilograms (kg) and pounds (lb), proportion (P) female, and mature by sex, spawning stock biomass (SSB; sum product of the proportion and maturity of each sex and the average weight), and natural mortality. All values were fixed model input.

| Age | TL (mm) | TL (in) | TL CV | W (kg) | W (lb) | P (female) | P (mature female) | P (mature male) | SSB (kg) | M |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 249 | 9.8 | 0.14 | 0.23 | 0.51 | 0.82 | 0.30 | 1 | 0.10 | 0.46 |
| 2 | 294 | 11.6 | 0.14 | 0.38 | 0.83 | 0.71 | 0.69 | 1 | 0.29 | 0.36 |
| 3 | 327 | 12.9 | 0.14 | 0.51 | 1.13 | 0.57 | 0.92 | 1 | 0.49 | 0.30 |
| 4 | 352 | 13.9 | 0.14 | 0.63 | 1.40 | 0.42 | 0.98 | 1 | 0.63 | 0.27 |
| 5 | 370 | 14.6 | 0.14 | 0.73 | 1.62 | 0.28 | 1.00 | 1 | 0.73 | 0.25 |
| 6 | 384 | 15.1 | 0.14 | 0.81 | 1.79 | 0.17 | 1.00 | 1 | 0.81 | 0.24 |
| 7 | 394 | 15.5 | 0.14 | 0.88 | 1.93 | 0.10 | 1.00 | 1 | 0.88 | 0.23 |
| 8 | 401 | 15.8 | 0.14 | 0.92 | 2.04 | 0.06 | 1.00 | 1 | 0.92 | 0.22 |
| 9 | 407 | 16.0 | 0.14 | 0.96 | 2.12 | 0.03 | 1.00 | 1 | 0.96 | 0.22 |
| 10 | 411 | 16.2 | 0.14 | 0.99 | 2.18 | 0.02 | 1.00 | 1 | 0.99 | 0.22 |
| 11 | 414 | 16.3 | 0.14 | 1.01 | 2.23 | 0.01 | 1.00 | 1 | 1.01 | 0.21 |
| 12 | 416 | 16.4 | 0.14 | 1.03 | 2.26 | 0.00 | 1.00 | 1 | 1.03 | 0.21 |
| 13 | 418 | 16.5 | 0.14 | 1.04 | 2.29 | 0.00 | 1.00 | 1 | 1.04 | 0.21 |
| 14 | 419 | 16.5 | 0.14 | 1.05 | 2.31 | 0.00 | 1.00 | 1 | 1.05 | 0.21 |

Table 16. Estimated total abundance at age (1000 fish) at start of year.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 3331.19 | 1850.29 | 1216.43 | 859.43 | 629.90 | 471.37 | 356.91 | 272.17 | 208.35 | 159.96 | 122.94 | 94.64 | 72.83 | 231.71 | 9878.10 |
| 1973 | 3197.94 | 2083.78 | 1233.52 | 853.91 | 622.29 | 465.31 | 352.76 | 269.52 | 206.76 | 159.07 | 122.49 | 94.43 | 72.76 | 234.61 | 9969.15 |
| 1974 | 3212.00 | 1991.94 | 1373.71 | 856.88 | 611.85 | 454.90 | 344.60 | 263.60 | 202.61 | 156.22 | 120.55 | 93.10 | 71.84 | 234.33 | 9988.13 |
| 1975 | 1786.68 | 2009.70 | 1320.19 | 955.67 | 614.87 | 447.92 | 337.38 | 257.88 | 198.46 | 153.30 | 118.55 | 91.76 | 70.94 | 233.76 | 8597.05 |
| 1976 | 2885.46 | 1117.15 | 1318.54 | 905.80 | 676.32 | 443.93 | 327.62 | 249.00 | 191.47 | 148.09 | 114.74 | 89.00 | 68.95 | 229.42 | 8765.50 |
| 1977 | 10259.32 | 1799.14 | 728.07 | 897.53 | 635.96 | 484.44 | 322.14 | 239.89 | 183.42 | 141.75 | 109.96 | 85.46 | 66.35 | 222.89 | 16176.31 |
| 1978 | 3664.74 | 6360.90 | 1153.80 | 485.62 | 617.46 | 446.36 | 344.46 | 231.13 | 173.15 | 133.05 | 103.14 | 80.25 | 62.43 | 211.71 | 14068.19 |
| 1979 | 2019.83 | 2300.56 | 4069.89 | 757.97 | 329.03 | 426.81 | 312.57 | 243.39 | 164.30 | 123.70 | 95.34 | 74.13 | 57.73 | 197.62 | 11172.87 |
| 1980 | 3572.65 | 1262.61 | 1438.90 | 2592.57 | 497.97 | 220.53 | 289.81 | 214.16 | 167.77 | 113.82 | 85.95 | 66.44 | 51.71 | 178.49 | 10753.37 |
| 1981 | 3002.69 | 2194.05 | 758.73 | 877.31 | 1630.23 | 319.45 | 143.32 | 190.05 | 141.29 | 111.24 | 75.69 | 57.33 | 44.36 | 154.01 | 9699.76 |
| 1982 | 1978.11 | 1836.23 | 1248.10 | 428.28 | 510.68 | 968.13 | 192.19 | 87.01 | 116.07 | 86.72 | 68.48 | 46.74 | 35.44 | 122.86 | 7725.03 |
| 1983 | 2930.16 | 1200.32 | 971.17 | 641.83 | 227.10 | 276.27 | 530.59 | 106.28 | 48.41 | 64.90 | 48.63 | 38.52 | 26.32 | 89.31 | 7199.80 |
| 1984 | 2659.18 | 1805.36 | 644.95 | 502.55 | 342.46 | 123.62 | 152.35 | 295.25 | 59.50 | 27.23 | 36.62 | 27.53 | 21.82 | 65.64 | 6764.07 |
| 1985 | 2745.00 | 1616.16 | 902.06 | 307.82 | 247.31 | 171.93 | 62.88 | 78.19 | 152.44 | 30.87 | 14.17 | 19.12 | 14.38 | 45.79 | 6408.14 |
| 1986 | 2468.55 | 1676.24 | 823.93 | 442.51 | 155.70 | 127.62 | 89.88 | 33.17 | 41.49 | 81.30 | 16.52 | 7.60 | 10.27 | 32.39 | 6007.17 |
| 1987 | 2160.14 | 1530.30 | 870.84 | 405.75 | 224.68 | 80.65 | 66.97 | 47.60 | 17.67 | 22.22 | 43.66 | 8.90 | 4.10 | 23.04 | 5506.51 |
| 1988 | 3671.58 | 1331.01 | 791.59 | 430.53 | 206.83 | 116.85 | 42.49 | 35.60 | 25.45 | 9.50 | 11.98 | 23.61 | 4.81 | 14.72 | 6716.56 |
| 1989 | 2748.34 | 2227.10 | 618.23 | 341.65 | 191.57 | 93.89 | 53.74 | 19.72 | 16.62 | 11.94 | 4.47 | 5.65 | 11.15 | 9.25 | 6353.32 |
| 1990 | 2482.05 | 1692.42 | 1072.59 | 275.86 | 157.16 | 89.91 | 44.64 | 25.78 | 9.52 | 8.06 | 5.81 | 2.18 | 2.76 | 9.99 | 5878.72 |
| 1991 | 2200.15 | 1535.94 | 768.64 | 435.68 | 115.50 | 67.14 | 38.91 | 19.49 | 11.32 | 4.20 | 3.57 | 2.58 | 0.97 | 5.68 | 5209.77 |
| 1992 | 1447.94 | 1363.14 | 756.73 | 351.23 | 205.24 | 55.51 | 32.69 | 19.11 | 9.63 | 5.63 | 2.09 | 1.78 | 1.29 | 3.33 | 4255.36 |
| 1993 | 1432.21 | 911.61 | 729.45 | 361.23 | 171.98 | 102.51 | 28.09 | 16.69 | 9.82 | 4.97 | 2.91 | 1.09 | 0.93 | 2.41 | 3775.91 |
| 1994 | 1595.39 | 903.32 | 503.82 | 371.60 | 189.15 | 91.86 | 55.47 | 15.34 | 9.17 | 5.42 | 2.75 | 1.62 | 0.60 | 1.86 | 3747.38 |
| 1995 | 2210.05 | 1006.19 | 495.63 | 252.79 | 191.56 | 99.46 | 48.94 | 29.82 | 8.29 | 4.98 | 2.95 | 1.51 | 0.89 | 1.35 | 4354.41 |
| 1996 | 1116.35 | 1392.87 | 552.65 | 250.80 | 131.50 | 101.65 | 53.47 | 26.55 | 16.27 | 4.55 | 2.74 | 1.63 | 0.83 | 1.24 | 3653.08 |
| 1997 | 1310.35 | 704.27 | 780.74 | 287.48 | 134.10 | 71.73 | 56.17 | 29.81 | 14.89 | 9.17 | 2.57 | 1.55 | 0.93 | 1.18 | 3404.94 |
| 1998 | 1460.59 | 827.67 | 399.99 | 418.59 | 158.64 | 75.49 | 40.91 | 32.32 | 17.26 | 8.66 | 5.35 | 1.51 | 0.91 | 1.23 | 3449.11 |
| 1999 | 999.40 | 922.29 | 485.03 | 223.73 | 240.90 | 93.13 | 44.90 | 24.55 | 19.52 | 10.47 | 5.27 | 3.27 | 0.92 | 1.31 | 3074.70 |
| 2000 | 552.64 | 629.95 | 616.06 | 318.84 | 139.64 | 148.08 | 57.58 | 27.98 | 15.39 | 12.29 | 6.62 | 3.34 | 2.07 | 1.42 | 2531.90 |
| 2001 | 1141.59 | 349.45 | 427.19 | 430.26 | 225.06 | 99.63 | 106.83 | 41.90 | 20.48 | 11.32 | 9.07 | 4.90 | 2.48 | 2.59 | 2872.76 |
| 2002 | 1696.23 | 715.08 | 228.63 | 281.00 | 274.89 | 142.37 | 63.44 | 68.57 | 27.05 | 13.29 | 7.37 | 5.92 | 3.20 | 3.32 | 3530.36 |
| 2003 | 1236.74 | 1067.84 | 453.60 | 142.10 | 168.05 | 161.78 | 84.23 | 37.82 | 41.12 | 16.30 | 8.03 | 4.47 | 3.59 | 3.96 | 3429.65 |
| 2004 | 1242.61 | 778.02 | 716.64 | 306.81 | 92.37 | 107.32 | 103.80 | 54.46 | 24.60 | 26.88 | 10.69 | 5.28 | 2.94 | 4.98 | 3477.40 |
| 2005 | 1507.88 | 780.47 | 520.38 | 482.67 | 196.86 | 57.83 | 67.39 | 65.67 | 34.66 | 15.73 | 17.24 | 6.88 | 3.40 | 5.11 | 3762.18 |
| 2006 | 1781.72 | 952.25 | 534.01 | 362.98 | 328.50 | 132.66 | 39.22 | 46.08 | 45.16 | 23.95 | 10.91 | 11.99 | 4.79 | 5.94 | 4280.16 |
| 2007 | 1199.71 | 1126.60 | 652.19 | 370.23 | 245.31 | 220.55 | 89.73 | 26.74 | 31.60 | 31.13 | 16.56 | 7.56 | 8.32 | 7.46 | 4033.70 |
| 2008 | 817.48 | 756.72 | 765.39 | 437.76 | 232.43 | 149.91 | 135.18 | 55.40 | 16.61 | 19.72 | 19.49 | 10.40 | 4.75 | 9.94 | 3431.18 |
| 2009 | 1214.85 | 513.99 | 506.61 | 501.58 | 265.68 | 136.80 | 88.42 | 80.32 | 33.11 | 9.98 | 11.88 | 11.78 | 6.29 | 8.90 | 3390.19 |
| 2010 | 1927.83 | 768.74 | 352.28 | 345.04 | 327.72 | 172.15 | 89.29 | 58.18 | 53.16 | 22.03 | 6.66 | 7.95 | 7.89 | 10.20 | 4149.11 |
| 2011 | 1851.05 | 1219.60 | 528.05 | 241.70 | 229.40 | 217.52 | 115.27 | 60.29 | 39.52 | 36.29 | 15.08 | 4.57 | 5.47 | 12.46 | 4576.28 |
| 2012 | 1156.83 | 1170.93 | 837.66 | 356.33 | 154.35 | 144.84 | 138.28 | 73.87 | 38.86 | 25.60 | 23.58 | 9.83 | 2.98 | 11.72 | 4145.67 |
| 2013 | 620.61 | 732.33 | 805.89 | 574.32 | 234.53 | 100.72 | 95.19 | 91.62 | 49.24 | 26.03 | 17.20 | 15.89 | 6.63 | 9.93 | 3380.14 |
| 2014 | 839.32 | 392.85 | 504.39 | 557.22 | 387.00 | 158.05 | 68.50 | 65.29 | 63.21 | 34.14 | 18.10 | 12.00 | 11.10 | 11.59 | 3122.76 |
| 2015 | 613.05 | 530.48 | 269.11 | 349.08 | 379.79 | 265.10 | 109.37 | 47.81 | 45.84 | 44.61 | 24.16 | 12.85 | 8.53 | 16.15 | 2715.93 |
| 2016 | 782.93 | 386.39 | 359.14 | 182.15 | 227.89 | 246.09 | 173.05 | 71.98 | 31.65 | 30.50 | 29.77 | 16.17 | 8.61 | 16.57 | 2562.89 |
| 2017 | 345.47 | 483.87 | 243.81 | 214.20 | 91.23 | 103.61 | 110.46 | 78.06 | 32.64 | 14.42 | 13.94 | 13.65 | 7.42 | 11.58 | 1764.35 |

Table 17. Estimated total abundance at age (mt) at start of year.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 775.76 | 696.15 | 624.15 | 544.09 | 461.65 | 383.27 | 312.68 | 251.69 | 200.41 | 158.37 | 124.33 | 97.21 | 75.68 | 242.81 | 4948.24 |
| 1973 | 744.73 | 784.00 | 632.92 | 540.60 | 456.07 | 378.34 | 309.04 | 249.23 | 198.89 | 157.49 | 123.87 | 96.99 | 75.60 | 245.85 | 4993.64 |
| 1974 | 748.00 | 749.45 | 704.86 | 542.48 | 448.42 | 369.88 | 301.89 | 243.76 | 194.89 | 154.66 | 121.90 | 95.63 | 74.65 | 245.56 | 4996.03 |
| 1975 | 416.08 | 756.13 | 677.39 | 605.02 | 450.63 | 364.20 | 295.57 | 238.47 | 190.90 | 151.78 | 119.89 | 94.25 | 73.71 | 244.95 | 4678.97 |
| 1976 | 671.96 | 420.32 | 676.55 | 573.45 | 495.67 | 360.96 | 287.02 | 230.26 | 184.18 | 146.62 | 116.03 | 91.42 | 71.64 | 240.41 | 4566.49 |
| 1977 | 2389.16 | 676.91 | 373.57 | 568.21 | 466.09 | 393.90 | 282.22 | 221.84 | 176.43 | 140.34 | 111.20 | 87.78 | 68.94 | 233.57 | 6190.16 |
| 1978 | 853.43 | 2393.22 | 592.02 | 307.44 | 452.53 | 362.93 | 301.77 | 213.73 | 166.56 | 131.73 | 104.30 | 82.43 | 64.86 | 221.86 | 6248.81 |
| 1979 | 470.37 | 865.56 | 2088.27 | 479.86 | 241.14 | 347.04 | 273.84 | 225.08 | 158.04 | 122.47 | 96.42 | 76.14 | 59.99 | 207.09 | 5711.30 |
| 1980 | 831.99 | 475.04 | 738.30 | 1641.32 | 364.95 | 179.31 | 253.90 | 198.04 | 161.38 | 112.69 | 86.92 | 68.25 | 53.73 | 187.04 | 5352.86 |
| 1981 | 699.26 | 825.49 | 389.31 | 555.41 | 1194.78 | 259.75 | 125.56 | 175.75 | 135.91 | 110.13 | 76.54 | 58.89 | 46.10 | 161.39 | 4814.26 |
| 1982 | 460.66 | 690.86 | 640.40 | 271.14 | 374.27 | 787.19 | 168.38 | 80.46 | 111.65 | 85.86 | 69.25 | 48.01 | 36.82 | 128.75 | 3953.68 |
| 1983 | 682.37 | 451.61 | 498.31 | 406.33 | 166.44 | 224.63 | 464.84 | 98.29 | 46.56 | 64.25 | 49.18 | 39.57 | 27.34 | 93.59 | 3313.30 |
| 1984 | 619.26 | 679.25 | 330.93 | 318.16 | 250.98 | 100.52 | 133.47 | 273.03 | 57.23 | 26.96 | 37.03 | 28.27 | 22.68 | 68.78 | 2946.56 |
| 1985 | 639.25 | 608.06 | 462.85 | 194.88 | 181.25 | 139.80 | 55.09 | 72.31 | 146.63 | 30.57 | 14.33 | 19.64 | 14.95 | 47.99 | 2627.58 |
| 1986 | 574.87 | 630.67 | 422.76 | 280.15 | 114.11 | 103.77 | 78.75 | 30.67 | 39.91 | 80.49 | 16.70 | 7.81 | 10.67 | 33.94 | 2425.27 |
| 1987 | 503.05 | 575.76 | 446.83 | 256.87 | 164.67 | 65.58 | 58.67 | 44.01 | 17.00 | 21.99 | 44.15 | 9.14 | 4.26 | 24.15 | 2236.12 |
| 1988 | 855.03 | 500.78 | 406.17 | 272.56 | 151.58 | 95.01 | 37.23 | 32.92 | 24.49 | 9.40 | 12.11 | 24.25 | 5.00 | 15.43 | 2441.95 |
| 1989 | 640.03 | 837.92 | 317.21 | 216.29 | 140.40 | 76.34 | 47.08 | 18.23 | 15.99 | 11.82 | 4.52 | 5.81 | 11.59 | 9.69 | 2352.92 |
| 1990 | 578.01 | 636.76 | 550.35 | 174.64 | 115.18 | 73.10 | 39.11 | 23.84 | 9.15 | 7.98 | 5.88 | 2.24 | 2.87 | 10.46 | 2229.58 |
| 1991 | 512.36 | 577.88 | 394.39 | 275.82 | 84.65 | 54.59 | 34.09 | 18.03 | 10.89 | 4.16 | 3.61 | 2.65 | 1.01 | 5.95 | 1980.08 |
| 1992 | 337.19 | 512.87 | 388.28 | 222.36 | 150.42 | 45.14 | 28.64 | 17.68 | 9.27 | 5.57 | 2.12 | 1.83 | 1.34 | 3.49 | 1726.19 |
| 1993 | 333.53 | 342.98 | 374.28 | 228.69 | 126.05 | 83.35 | 24.61 | 15.43 | 9.44 | 4.92 | 2.95 | 1.12 | 0.96 | 2.52 | 1550.85 |
| 1994 | 371.53 | 339.87 | 258.51 | 235.25 | 138.62 | 74.69 | 48.60 | 14.18 | 8.82 | 5.37 | 2.79 | 1.66 | 0.63 | 1.95 | 1502.47 |
| 1995 | 514.67 | 378.57 | 254.31 | 160.04 | 140.39 | 80.87 | 42.87 | 27.58 | 7.98 | 4.93 | 2.99 | 1.55 | 0.92 | 1.42 | 1619.08 |
| 1996 | 259.97 | 524.05 | 283.57 | 158.78 | 96.37 | 82.65 | 46.84 | 24.55 | 15.65 | 4.50 | 2.77 | 1.67 | 0.86 | 1.30 | 1503.54 |
| 1997 | 305.15 | 264.97 | 400.60 | 182.00 | 98.28 | 58.32 | 49.21 | 27.57 | 14.32 | 9.08 | 2.60 | 1.60 | 0.96 | 1.23 | 1415.90 |
| 1998 | 340.14 | 311.40 | 205.23 | 265.00 | 116.26 | 61.38 | 35.84 | 29.89 | 16.60 | 8.58 | 5.41 | 1.55 | 0.95 | 1.29 | 1399.53 |
| 1999 | 232.74 | 347.00 | 248.87 | 141.64 | 176.55 | 75.73 | 39.34 | 22.70 | 18.77 | 10.37 | 5.33 | 3.36 | 0.96 | 1.38 | 1324.73 |
| 2000 | 128.70 | 237.01 | 316.10 | 201.85 | 102.34 | 120.41 | 50.44 | 25.87 | 14.80 | 12.17 | 6.69 | 3.43 | 2.15 | 1.49 | 1223.47 |
| 2001 | 265.85 | 131.48 | 219.19 | 272.39 | 164.94 | 81.01 | 93.59 | 38.75 | 19.70 | 11.21 | 9.17 | 5.03 | 2.57 | 2.72 | 1317.61 |
| 2002 | 395.01 | 269.04 | 117.31 | 177.89 | 201.46 | 115.76 | 55.58 | 63.41 | 26.02 | 13.16 | 7.45 | 6.08 | 3.33 | 3.48 | 1455.00 |
| 2003 | 288.01 | 401.76 | 232.74 | 89.96 | 123.16 | 131.54 | 73.79 | 34.98 | 39.56 | 16.14 | 8.12 | 4.59 | 3.73 | 4.15 | 1452.25 |
| 2004 | 289.38 | 292.72 | 367.71 | 194.24 | 67.70 | 87.26 | 90.94 | 50.36 | 23.66 | 26.61 | 10.81 | 5.43 | 3.05 | 5.22 | 1515.09 |
| 2005 | 351.15 | 293.64 | 267.01 | 305.57 | 144.28 | 47.02 | 59.04 | 60.73 | 33.34 | 15.58 | 17.43 | 7.06 | 3.53 | 5.36 | 1610.75 |
| 2006 | 414.92 | 358.27 | 274.00 | 229.80 | 240.76 | 107.87 | 34.36 | 42.61 | 43.44 | 23.71 | 11.03 | 12.31 | 4.97 | 6.22 | 1804.28 |
| 2007 | 279.39 | 423.87 | 334.64 | 234.39 | 179.79 | 179.33 | 78.61 | 24.73 | 30.40 | 30.82 | 16.75 | 7.77 | 8.65 | 7.82 | 1836.94 |
| 2008 | 190.37 | 284.71 | 392.72 | 277.14 | 170.35 | 121.89 | 118.43 | 51.23 | 15.98 | 19.53 | 19.71 | 10.68 | 4.94 | 10.41 | 1688.08 |
| 2009 | 282.91 | 193.38 | 259.94 | 317.54 | 194.71 | 111.23 | 77.47 | 74.28 | 31.85 | 9.88 | 12.02 | 12.10 | 6.53 | 9.33 | 1593.17 |
| 2010 | 448.95 | 289.23 | 180.76 | 218.44 | 240.18 | 139.97 | 78.22 | 53.81 | 51.14 | 21.81 | 6.73 | 8.17 | 8.20 | 10.69 | 1756.28 |
| 2011 | 431.07 | 458.86 | 270.95 | 153.02 | 168.13 | 176.86 | 100.99 | 55.75 | 38.02 | 35.93 | 15.25 | 4.69 | 5.68 | 13.05 | 1928.25 |
| 2012 | 269.40 | 440.55 | 429.81 | 225.59 | 113.12 | 117.77 | 121.14 | 68.31 | 37.38 | 25.35 | 23.85 | 10.10 | 3.10 | 12.28 | 1897.74 |
| 2013 | 144.53 | 275.53 | 413.51 | 363.59 | 171.88 | 81.90 | 83.40 | 84.73 | 47.36 | 25.77 | 17.40 | 16.32 | 6.89 | 10.41 | 1743.21 |
| 2014 | 195.46 | 147.81 | 258.81 | 352.77 | 283.63 | 128.51 | 60.01 | 60.38 | 60.80 | 33.80 | 18.31 | 12.32 | 11.53 | 12.14 | 1636.27 |
| 2015 | 142.77 | 199.59 | 138.08 | 220.99 | 278.34 | 215.55 | 95.82 | 44.21 | 44.10 | 44.16 | 24.44 | 13.20 | 8.86 | 16.93 | 1487.04 |
| 2016 | 182.33 | 145.37 | 184.28 | 115.32 | 167.02 | 200.09 | 151.61 | 66.56 | 30.45 | 30.20 | 30.10 | 16.61 | 8.95 | 17.36 | 1346.24 |
| 2017 | 80.45 | 182.05 | 125.10 | 135.61 | 66.86 | 84.25 | 96.77 | 72.18 | 31.40 | 14.28 | 14.10 | 14.02 | 7.71 | 12.13 | 936.90 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 1710.20 | 1534.70 | 1376.00 | 1199.50 | 1017.80 | 845.00 | 689.30 | 554.90 | 441.80 | 349.10 | 274.10 | 214.30 | 166.80 | 535.30 | 10908.90 |
| 1973 | 1641.80 | 1728.40 | 1395.30 | 1191.80 | 1005.50 | 834.10 | 681.30 | 549.50 | 438.50 | 347.20 | 273.10 | 213.80 | 166.70 | 542.00 | 11009.00 |
| 1974 | 1649.00 | 1652.20 | 1553.90 | 1196.00 | 988.60 | 815.40 | 665.50 | 537.40 | 429.70 | 341.00 | 268.70 | 210.80 | 164.60 | 541.40 | 11014.20 |
| 1975 | 917.30 | 1667.00 | 1493.40 | 1333.80 | 993.50 | 802.90 | 651.60 | 525.70 | 420.90 | 334.60 | 264.30 | 207.80 | 162.50 | 540.00 | 10315.30 |
| 1976 | 1481.40 | 926.60 | 1491.50 | 1264.20 | 1092.80 | 795.80 | 632.80 | 507.60 | 406.00 | 323.20 | 255.80 | 201.50 | 157.90 | 530.00 | 10067.30 |
| 1977 | 5267.10 | 1492.30 | 823.60 | 1252.70 | 1027.50 | 868.40 | 622.20 | 489.10 | 389.00 | 309.40 | 245.20 | 193.50 | 152.00 | 514.90 | 13646.80 |
| 1978 | 1881.50 | 5276.10 | 1305.20 | 677.80 | 997.60 | 800.10 | 665.30 | 471.20 | 367.20 | 290.40 | 229.90 | 181.70 | 143.00 | 489.10 | 13776.10 |
| 1979 | 1037.00 | 1908.20 | 4603.80 | 1057.90 | 531.60 | 765.10 | 603.70 | 496.20 | 348.40 | 270.00 | 212.60 | 167.90 | 132.30 | 456.60 | 12591.10 |
| 1980 | 1834.20 | 1047.30 | 1627.70 | 3618.50 | 804.60 | 395.30 | 559.70 | 436.60 | 355.80 | 248.40 | 191.60 | 150.50 | 118.50 | 412.30 | 11800.90 |
| 1981 | 1541.60 | 1819.90 | 858.30 | 1224.50 | 2634.00 | 572.60 | 276.80 | 387.50 | 299.60 | 242.80 | 168.70 | 129.80 | 101.60 | 355.80 | 10613.50 |
| 1982 | 1015.60 | 1523.10 | 1411.80 | 597.80 | 825.10 | 1735.40 | 371.20 | 177.40 | 246.10 | 189.30 | 152.70 | 105.80 | 81.20 | 283.80 | 8716.30 |
| 1983 | 1504.40 | 995.60 | 1098.60 | 895.80 | 366.90 | 495.20 | 1024.80 | 216.70 | 102.60 | 141.60 | 108.40 | 87.20 | 60.30 | 206.30 | 7304.50 |
| 1984 | 1365.20 | 1497.50 | 729.60 | 701.40 | 553.30 | 221.60 | 294.20 | 601.90 | 126.20 | 59.40 | 81.60 | 62.30 | 50.00 | 151.60 | 6496.00 |
| 1985 | 1409.30 | 1340.50 | 1020.40 | 429.60 | 399.60 | 308.20 | 121.50 | 159.40 | 323.30 | 67.40 | 31.60 | 43.30 | 33.00 | 105.80 | 5792.80 |
| 1986 | 1267.40 | 1390.40 | 932.00 | 617.60 | 251.60 | 228.80 | 173.60 | 67.60 | 88.00 | 177.40 | 36.80 | 17.20 | 23.50 | 74.80 | 5346.80 |
| 1987 | 1109.00 | 1269.30 | 985.10 | 566.30 | 363.00 | 144.60 | 129.30 | 97.00 | 37.50 | 48.50 | 97.30 | 20.20 | 9.40 | 53.20 | 4929.80 |
| 1988 | 1885.00 | 1104.00 | 895.40 | 600.90 | 334.20 | 209.50 | 82.10 | 72.60 | 54.00 | 20.70 | 26.70 | 53.50 | 11.00 | 34.00 | 5383.50 |
| 1989 | 1411.00 | 1847.30 | 699.30 | 476.80 | 309.50 | 168.30 | 103.80 | 40.20 | 35.30 | 26.10 | 10.00 | 12.80 | 25.60 | 21.40 | 5187.20 |
| 1990 | 1274.30 | 1403.80 | 1213.30 | 385.00 | 253.90 | 161.20 | 86.20 | 52.60 | 20.20 | 17.60 | 13.00 | 4.90 | 6.30 | 23.10 | 4915.30 |
| 1991 | 1129.50 | 1274.00 | 869.50 | 608.10 | 186.60 | 120.30 | 75.20 | 39.70 | 24.00 | 9.20 | 8.00 | 5.80 | 2.20 | 13.10 | 4365.30 |
| 1992 | 743.40 | 1130.70 | 856.00 | 490.20 | 331.60 | 99.50 | 63.10 | 39.00 | 20.40 | 12.30 | 4.70 | 4.00 | 3.00 | 7.70 | 3805.60 |
| 1993 | 735.30 | 756.10 | 825.10 | 504.20 | 277.90 | 183.80 | 54.30 | 34.00 | 20.80 | 10.80 | 6.50 | 2.50 | 2.10 | 5.60 | 3419.00 |
| 1994 | 819.10 | 749.30 | 569.90 | 518.60 | 305.60 | 164.70 | 107.10 | 31.30 | 19.40 | 11.80 | 6.20 | 3.70 | 1.40 | 4.30 | 3312.30 |
| 1995 | 1134.60 | 834.60 | 560.70 | 352.80 | 309.50 | 178.30 | 94.50 | 60.80 | 17.60 | 10.90 | 6.60 | 3.40 | 2.00 | 3.10 | 3569.40 |
| 1996 | 573.10 | 1155.30 | 625.20 | 350.00 | 212.50 | 182.20 | 103.30 | 54.10 | 34.50 | 9.90 | 6.10 | 3.70 | 1.90 | 2.90 | 3314.70 |
| 1997 | 672.70 | 584.20 | 883.20 | 401.20 | 216.70 | 128.60 | 108.50 | 60.80 | 31.60 | 20.00 | 5.70 | 3.50 | 2.10 | 2.70 | 3121.50 |
| 1998 | 749.90 | 686.50 | 452.50 | 584.20 | 256.30 | 135.30 | 79.00 | 65.90 | 36.60 | 18.90 | 11.90 | 3.40 | 2.10 | 2.80 | 3085.40 |
| 1999 | 513.10 | 765.00 | 548.70 | 312.30 | 389.20 | 167.00 | 86.70 | 50.00 | 41.40 | 22.90 | 11.80 | 7.40 | 2.10 | 3.00 | 2920.50 |
| 2000 | 283.70 | 522.50 | 696.90 | 445.00 | 225.60 | 265.50 | 111.20 | 57.00 | 32.60 | 26.80 | 14.70 | 7.60 | 4.70 | 3.30 | 2697.30 |
| 2001 | 586.10 | 289.90 | 483.20 | 600.50 | 363.60 | 178.60 | 206.30 | 85.40 | 43.40 | 24.70 | 20.20 | 11.10 | 5.70 | 6.00 | 2904.80 |
| 2002 | 870.80 | 593.10 | 258.60 | 392.20 | 444.10 | 255.20 | 122.50 | 139.80 | 57.40 | 29.00 | 16.40 | 13.40 | 7.30 | 7.70 | 3207.70 |
| 2003 | 634.90 | 885.70 | 513.10 | 198.30 | 271.50 | 290.00 | 162.70 | 77.10 | 87.20 | 35.60 | 17.90 | 10.10 | 8.20 | 9.10 | 3201.60 |
| 2004 | 638.00 | 645.30 | 810.70 | 428.20 | 149.30 | 192.40 | 200.50 | 111.00 | 52.20 | 58.70 | 23.80 | 12.00 | 6.70 | 11.50 | 3340.20 |
| 2005 | 774.10 | 647.40 | 588.70 | 673.70 | 318.10 | 103.70 | 130.20 | 133.90 | 73.50 | 34.30 | 38.40 | 15.60 | 7.80 | 11.80 | 3551.10 |
| 2006 | 914.70 | 789.80 | 604.10 | 506.60 | 530.80 | 237.80 | 75.80 | 93.90 | 95.80 | 52.30 | 24.30 | 27.10 | 11.00 | 13.70 | 3977.70 |
| 2007 | 615.90 | 934.50 | 737.70 | 516.70 | 396.40 | 395.40 | 173.30 | 54.50 | 67.00 | 67.90 | 36.90 | 17.10 | 19.10 | 17.20 | 4049.70 |
| 2008 | 419.70 | 627.70 | 865.80 | 611.00 | 375.60 | 268.70 | 261.10 | 112.90 | 35.20 | 43.10 | 43.50 | 23.50 | 10.90 | 22.90 | 3721.50 |
| 2009 | 623.70 | 426.30 | 573.10 | 700.00 | 429.30 | 245.20 | 170.80 | 163.80 | 70.20 | 21.80 | 26.50 | 26.70 | 14.40 | 20.60 | 3512.30 |
| 2010 | 989.80 | 637.60 | 398.50 | 481.60 | 529.50 | 308.60 | 172.40 | 118.60 | 112.70 | 48.10 | 14.80 | 18.00 | 18.10 | 23.60 | 3871.90 |
| 2011 | 950.30 | 1011.60 | 597.30 | 337.30 | 370.70 | 389.90 | 222.60 | 122.90 | 83.80 | 79.20 | 33.60 | 10.30 | 12.50 | 28.80 | 4251.00 |
| 2012 | 593.90 | 971.20 | 947.60 | 497.30 | 249.40 | 259.60 | 267.10 | 150.60 | 82.40 | 55.90 | 52.60 | 22.30 | 6.80 | 27.10 | 4183.80 |
| 2013 | 318.60 | 607.40 | 911.60 | 801.60 | 378.90 | 180.60 | 183.90 | 186.80 | 104.40 | 56.80 | 38.40 | 36.00 | 15.20 | 22.90 | 3843.10 |
| 2014 | 430.90 | 325.90 | 570.60 | 777.70 | 625.30 | 283.30 | 132.30 | 133.10 | 134.00 | 74.50 | 40.40 | 27.20 | 25.40 | 26.80 | 3607.30 |
| 2015 | 314.80 | 440.00 | 304.40 | 487.20 | 613.60 | 475.20 | 211.20 | 97.50 | 97.20 | 97.40 | 53.90 | 29.10 | 19.50 | 37.30 | 3278.30 |
| 2016 | 402.00 | 320.50 | 406.30 | 254.20 | 368.20 | 441.10 | 334.20 | 146.70 | 67.10 | 66.60 | 66.40 | 36.60 | 19.70 | 38.30 | 2967.90 |
| 2017 | 177.40 | 401.30 | 275.80 | 299.00 | 147.40 | 185.70 | 213.30 | 159.10 | 69.20 | 31.50 | 31.10 | 30.90 | 17.00 | 26.70 | 2065.50 |

Table 19. Estimated time series of status indicators. Fishing mortality rate is apical $F$, which includes discard mortalities. Total and spawning stock biomass ( $B$ and SSB, mt) are at the start of the year. The MSST is defined by MSST $=(1-M) \mathrm{SSB}_{\mathrm{MSY}}$, with constant $M=0.22 . S P R$ is static spawning potential ratio.

| Year | $F$ | $F / F_{\text {MSY }}$ | B | $B / B_{\text {unfished }}$ | SSB | $\mathrm{SSB} / \mathrm{SSB}_{\mathrm{MSY}}$ | SSB/MSST | SPR |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 0.0519 | 0.288 | 4948 | 0.632 | 4089 | 1.418 | 1.818 | 0.778 |
| 1973 | 0.0623 | 0.346 | 4994 | 0.638 | 4125 | 1.430 | 1.834 | 0.743 |
| 1974 | 0.0609 | 0.338 | 4996 | 0.638 | 4130 | 1.432 | 1.836 | 0.752 |
| 1975 | 0.0747 | 0.415 | 4679 | 0.598 | 4003 | 1.388 | 1.780 | 0.715 |
| 1976 | 0.0827 | 0.459 | 4566 | 0.583 | 3825 | 1.327 | 1.701 | 0.693 |
| 1977 | 0.1030 | 0.572 | 6190 | 0.791 | 4367 | 1.515 | 1.942 | 0.645 |
| 1978 | 0.1183 | 0.657 | 6249 | 0.798 | 4868 | 1.688 | 2.164 | 0.627 |
| 1979 | 0.1491 | 0.828 | 5711 | 0.730 | 4799 | 1.664 | 2.134 | 0.572 |
| 1980 | 0.1929 | 1.072 | 5353 | 0.684 | 4380 | 1.519 | 1.947 | 0.501 |
| 1981 | 0.2701 | 1.501 | 4814 | 0.615 | 3858 | 1.338 | 1.715 | 0.425 |
| 1982 | 0.3634 | 2.019 | 3954 | 0.505 | 3166 | 1.098 | 1.408 | 0.360 |
| 1983 | 0.3572 | 1.984 | 3313 | 0.423 | 2533 | 0.878 | 1.126 | 0.371 |
| 1984 | 0.4381 | 2.434 | 2947 | 0.377 | 2170 | 0.752 | 0.965 | 0.325 |
| 1985 | 0.4106 | 2.281 | 2628 | 0.336 | 1891 | 0.656 | 0.841 | 0.339 |
| 1986 | 0.4068 | 2.260 | 2425 | 0.310 | 1740 | 0.604 | 0.774 | 0.348 |
| 1987 | 0.4028 | 2.238 | 2236 | 0.286 | 1616 | 0.560 | 0.719 | 0.347 |
| 1988 | 0.5388 | 2.993 | 2442 | 0.312 | 1613 | 0.559 | 0.717 | 0.290 |
| 1989 | 0.5055 | 2.808 | 2353 | 0.301 | 1587 | 0.550 | 0.706 | 0.306 |
| 1990 | 0.5996 | 3.331 | 2230 | 0.285 | 1518 | 0.526 | 0.675 | 0.281 |
| 1991 | 0.4817 | 2.676 | 1980 | 0.253 | 1370 | 0.475 | 0.609 | 0.317 |
| 1992 | 0.4432 | 2.462 | 1726 | 0.221 | 1254 | 0.435 | 0.558 | 0.349 |
| 1993 | 0.3761 | 2.089 | 1551 | 0.198 | 1143 | 0.396 | 0.508 | 0.379 |
| 1994 | 0.3917 | 2.176 | 1502 | 0.192 | 1084 | 0.376 | 0.482 | 0.372 |
| 1995 | 0.3827 | 2.126 | 1619 | 0.207 | 1112 | 0.386 | 0.494 | 0.375 |
| 1996 | 0.3551 | 1.973 | 1504 | 0.192 | 1110 | 0.385 | 0.493 | 0.391 |
| 1997 | 0.3236 | 1.798 | 1416 | 0.181 | 1057 | 0.367 | 0.470 | 0.409 |
| 1998 | 0.2816 | 1.564 | 1400 | 0.179 | 1030 | 0.357 | 0.458 | 0.440 |
| 1999 | 0.2442 | 1.357 | 1325 | 0.169 | 1024 | 0.355 | 0.455 | 0.555 |
| 2000 | 0.0889 | 0.494 | 1223 | 0.156 | 1019 | 0.353 | 0.453 | 0.733 |
| 2001 | 0.2145 | 1.192 | 1318 | 0.168 | 1045 | 0.362 | 0.465 | 0.562 |
| 2002 | 0.2884 | 1.602 | 1455 | 0.186 | 1071 | 0.372 | 0.476 | 0.495 |
| 2003 | 0.2073 | 1.152 | 1452 | 0.186 | 1107 | 0.384 | 0.492 | 0.593 |
| 2004 | 0.2291 | 1.273 | 1515 | 0.194 | 1180 | 0.409 | 0.524 | 0.575 |
| 2005 | 0.1514 | 0.841 | 1611 | 0.206 | 1248 | 0.433 | 0.555 | 0.665 |
| 2006 | 0.1541 | 0.856 | 1804 | 0.231 | 1381 | 0.479 | 0.614 | 0.659 |
| 2007 | 0.2534 | 1.408 | 1837 | 0.235 | 1456 | 0.505 | 0.647 | 0.567 |
| 2008 | 0.2918 | 1.621 | 1688 | 0.216 | 1383 | 0.480 | 0.615 | 0.531 |
| 2009 | 0.1897 | 1.054 | 1593 | 0.204 | 1283 | 0.445 | 0.570 | 0.619 |
| 2010 | 0.1638 | 0.910 | 1756 | 0.224 | 1334 | 0.463 | 0.593 | 0.644 |
| 2011 | 0.2161 | 1.201 | 1928 | 0.246 | 1461 | 0.507 | 0.649 | 0.596 |
| 2012 | 0.1827 | 1.015 | 1898 | 0.242 | 1521 | 0.527 | 0.676 | 0.630 |
| 2013 | 0.1482 | 0.823 | 1743 | 0.223 | 1481 | 0.513 | 0.658 | 0.665 |
| 2014 | 0.1306 | 0.726 | 1636 | 0.209 | 1388 | 0.481 | 0.617 | 0.679 |
| 2015 | 0.1895 | 1.053 | 1487 | 0.190 | 1262 | 0.438 | 0.561 | 0.608 |
| 2016 | 0.5679 | 3.155 | 1346 | 0.172 | 1070 | 0.371 | 0.476 | 0.409 |
| 2017 | 0.2807 | 1.559 | 937 | 0.120 | 780 | 0.271 | 0.347 | 0.542 |

Table 20. Selectivity at age for landings from commercial handline ( $c \mathrm{Hl}$ ), commercial trawl ( $c \mathrm{Tw}$ ), recreational headboat ( rHb ), and recreational MRIP (rGe) fleets, discards for commercial handline (cHl.D), recreational headboat (rHb.D), and recreational MRIP (rGe.D) fleets, selectivity for the SERFS chevron trap (sCT) survey, selectivity of landings averaged across fisheries (L.avg), selectivity of discard mortalities averaged across fisheries (D.avg), and selectivity of total removals (Total = L.avg + D.avg). Selectivities of landings and discards from the MRIP fleet were assumed equal to those from the headboat fleet. For time-varying selectivities ( $\mathrm{cHl}, \mathrm{rHb}$, and rGe), values shown are from the first year of each constant selectivity time period.

| Age | TL <br> (mm) | $\begin{aligned} & \mathrm{TL} \\ & \text { (in) } \end{aligned}$ | $\begin{aligned} & \text { cHl- } \\ & 1972 \end{aligned}$ | $\begin{aligned} & \text { cHl- } \\ & 1999 \end{aligned}$ | cTw | $\begin{aligned} & \text { rHb- } \\ & 1972 \end{aligned}$ | $\begin{aligned} & \text { rHb- } \\ & 1992 \end{aligned}$ | $\begin{aligned} & \text { rHb- } \\ & 1999 \end{aligned}$ | $\begin{aligned} & \text { rGe- } \\ & 1972 \end{aligned}$ | rGe1992 | $\begin{aligned} & \text { rGe- } \\ & 1999 \end{aligned}$ | cHl.D | rHb.D | rGe.D | sCT | L.avg | D.avg | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 249 | 9.81 | 0.01 | 0.01 | 0.62 | 0.30 | 0.03 | 0.01 | 0.30 | 0.03 | 0.01 | 0.01 | 0.30 | 0.30 | 0.14 | 0.01 | 0.03 | 0.03 |
| 2 | 294 | 11.58 | 0.69 | 0.08 | 0.99 | 1.00 | 0.46 | 0.04 | 1.00 | 0.46 | 0.04 | 0.69 | 1.00 | 1.00 | 0.38 | 0.05 | 0.12 | 0.16 |
| 3 | 327 | 12.89 | 1.00 | 0.54 | 1.00 | 1.00 | 0.97 | 0.21 | 1.00 | 0.97 | 0.21 | 1.00 | 1.00 | 1.00 | 0.71 | 0.29 | 0.13 | 0.41 |
| 4 | 352 | 13.86 | 1.00 | 0.94 | 1.00 | 1.00 | 1.00 | 0.64 | 1.00 | 1.00 | 0.64 | 1.00 | 1.00 | 1.00 | 0.91 | 0.66 | 0.13 | 0.78 |
| 5 | 370 | 14.58 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.92 | 1.00 | 1.00 | 0.92 | 1.00 | 1.00 | 1.00 | 0.98 | 0.83 | 0.13 | 0.96 |
| 6 | 384 | 15.11 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.99 | 1.00 | 1.00 | 0.99 | 1.00 | 1.00 | 1.00 | 0.99 | 0.87 | 0.13 | 0.99 |
| 7 | 394 | 15.51 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.87 | 0.13 | 1.00 |
| 8 | 401 | 15.80 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.87 | 0.13 | 1.00 |
| 9 | 407 | 16.02 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.87 | 0.13 | 1.00 |
| 10 | 411 | 16.18 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.87 | 0.13 | 1.00 |
| 11 | 414 | 16.30 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.87 | 0.13 | 1.00 |
| 12 | 416 | 16.39 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.87 | 0.13 | 1.00 |
| 13 | 418 | 16.45 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.87 | 0.13 | 1.00 |
| 14 | 419 | 16.50 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.87 | 0.13 | 1.00 |

Table 21. Estimated time series of fully selected fishing mortality rates for commercial handline (F.cHl), commercial trawl (F.cTw), recreational headboat (F.rHb), recreational MRIP (F.rGe), commercial handline dead discards (F.cHl.D), recreational headboat dead discards (F.rHb.D), MRIP dead discards (F.rGe.D). Also shown is apical F, the maximum $F$ at age summed across fleets.

| Year | F.cHl | F.cTw | F.rHb | F.rGe | F.cHl.D | F.rHb.D | F.rGe.D | Apical F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 0.004 | 0.000 | 0.035 | 0.013 | 0.000 | 0.000 | 0.000 | 0.052 |
| 1973 | 0.002 | 0.001 | 0.046 | 0.013 | 0.000 | 0.000 | 0.000 | 0.062 |
| 1974 | 0.014 | 0.000 | 0.034 | 0.013 | 0.000 | 0.000 | 0.000 | 0.061 |
| 1975 | 0.026 | 0.000 | 0.035 | 0.013 | 0.000 | 0.000 | 0.000 | 0.075 |
| 1976 | 0.030 | 0.005 | 0.033 | 0.014 | 0.000 | 0.000 | 0.000 | 0.083 |
| 1977 | 0.043 | 0.015 | 0.033 | 0.011 | 0.000 | 0.000 | 0.000 | 0.103 |
| 1978 | 0.085 | 0.001 | 0.024 | 0.009 | 0.000 | 0.000 | 0.000 | 0.118 |
| 1979 | 0.111 | 0.008 | 0.020 | 0.010 | 0.000 | 0.000 | 0.000 | 0.149 |
| 1980 | 0.122 | 0.033 | 0.026 | 0.012 | 0.000 | 0.000 | 0.000 | 0.193 |
| 1981 | 0.193 | 0.039 | 0.029 | 0.010 | 0.000 | 0.000 | 0.000 | 0.270 |
| 1982 | 0.258 | 0.036 | 0.058 | 0.012 | 0.000 | 0.000 | 0.000 | 0.363 |
| 1983 | 0.289 | 0.022 | 0.040 | 0.005 | 0.000 | 0.000 | 0.000 | 0.357 |
| 1984 | 0.315 | 0.014 | 0.036 | 0.072 | 0.000 | 0.000 | 0.000 | 0.438 |
| 1985 | 0.292 | 0.004 | 0.053 | 0.059 | 0.000 | 0.000 | 0.002 | 0.411 |
| 1986 | 0.340 | 0.004 | 0.051 | 0.011 | 0.000 | 0.000 | 0.000 | 0.407 |
| 1987 | 0.314 | 0.003 | 0.059 | 0.024 | 0.000 | 0.000 | 0.003 | 0.403 |
| 1988 | 0.404 | 0.007 | 0.057 | 0.071 | 0.000 | 0.000 | 0.000 | 0.539 |
| 1989 | 0.413 | 0.000 | 0.046 | 0.044 | 0.000 | 0.000 | 0.003 | 0.505 |
| 1990 | 0.526 | 0.000 | 0.036 | 0.037 | 0.000 | 0.000 | 0.000 | 0.600 |
| 1991 | 0.410 | 0.000 | 0.049 | 0.023 | 0.000 | 0.000 | 0.000 | 0.482 |
| 1992 | 0.263 | 0.000 | 0.057 | 0.105 | 0.012 | 0.004 | 0.002 | 0.443 |
| 1993 | 0.260 | 0.000 | 0.059 | 0.040 | 0.012 | 0.004 | 0.002 | 0.376 |
| 1994 | 0.262 | 0.000 | 0.056 | 0.056 | 0.012 | 0.004 | 0.001 | 0.392 |
| 1995 | 0.267 | 0.000 | 0.058 | 0.037 | 0.012 | 0.004 | 0.004 | 0.383 |
| 1996 | 0.241 | 0.000 | 0.048 | 0.049 | 0.012 | 0.004 | 0.001 | 0.355 |
| 1997 | 0.250 | 0.000 | 0.041 | 0.016 | 0.012 | 0.004 | 0.000 | 0.324 |
| 1998 | 0.194 | 0.000 | 0.045 | 0.026 | 0.012 | 0.004 | 0.001 | 0.282 |
| 1999 | 0.091 | 0.000 | 0.056 | 0.050 | 0.029 | 0.004 | 0.013 | 0.244 |
| 2000 | 0.020 | 0.000 | 0.012 | 0.016 | 0.031 | 0.004 | 0.005 | 0.089 |
| 2001 | 0.046 | 0.000 | 0.040 | 0.056 | 0.033 | 0.011 | 0.029 | 0.215 |
| 2002 | 0.045 | 0.000 | 0.032 | 0.087 | 0.104 | 0.007 | 0.014 | 0.288 |
| 2003 | 0.041 | 0.000 | 0.033 | 0.092 | 0.016 | 0.006 | 0.019 | 0.207 |
| 2004 | 0.036 | 0.000 | 0.036 | 0.114 | 0.013 | 0.012 | 0.018 | 0.229 |
| 2005 | 0.031 | 0.000 | 0.033 | 0.066 | 0.008 | 0.003 | 0.009 | 0.151 |
| 2006 | 0.050 | 0.000 | 0.048 | 0.036 | 0.012 | 0.007 | 0.001 | 0.154 |
| 2007 | 0.082 | 0.000 | 0.087 | 0.063 | 0.007 | 0.007 | 0.008 | 0.253 |
| 2008 | 0.095 | 0.000 | 0.037 | 0.123 | 0.011 | 0.005 | 0.020 | 0.292 |
| 2009 | 0.092 | 0.000 | 0.022 | 0.060 | 0.011 | 0.003 | 0.002 | 0.190 |
| 2010 | 0.092 | 0.000 | 0.025 | 0.033 | 0.007 | 0.002 | 0.004 | 0.164 |
| 2011 | 0.120 | 0.000 | 0.025 | 0.062 | 0.003 | 0.002 | 0.003 | 0.216 |
| 2012 | 0.089 | 0.000 | 0.027 | 0.057 | 0.007 | 0.003 | 0.001 | 0.183 |
| 2013 | 0.083 | 0.000 | 0.018 | 0.036 | 0.007 | 0.002 | 0.002 | 0.148 |
| 2014 | 0.074 | 0.000 | 0.016 | 0.022 | 0.008 | 0.003 | 0.007 | 0.131 |
| 2015 | 0.079 | 0.000 | 0.016 | 0.066 | 0.010 | 0.004 | 0.015 | 0.190 |
| 2016 | 0.087 | 0.000 | 0.021 | 0.374 | 0.007 | 0.004 | 0.075 | 0.568 |
| 2017 | 0.119 | 0.000 | 0.023 | 0.113 | 0.010 | 0.004 | 0.012 | 0.281 |

Table 23. Estimated total landings at age in numbers (1000 fish)

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 37.62 | 76.77 | 53.12 | 38.09 | 28.18 | 21.22 | 16.14 | 12.34 | 9.47 | 7.28 | 5.60 | 4.32 | 3.32 | 10.58 |
| 1973 | 46.86 | 105.09 | 64.44 | 45.27 | 33.30 | 25.06 | 19.08 | 14.62 | 11.24 | 8.66 | 6.68 | 5.15 | 3.97 | 12.81 |
| 1974 | 35.66 | 92.01 | 70.11 | 44.39 | 32.00 | 23.94 | 18.21 | 13.97 | 10.76 | 8.31 | 6.42 | 4.96 | 3.83 | 12.50 |
| 1975 | 20.79 | 108.99 | 82.18 | 60.39 | 39.22 | 28.75 | 21.75 | 16.67 | 12.86 | 9.95 | 7.70 | 5.97 | 4.62 | 15.21 |
| 1976 | 39.96 | 66.47 | 90.45 | 63.08 | 47.54 | 31.40 | 23.27 | 17.74 | 13.67 | 10.59 | 8.22 | 6.38 | 4.94 | 16.45 |
| 1977 | 187.71 | 129.64 | 61.64 | 77.14 | 55.18 | 42.29 | 28.24 | 21.09 | 16.16 | 12.51 | 9.72 | 7.56 | 5.87 | 19.73 |
| 1978 | 31.07 | 469.71 | 111.33 | 47.58 | 61.07 | 44.42 | 34.42 | 23.16 | 17.39 | 13.38 | 10.39 | 8.09 | 6.30 | 21.36 |
| 1979 | 23.91 | 209.83 | 487.99 | 92.28 | 40.43 | 52.77 | 38.81 | 30.30 | 20.50 | 15.46 | 11.93 | 9.28 | 7.24 | 24.77 |
| 1980 | 92.06 | 152.64 | 218.80 | 400.20 | 77.58 | 34.57 | 45.61 | 33.80 | 26.54 | 18.03 | 13.64 | 10.55 | 8.22 | 28.36 |
| 1981 | 87.21 | 350.52 | 155.89 | 182.95 | 343.08 | 67.63 | 30.47 | 40.51 | 30.19 | 23.80 | 16.22 | 12.29 | 9.52 | 33.04 |
| 1982 | 69.28 | 382.70 | 330.81 | 115.19 | 138.58 | 264.26 | 52.67 | 23.91 | 31.97 | 23.92 | 18.91 | 12.91 | 9.80 | 33.98 |
| 1983 | 68.00 | 237.91 | 253.75 | 170.18 | 60.76 | 74.35 | 143.37 | 28.80 | 13.14 | 17.65 | 13.24 | 10.49 | 7.18 | 24.35 |
| 1984 | 89.89 | 440.90 | 199.40 | 157.63 | 108.37 | 39.35 | 48.68 | 94.60 | 19.11 | 8.76 | 11.79 | 8.87 | 7.04 | 21.16 |
| 1985 | 81.63 | 372.59 | 263.48 | 91.22 | 73.94 | 51.70 | 18.98 | 23.67 | 46.25 | 9.38 | 4.31 | 5.82 | 4.38 | 13.95 |
| 1986 | 45.41 | 368.70 | 239.85 | 130.71 | 46.40 | 38.26 | 27.05 | 10.01 | 12.55 | 24.62 | 5.01 | 2.31 | 3.12 | 9.83 |
| 1987 | 48.45 | 337.35 | 249.58 | 117.99 | 65.92 | 23.80 | 19.84 | 14.14 | 5.26 | 6.62 | 13.03 | 2.66 | 1.23 | 6.89 |
| 1988 | 129.72 | 382.72 | 288.18 | 158.96 | 77.03 | 43.77 | 15.98 | 13.42 | 9.62 | 3.59 | 4.54 | 8.95 | 1.83 | 5.58 |
| 1989 | 63.23 | 589.27 | 213.09 | 119.45 | 67.57 | 33.30 | 19.14 | 7.04 | 5.95 | 4.28 | 1.60 | 2.03 | 4.01 | 3.32 |
| 1990 | 49.15 | 508.25 | 423.39 | 110.44 | 63.46 | 36.51 | 18.20 | 10.54 | 3.90 | 3.31 | 2.39 | 0.90 | 1.14 | 4.11 |
| 1991 | 41.48 | 388.41 | 256.44 | 147.46 | 39.44 | 23.05 | 13.41 | 6.74 | 3.92 | 1.46 | 1.24 | 0.90 | 0.34 | 1.98 |
| 1992 | 6.53 | 258.89 | 224.07 | 106.56 | 62.84 | 17.10 | 10.11 | 5.93 | 2.99 | 1.75 | 0.65 | 0.56 | 0.40 | 1.04 |
| 1993 | 4.63 | 153.91 | 188.13 | 95.24 | 45.76 | 27.44 | 7.55 | 4.50 | 2.65 | 1.35 | 0.79 | 0.29 | 0.25 | 0.65 |
| 1994 | 5.63 | 157.67 | 134.98 | 101.83 | 52.31 | 25.55 | 15.49 | 4.30 | 2.57 | 1.52 | 0.78 | 0.46 | 0.17 | 0.52 |
| 1995 | 7.05 | 172.07 | 128.89 | 67.18 | 51.38 | 26.83 | 13.26 | 8.10 | 2.26 | 1.36 | 0.81 | 0.41 | 0.24 | 0.37 |
| 1996 | 3.47 | 222.05 | 135.50 | 62.89 | 33.28 | 25.88 | 13.67 | 6.80 | 4.18 | 1.17 | 0.71 | 0.42 | 0.21 | 0.32 |
| 1997 | 3.09 | 106.38 | 177.01 | 66.50 | 31.31 | 16.84 | 13.24 | 7.05 | 3.53 | 2.18 | 0.61 | 0.37 | 0.22 | 0.28 |
| 1998 | 3.43 | 106.05 | 79.27 | 84.84 | 32.45 | 15.54 | 8.45 | 6.70 | 3.58 | 1.80 | 1.11 | 0.31 | 0.19 | 0.26 |
| 1999 | 0.95 | 8.50 | 28.16 | 27.46 | 36.07 | 14.52 | 7.07 | 3.88 | 3.09 | 1.66 | 0.84 | 0.52 | 0.15 | 0.21 |
| 2000 | 0.13 | 1.37 | 8.52 | 9.91 | 5.46 | 6.06 | 2.38 | 1.16 | 0.64 | 0.51 | 0.28 | 0.14 | 0.09 | 0.06 |
| 2001 | 0.76 | 2.05 | 15.54 | 36.29 | 24.29 | 11.30 | 12.25 | 4.82 | 2.36 | 1.31 | 1.05 | 0.57 | 0.29 | 0.30 |
| 2002 | 1.30 | 4.58 | 8.87 | 25.95 | 32.89 | 17.94 | 8.09 | 8.77 | 3.47 | 1.71 | 0.95 | 0.76 | 0.41 | 0.43 |
| 2003 | 0.96 | 6.98 | 18.05 | 13.72 | 21.20 | 21.54 | 11.34 | 5.11 | 5.57 | 2.21 | 1.09 | 0.61 | 0.49 | 0.54 |
| 2004 | 1.07 | 5.41 | 29.89 | 32.21 | 12.88 | 15.83 | 15.49 | 8.16 | 3.70 | 4.04 | 1.61 | 0.80 | 0.44 | 0.75 |
| 2005 | 0.92 | 4.01 | 16.36 | 37.26 | 19.99 | 6.20 | 7.31 | 7.16 | 3.79 | 1.72 | 1.89 | 0.75 | 0.37 | 0.56 |
| 2006 | 1.14 | 5.62 | 19.79 | 30.26 | 34.51 | 14.60 | 4.36 | 5.14 | 5.05 | 2.68 | 1.22 | 1.35 | 0.54 | 0.67 |
| 2007 | 1.32 | 11.28 | 40.38 | 51.18 | 42.60 | 40.10 | 16.48 | 4.93 | 5.84 | 5.76 | 3.07 | 1.40 | 1.54 | 1.38 |
| 2008 | 0.99 | 8.43 | 52.67 | 66.20 | 43.81 | 29.54 | 26.90 | 11.07 | 3.33 | 3.95 | 3.91 | 2.09 | 0.96 | 2.00 |
| 2009 | 1.02 | 4.42 | 27.88 | 56.70 | 36.03 | 19.25 | 12.55 | 11.44 | 4.73 | 1.43 | 1.70 | 1.69 | 0.90 | 1.28 |
| 2010 | 1.42 | 6.07 | 18.06 | 35.12 | 39.17 | 21.25 | 11.11 | 7.27 | 6.65 | 2.76 | 0.84 | 1.00 | 0.99 | 1.28 |
| 2011 | 1.87 | 12.96 | 35.97 | 32.82 | 36.77 | 36.05 | 19.26 | 10.11 | 6.64 | 6.11 | 2.54 | 0.77 | 0.92 | 2.10 |
| 2012 | 0.96 | 9.90 | 45.24 | 39.80 | 20.77 | 20.24 | 19.49 | 10.45 | 5.51 | 3.64 | 3.35 | 1.40 | 0.42 | 1.67 |
| 2013 | 0.42 | 5.24 | 37.48 | 53.32 | 25.63 | 11.38 | 10.84 | 10.47 | 5.64 | 2.99 | 1.98 | 1.83 | 0.76 | 1.14 |
| 2014 | 0.46 | 2.38 | 20.02 | 43.37 | 35.02 | 14.74 | 6.44 | 6.16 | 5.98 | 3.23 | 1.72 | 1.14 | 1.05 | 1.10 |
| 2015 | 0.47 | 4.09 | 13.22 | 35.94 | 47.44 | 34.44 | 14.34 | 6.29 | 6.05 | 5.89 | 3.20 | 1.70 | 1.13 | 2.14 |
| 2016 | 1.74 | 6.68 | 36.07 | 43.96 | 71.11 | 80.74 | 57.38 | 23.96 | 10.56 | 10.19 | 9.96 | 5.41 | 2.88 | 5.55 |
| 2017 | 0.42 | 5.81 | 18.36 | 33.63 | 17.37 | 20.52 | 22.07 | 15.65 | 6.56 | 2.90 | 2.81 | 2.75 | 1.50 | 2.34 |

Table 24. Estimated total landings at age in whole weight (1000 lb)

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 19.31 | 63.67 | 60.09 | 53.16 | 45.53 | 38.04 | 31.17 | 25.16 | 20.08 | 15.89 | 12.49 | 9.77 | 7.62 | 24.43 |
| 1973 | 24.06 | 87.16 | 72.90 | 63.18 | 53.81 | 44.91 | 36.84 | 29.80 | 23.84 | 18.90 | 14.89 | 11.66 | 9.10 | 29.59 |
| 1974 | 18.31 | 76.32 | 79.31 | 61.96 | 51.70 | 42.91 | 35.17 | 28.48 | 22.83 | 18.14 | 14.32 | 11.24 | 8.78 | 28.88 |
| 1975 | 10.67 | 90.40 | 92.97 | 84.29 | 63.38 | 51.54 | 42.00 | 33.99 | 27.27 | 21.71 | 17.18 | 13.51 | 10.58 | 35.14 |
| 1976 | 20.51 | 55.13 | 102.32 | 88.04 | 76.82 | 56.29 | 44.95 | 36.16 | 28.99 | 23.11 | 18.32 | 14.44 | 11.33 | 38.01 |
| 1977 | 96.37 | 107.53 | 69.73 | 107.67 | 89.15 | 75.81 | 54.55 | 43.00 | 34.28 | 27.30 | 21.67 | 17.11 | 13.45 | 45.57 |
| 1978 | 15.95 | 389.61 | 125.94 | 66.41 | 98.67 | 79.62 | 66.48 | 47.22 | 36.88 | 29.21 | 23.16 | 18.31 | 14.43 | 49.34 |
| 1979 | 12.28 | 174.05 | 552.01 | 128.80 | 65.33 | 94.59 | 74.95 | 61.78 | 43.48 | 33.74 | 26.60 | 21.02 | 16.57 | 57.22 |
| 1980 | 47.26 | 126.61 | 247.51 | 558.56 | 125.35 | 61.96 | 88.10 | 68.91 | 56.28 | 39.36 | 30.40 | 23.88 | 18.82 | 65.51 |
| 1981 | 44.78 | 290.75 | 176.34 | 255.35 | 554.33 | 121.23 | 58.85 | 82.59 | 64.02 | 51.95 | 36.15 | 27.83 | 21.80 | 76.33 |
| 1982 | 35.57 | 317.44 | 374.21 | 160.77 | 223.91 | 473.71 | 101.74 | 48.75 | 67.80 | 52.21 | 42.17 | 29.24 | 22.45 | 78.50 |
| 1983 | 34.91 | 197.34 | 287.04 | 237.53 | 98.17 | 133.27 | 276.91 | 58.71 | 27.88 | 38.52 | 29.52 | 23.76 | 16.44 | 56.26 |
| 1984 | 46.15 | 365.71 | 225.56 | 220.00 | 175.09 | 70.53 | 94.03 | 192.85 | 40.52 | 19.11 | 26.29 | 20.08 | 16.12 | 48.89 |
| 1985 | 41.91 | 309.05 | 298.05 | 127.32 | 119.47 | 92.68 | 36.67 | 48.26 | 98.09 | 20.47 | 9.61 | 13.18 | 10.04 | 32.23 |
| 1986 | 23.32 | 305.83 | 271.32 | 182.43 | 74.97 | 68.58 | 52.25 | 20.41 | 26.61 | 53.75 | 11.17 | 5.23 | 7.14 | 22.72 |
| 1987 | 24.87 | 279.82 | 282.32 | 164.68 | 106.51 | 42.66 | 38.33 | 28.83 | 11.16 | 14.46 | 29.06 | 6.02 | 2.81 | 15.91 |
| 1988 | 66.60 | 317.46 | 325.99 | 221.87 | 124.47 | 78.45 | 30.86 | 27.37 | 20.40 | 7.84 | 10.12 | 20.26 | 4.18 | 12.90 |
| 1989 | 32.46 | 488.78 | 241.04 | 166.72 | 109.17 | 59.70 | 36.96 | 14.35 | 12.61 | 9.34 | 3.57 | 4.59 | 9.18 | 7.68 |
| 1990 | 25.24 | 421.58 | 478.94 | 154.14 | 102.53 | 65.44 | 35.15 | 21.48 | 8.27 | 7.22 | 5.32 | 2.03 | 2.60 | 9.49 |
| 1991 | 21.30 | 322.17 | 290.08 | 205.81 | 63.72 | 41.33 | 25.91 | 13.74 | 8.32 | 3.18 | 2.76 | 2.03 | 0.77 | 4.56 |
| 1992 | 3.35 | 214.74 | 253.47 | 148.73 | 101.54 | 30.64 | 19.52 | 12.08 | 6.35 | 3.82 | 1.45 | 1.26 | 0.92 | 2.40 |
| 1993 | 2.38 | 127.66 | 212.81 | 132.93 | 73.94 | 49.18 | 14.58 | 9.17 | 5.62 | 2.94 | 1.76 | 0.67 | 0.58 | 1.51 |
| 1994 | 2.89 | 130.78 | 152.69 | 142.13 | 84.52 | 45.81 | 29.93 | 8.76 | 5.46 | 3.33 | 1.73 | 1.03 | 0.39 | 1.21 |
| 1995 | 3.62 | 142.73 | 145.80 | 93.77 | 83.01 | 48.10 | 25.60 | 16.51 | 4.79 | 2.96 | 1.80 | 0.93 | 0.55 | 0.85 |
| 1996 | 1.78 | 184.18 | 153.28 | 87.78 | 53.77 | 46.39 | 26.40 | 13.87 | 8.86 | 2.55 | 1.57 | 0.95 | 0.49 | 0.74 |
| 1997 | 1.58 | 88.24 | 200.23 | 92.81 | 50.58 | 30.19 | 25.58 | 14.37 | 7.48 | 4.75 | 1.36 | 0.84 | 0.50 | 0.65 |
| 1998 | 1.76 | 87.97 | 89.67 | 118.41 | 52.44 | 27.85 | 16.33 | 13.65 | 7.60 | 3.93 | 2.49 | 0.71 | 0.44 | 0.59 |
| 1999 | 0.49 | 7.05 | 31.85 | 38.32 | 58.29 | 26.03 | 13.65 | 7.90 | 6.55 | 3.62 | 1.87 | 1.18 | 0.34 | 0.48 |
| 2000 | 0.07 | 1.13 | 9.63 | 13.83 | 8.82 | 10.87 | 4.60 | 2.37 | 1.36 | 1.12 | 0.62 | 0.32 | 0.20 | 0.14 |
| 2001 | 0.39 | 1.70 | 17.58 | 50.65 | 39.25 | 20.26 | 23.66 | 9.83 | 5.01 | 2.85 | 2.34 | 1.28 | 0.66 | 0.69 |
| 2002 | 0.67 | 3.80 | 10.04 | 36.22 | 53.15 | 32.17 | 15.62 | 17.88 | 7.36 | 3.73 | 2.11 | 1.73 | 0.94 | 0.99 |
| 2003 | 0.49 | 5.79 | 20.42 | 19.14 | 34.26 | 38.60 | 21.90 | 10.42 | 11.82 | 4.83 | 2.43 | 1.38 | 1.12 | 1.25 |
| 2004 | 0.55 | 4.49 | 33.81 | 44.95 | 20.81 | 28.37 | 29.92 | 16.64 | 7.84 | 8.82 | 3.59 | 1.80 | 1.02 | 1.74 |
| 2005 | 0.47 | 3.33 | 18.50 | 52.01 | 32.30 | 11.12 | 14.13 | 14.59 | 8.03 | 3.76 | 4.21 | 1.71 | 0.85 | 1.30 |
| 2006 | 0.58 | 4.66 | 22.39 | 42.23 | 55.75 | 26.18 | 8.43 | 10.49 | 10.72 | 5.86 | 2.73 | 3.05 | 1.23 | 1.54 |
| 2007 | 0.68 | 9.36 | 45.68 | 71.44 | 68.84 | 71.89 | 31.84 | 10.05 | 12.39 | 12.58 | 6.84 | 3.18 | 3.54 | 3.20 |
| 2008 | 0.51 | 6.99 | 59.58 | 92.39 | 70.79 | 52.95 | 51.96 | 22.56 | 7.05 | 8.63 | 8.72 | 4.73 | 2.19 | 4.62 |
| 2009 | 0.53 | 3.67 | 31.54 | 79.13 | 58.21 | 34.50 | 24.24 | 23.32 | 10.03 | 3.11 | 3.79 | 3.82 | 2.07 | 2.95 |
| 2010 | 0.73 | 5.04 | 20.43 | 49.02 | 63.28 | 38.09 | 21.46 | 14.81 | 14.11 | 6.03 | 1.86 | 2.26 | 2.27 | 2.96 |
| 2011 | 0.96 | 10.75 | 40.69 | 45.81 | 59.40 | 64.62 | 37.20 | 20.61 | 14.09 | 13.33 | 5.67 | 1.75 | 2.11 | 4.86 |
| 2012 | 0.49 | 8.21 | 51.18 | 55.55 | 33.56 | 36.28 | 37.65 | 21.31 | 11.69 | 7.94 | 7.48 | 3.17 | 0.97 | 3.85 |
| 2013 | 0.21 | 4.35 | 42.40 | 74.42 | 41.42 | 20.40 | 20.94 | 21.35 | 11.96 | 6.52 | 4.41 | 4.14 | 1.75 | 2.64 |
| 2014 | 0.24 | 1.97 | 22.64 | 60.54 | 56.58 | 26.42 | 12.43 | 12.55 | 12.67 | 7.05 | 3.83 | 2.58 | 2.41 | 2.54 |
| 2015 | 0.24 | 3.39 | 14.95 | 50.16 | 76.65 | 61.73 | 27.70 | 12.82 | 12.82 | 12.86 | 7.12 | 3.85 | 2.59 | 4.94 |
| 2016 | 0.89 | 5.54 | 40.80 | 61.35 | 114.90 | 144.73 | 110.83 | 48.84 | 22.39 | 22.24 | 22.20 | 12.25 | 6.61 | 12.82 |
| 2017 | 0.22 | 4.82 | 20.76 | 46.94 | 28.07 | 36.78 | 42.63 | 31.91 | 13.92 | 6.34 | 6.27 | 6.23 | 3.43 | 5.40 |

Table 25. Estimated time series of landings in numbers (1000 fish) for commercial handline (L.cHl), commercial trawl (L.cTw), recreational headboat (L.rHb), and MRIP (L.rGe)

| Year | L.cHl | L.cTw | L.rHb | L.rGe |
| ---: | ---: | ---: | ---: | ---: |
| 1972 | 21.89 | 0.56 | 220.03 | 81.56 |
| 1973 | 10.09 | 10.75 | 299.82 | 81.56 |
| 1974 | 74.92 | 0.56 | 220.04 | 81.57 |
| 1975 | 136.89 | 0.91 | 215.68 | 81.57 |
| 1976 | 140.57 | 31.27 | 186.77 | 81.55 |
| 1977 | 195.25 | 154.16 | 243.54 | 81.53 |
| 1978 | 587.17 | 7.15 | 223.80 | 81.55 |
| 1979 | 756.83 | 70.48 | 156.63 | 81.57 |
| 1980 | 663.61 | 246.86 | 168.55 | 81.57 |
| 1981 | 899.97 | 258.72 | 168.44 | 56.19 |
| 1982 | 993.15 | 187.60 | 273.23 | 54.91 |
| 1983 | 842.36 | 103.66 | 155.85 | 21.28 |
| 1984 | 806.95 | 60.45 | 129.95 | 258.18 |
| 1985 | 673.01 | 16.34 | 176.46 | 195.53 |
| 1986 | 750.58 | 15.89 | 161.06 | 36.30 |
| 1987 | 658.03 | 10.17 | 173.90 | 70.66 |
| 1988 | 737.26 | 28.89 | 168.81 | 208.93 |
| 1989 | 848.22 | 0.00 | 146.52 | 138.53 |
| 1990 | 1023.57 | 0.00 | 104.71 | 107.37 |
| 1991 | 736.14 | 0.00 | 129.65 | 60.48 |
| 1992 | 455.22 | 0.00 | 85.82 | 158.38 |
| 1993 | 396.86 | 0.00 | 81.70 | 54.58 |
| 1994 | 363.06 | 0.00 | 70.40 | 70.33 |
| 1995 | 364.60 | 0.00 | 70.68 | 44.93 |
| 1996 | 379.76 | 0.00 | 64.85 | 65.94 |
| 1997 | 353.96 | 0.00 | 53.89 | 20.75 |
| 1998 | 259.06 | 0.00 | 53.84 | 31.09 |
| 1999 | 72.85 | 0.00 | 31.88 | 28.33 |
| 2000 | 18.11 | 0.00 | 8.03 | 10.55 |
| 2001 | 43.98 | 0.00 | 28.85 | 40.35 |
| 2002 | 37.52 | 0.00 | 20.93 | 57.68 |
| 2003 | 33.48 | 0.00 | 20.17 | 55.76 |
| 2004 | 34.26 | 0.00 | 23.46 | 74.56 |
| 2005 | 33.23 | 0.00 | 24.83 | 50.24 |
| 2006 | 56.50 | 0.00 | 40.38 | 30.06 |
| 2007 | 97.42 | 0.00 | 75.20 | 54.67 |
| 2008 | 116.24 | 0.00 | 32.52 | 107.07 |
| 2009 | 108.47 | 0.00 | 19.54 | 53.00 |
| 2010 | 101.86 | 0.00 | 21.91 | 29.22 |
| 2011 | 131.73 | 0.00 | 21.05 | 52.11 |
| 2012 | 109.96 | 0.00 | 23.21 | 49.69 |
| 2013 | 115.63 | 0.00 | 17.72 | 35.78 |
| 2014 | 101.89 | 0.00 | 17.18 | 23.72 |
| 2015 | 94.94 | 0.00 | 15.54 | 65.84 |
| 2016 | 77.33 | 0.00 | 15.30 | 273.55 |
| 2017 | 80.15 | 0.00 | 12.32 | 60.23 |
|  |  |  |  |  |

Table 26. Estimated time series of landings in weight (1000 lb) for commercial handline (L.cHl), commercial trawl (L.cTw), recreational headboat (L.rHb), and MRIP (L.rGe)

| Year | L.cHl | L.cTw | L.rHb | L.rGe |
| ---: | ---: | ---: | ---: | ---: |
| 1972 | 32.17 | 0.67 | 287.14 | 106.43 |
| 1973 | 14.65 | 12.95 | 387.61 | 105.44 |
| 1974 | 108.37 | 0.67 | 284.01 | 105.29 |
| 1975 | 197.85 | 1.16 | 287.06 | 108.56 |
| 1976 | 211.77 | 39.26 | 252.94 | 110.45 |
| 1977 | 288.59 | 148.45 | 274.31 | 91.83 |
| 1978 | 719.96 | 7.44 | 244.69 | 89.16 |
| 1979 | 987.84 | 83.14 | 191.64 | 99.81 |
| 1980 | 944.75 | 293.14 | 216.06 | 104.57 |
| 1981 | 1275.95 | 303.49 | 212.10 | 70.76 |
| 1982 | 1390.24 | 223.53 | 345.28 | 69.39 |
| 1983 | 1187.81 | 113.78 | 188.85 | 25.79 |
| 1984 | 1062.43 | 62.06 | 146.12 | 290.31 |
| 1985 | 844.60 | 15.83 | 188.14 | 208.47 |
| 1986 | 906.14 | 15.06 | 166.89 | 37.62 |
| 1987 | 784.72 | 9.68 | 179.93 | 73.10 |
| 1988 | 879.70 | 24.71 | 162.82 | 201.52 |
| 1989 | 928.77 | 0.00 | 137.44 | 129.94 |
| 1990 | 1136.59 | 0.00 | 100.14 | 102.68 |
| 1991 | 822.40 | 0.00 | 124.98 | 58.31 |
| 1992 | 515.17 | 0.00 | 100.20 | 184.91 |
| 1993 | 470.26 | 0.00 | 99.19 | 66.27 |
| 1994 | 436.81 | 0.00 | 86.96 | 86.87 |
| 1995 | 430.89 | 0.00 | 85.67 | 54.46 |
| 1996 | 429.06 | 0.00 | 76.14 | 77.42 |
| 1997 | 427.22 | 0.00 | 66.39 | 25.56 |
| 1998 | 317.02 | 0.00 | 67.71 | 39.10 |
| 1999 | 104.75 | 0.00 | 49.18 | 43.69 |
| 2000 | 26.19 | 0.00 | 12.48 | 16.39 |
| 2001 | 66.16 | 0.00 | 45.86 | 64.13 |
| 2002 | 58.22 | 0.00 | 34.13 | 94.04 |
| 2003 | 50.37 | 0.00 | 32.81 | 90.68 |
| 2004 | 49.71 | 0.00 | 37.01 | 117.62 |
| 2005 | 48.79 | 0.00 | 38.86 | 78.64 |
| 2006 | 84.13 | 0.00 | 64.03 | 47.67 |
| 2007 | 144.62 | 0.00 | 119.79 | 87.08 |
| 2008 | 171.94 | 0.00 | 51.65 | 170.07 |
| 2009 | 164.46 | 0.00 | 31.37 | 85.09 |
| 2010 | 158.49 | 0.00 | 35.94 | 47.93 |
| 2011 | 201.63 | 0.00 | 34.60 | 85.63 |
| 2012 | 162.36 | 0.00 | 37.23 | 79.72 |
| 2013 | 171.72 | 0.00 | 28.21 | 56.96 |
| 2014 | 157.78 | 0.00 | 28.00 | 38.67 |
| 2015 | 154.07 | 0.00 | 26.31 | 111.46 |
| 2016 | 126.69 | 0.00 | 26.47 | 473.23 |
| 2017 | 129.21 | 0.00 | 21.15 | 103.36 |
|  |  |  |  |  |

Table 27. Estimated time series of dead discards in numbers (1000 fish) for commercial handline (L.cHl), recreational headboat (L.rHb), and MRIP (L.rGe)

| Year | D.cHl | D.rHb | D.rGe |
| ---: | ---: | ---: | ---: |
| 1972 | 0.00 | 0.00 | 0.00 |
| 1973 | 0.00 | 0.00 | 0.00 |
| 1974 | 0.00 | 0.00 | 0.00 |
| 1975 | 0.00 | 0.00 | 0.00 |
| 1976 | 0.00 | 0.00 | 0.00 |
| 1977 | 0.00 | 0.00 | 0.00 |
| 1978 | 0.00 | 0.00 | 0.00 |
| 1979 | 0.00 | 0.00 | 0.00 |
| 1980 | 0.00 | 0.00 | 0.00 |
| 1981 | 0.00 | 0.00 | 0.52 |
| 1982 | 0.00 | 0.00 | 0.81 |
| 1983 | 0.00 | 0.00 | 0.29 |
| 1984 | 0.00 | 0.00 | 0.61 |
| 1985 | 0.00 | 0.00 | 6.17 |
| 1986 | 0.00 | 0.00 | 0.23 |
| 1987 | 0.00 | 0.00 | 9.28 |
| 1988 | 0.00 | 0.00 | 0.27 |
| 1989 | 0.00 | 0.00 | 8.64 |
| 1990 | 0.00 | 0.00 | 0.08 |
| 1991 | 0.00 | 0.00 | 0.08 |
| 1992 | 20.57 | 10.55 | 5.34 |
| 1993 | 18.17 | 9.19 | 3.35 |
| 1994 | 16.46 | 8.71 | 1.06 |
| 1995 | 16.22 | 9.36 | 9.10 |
| 1996 | 18.76 | 9.60 | 2.87 |
| 1997 | 16.86 | 8.37 | 0.72 |
| 1998 | 15.90 | 8.31 | 2.66 |
| 1999 | 42.31 | 8.50 | 24.85 |
| 2000 | 46.43 | 7.93 | 9.27 |
| 2001 | 43.26 | 17.96 | 48.28 |
| 2002 | 133.27 | 13.02 | 25.39 |
| 2003 | 24.34 | 12.56 | 39.75 |
| 2004 | 20.73 | 26.14 | 37.65 |
| 2005 | 13.27 | 7.70 | 21.12 |
| 2006 | 21.34 | 17.76 | 3.37 |
| 2007 | 13.43 | 17.46 | 21.36 |
| 2008 | 21.30 | 11.41 | 46.19 |
| 2009 | 17.80 | 6.05 | 4.52 |
| 2010 | 11.14 | 5.30 | 9.52 |
| 2011 | 6.33 | 6.19 | 9.00 |
| 2012 | 14.75 | 6.93 | 3.58 |
| 2013 | 13.97 | 5.70 | 5.92 |
| 2014 | 14.89 | 7.31 | 14.58 |
| 2015 | 16.08 | 7.70 | 27.49 |
| 2016 | 8.88 | 6.33 | 111.57 |
| 2017 | 9.76 | 4.59 | 14.70 |
|  |  |  |  |

Table 28. Estimated time series of dead discards in weight (1000 lb) for commercial handline (L.cHl), recreational headboat (L.rHb), and MRIP (L.rGe)

| Year | D.cHl | D.rHb | D.rGe |
| ---: | ---: | ---: | ---: |
| 1972 | 0.00 | 0.00 | 0.00 |
| 1973 | 0.00 | 0.00 | 0.00 |
| 1974 | 0.00 | 0.00 | 0.00 |
| 1975 | 0.00 | 0.00 | 0.00 |
| 1976 | 0.00 | 0.00 | 0.00 |
| 1977 | 0.00 | 0.00 | 0.00 |
| 1978 | 0.00 | 0.00 | 0.00 |
| 1979 | 0.00 | 0.00 | 0.00 |
| 1980 | 0.00 | 0.00 | 0.00 |
| 1981 | 0.00 | 0.00 | 0.65 |
| 1982 | 0.00 | 0.00 | 1.02 |
| 1983 | 0.00 | 0.00 | 0.35 |
| 1984 | 0.00 | 0.00 | 0.69 |
| 1985 | 0.00 | 0.00 | 6.58 |
| 1986 | 0.00 | 0.00 | 0.23 |
| 1987 | 0.00 | 0.00 | 9.60 |
| 1988 | 0.00 | 0.00 | 0.26 |
| 1989 | 0.00 | 0.00 | 8.10 |
| 1990 | 0.00 | 0.00 | 0.08 |
| 1991 | 0.00 | 0.00 | 0.08 |
| 1992 | 23.27 | 10.59 | 5.36 |
| 1993 | 21.53 | 9.56 | 3.48 |
| 1994 | 19.80 | 9.00 | 1.09 |
| 1995 | 19.16 | 9.17 | 8.92 |
| 1996 | 21.19 | 9.70 | 2.90 |
| 1997 | 20.34 | 8.89 | 0.77 |
| 1998 | 19.46 | 8.75 | 2.81 |
| 1999 | 51.02 | 9.15 | 26.74 |
| 2000 | 58.75 | 9.29 | 10.86 |
| 2001 | 59.21 | 21.64 | 58.16 |
| 2002 | 176.65 | 14.39 | 28.05 |
| 2003 | 29.84 | 13.51 | 42.76 |
| 2004 | 25.89 | 29.00 | 41.76 |
| 2005 | 16.95 | 8.60 | 23.58 |
| 2006 | 27.21 | 19.62 | 3.72 |
| 2007 | 16.87 | 19.70 | 24.11 |
| 2008 | 27.63 | 13.60 | 55.08 |
| 2009 | 24.24 | 7.34 | 5.48 |
| 2010 | 15.02 | 6.04 | 10.85 |
| 2011 | 8.00 | 6.76 | 9.82 |
| 2012 | 18.41 | 7.83 | 4.04 |
| 2013 | 18.39 | 7.02 | 7.29 |
| 2014 | 21.17 | 9.57 | 19.08 |
| 2015 | 23.41 | 10.30 | 36.77 |
| 2016 | 12.95 | 8.25 | 145.36 |
| 2017 | 13.47 | 5.82 | 18.62 |
|  |  |  |  |

Table 29. Estimated status indicators, benchmarks, and related quantities from the base run of the Beaufort catch-age model, conditional on estimated current selectivities averaged across fleets. Also presented are median values and measures of precision (standard errors, SE) from the Monte Carlo/Bootstrap analysis. Rate estimates (F) are in units of $\mathrm{y}^{-1}$; status indicators are dimensionless; and biomass estimates are in units of metric tons or pounds, as indicated. Spawning stock biomass (SSB) is measured in metric tons

| Quantity | Units | Estimate | Median | SE |
| :--- | :--- | ---: | ---: | ---: |
| $F_{\text {MSY }}$ | $\mathrm{y}^{-1}$ | 0.18 | 0.18 | 0.027 |
| $85 \% F_{\text {MSY }}$ | $\mathrm{y}^{-1}$ | 0.153 | 0.153 | 0.023 |
| $75 \% F_{\text {MSY }}$ | $\mathrm{y}^{-1}$ | 0.135 | 0.135 | 0.02 |
| $65 \% F_{\text {MSY }}$ | $\mathrm{y}^{-1}$ | 0.117 | 0.117 | 0.018 |
| $F_{20 \%}$ | $\mathrm{y}^{-1}$ | 3.115 | 3.022 | 1.425 |
| $F_{30 \%}$ | $\mathrm{y}^{-1}$ | 1.21 | 1.175 | 0.796 |
| $F_{40 \%}$ | $\mathrm{y}^{-1}$ | 0.6 | 0.585 | 0.326 |
| $B_{\text {MSY }}$ | metric tons | 3604.6 | 3593.7 | 487.3 |
| $\mathrm{SSB}_{\text {MSY }}$ | metric tons | 2883.7 | 2902.6 | 337.3 |
| $\mathrm{MSST}^{\text {MSY }}$ | metric tons | 2249.3 | 2261 | 219 |
| $D_{\text {MSY }}$ | 1000 lb whole | 531.4 | 538.2 | 59.3 |
| $D_{\text {MSY }}$ | 1000 lb dead fish | 126.8 | 118.5 | 118.5 |
| $R_{\text {MSY }}$ | 1000 dead fish | 104 | 96.7 | 96.7 |
| $L_{85 \% \text { MSY }}$ | 1000 fish | 2641.2 | 2554.1 | 796.3 |
| $L_{75 \% \text { MSY }}$ | 1000 lb whole | 527.2 | 532.8 | 59.4 |
| $L_{65 \% \text { MSY }}$ | 1000 lb whole | 515.7 | 521.9 | 59.4 |
| $F_{2015-2017} / F_{\text {MSY }}$ | 1000 lb whole | 494.9 | 502.9 | 59.1 |
| SSB $_{2017} /$ MSST | - | 1.73 | 1.664 | 0.304 |
| SSB $_{2017} /$ SSB $_{\text {MSY }}$ | - | 0.347 | 0.369 | 0.101 |

Table 30. Results from sensitivity runs of the Beaufort catch-age model. Current $F$ represented by geometric mean of last three assessment years $\left(F / F_{\mathrm{MSY}}=F_{2015-2017} / F_{\mathrm{MSY}}\right)$. Stock and rebuild status based on terminal year ( $\mathrm{SSB} / \mathrm{MSST}=\mathrm{SSB}_{2017} / \mathrm{MSST}$; $\mathrm{SSB} / \mathrm{SSB}_{\mathrm{MSY}}=$ $\left.\mathrm{SSB}_{2017} / \mathrm{SSB}_{\mathrm{MSY}}\right) . h=$ Beverton-Holt steepness. See text for full description of sensitivity runs.

| Description | $F_{\text {MSY }}$ | $\begin{aligned} & \mathrm{SSB}_{\mathrm{MSY}} \\ & (m t) \end{aligned}$ | $\begin{aligned} & B_{\mathrm{MSY}} \\ & (m t) \end{aligned}$ | $\begin{aligned} & \text { MSY } \\ & (1000 \mathrm{lb}) \end{aligned}$ | $\begin{aligned} & D_{\text {MSY }} \\ & (1000 \mathrm{lb}) \end{aligned}$ | $\begin{aligned} & D_{\text {MSY }} \\ & (1000 \text { fish }) \end{aligned}$ | $F / F_{\text {MSY }}$ | SSB/MSST | $\mathrm{SSB} / \mathrm{SSB}_{\mathrm{MSY}}$ | $h$ | $R_{0}(1000$ fish $)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Base | 0.180 | 2884 | 3605 | 531 | 127 | 104 | 1.73 | 0.35 | 0.27 | 0.38 | 3430 |
| S1 set M constant lo | 0.190 | 2728 | 3203 | 619 | 138 | 104 | 2.05 | 0.25 | 0.21 | 0.56 | 1801 |
| S2 set M constant up | 0.125 | 3872 | 5225 | 418 | 104 | 91 | 1.64 | 0.50 | 0.34 | 0.26 | 7179 |
| S3 set steep lo | 0.070 | 8438 | 10105 | 686 | 156 | 116 | 4.32 | 0.12 | 0.09 | 0.25 | 6348 |
| S4 set steep up | 0.325 | 1732 | 2261 | 485 | 134 | 119 | 0.94 | 0.59 | 0.46 | 0.51 | 2218 |
| S5 set log R0 lo | 0.555 | 1317 | 1810 | 503 | 170 | 165 | 0.56 | 0.75 | 0.58 | 0.71 | 1759 |
| S6 include sFT | 0.175 | 2947 | 3677 | 532 | 126 | 104 | 1.78 | 0.34 | 0.26 | 0.37 | 3475 |
| S7 FMatTimeVarying | 0.200 | 2788 | 3410 | 541 | 132 | 110 | 1.55 | 0.37 | 0.29 | 0.39 | 3253 |
| S8 wcpuerHb2 | 0.200 | 2419 | 3043 | 482 | 118 | 98 | 1.57 | 0.41 | 0.32 | 0.39 | 2825 |
| S9 wcpuerHb3 | 0.355 | 1615 | 2124 | 474 | 136 | 118 | 0.90 | 0.60 | 0.47 | 0.52 | 2024 |
| S10 smoothMRIP2016 | 0.175 | 2929 | 3660 | 547 | 116 | 95 | 1.38 | 0.36 | 0.28 | 0.38 | 3501 |


| 边 | Table 31. Projection results with fishing mortality rate fixed at $F=F_{\mathrm{P}_{50 \%}^{*}}$ starting in 2021 and projecting forward to 2026. From 2018 to 2020 the fishing mortality rate was fixed at $F_{\text {current }}=0.31 . R=$ number of age-1 recruits (in 1000s), $F=$ fishing mortality rate (per year), $S=$ spawning stock ( $m t$ ), $L=$ landings expressed in numbers ( $n$, in 1000s) or whole weight ( $w$, in 1000 lb ), and $D=$ dead discards expressed in numbers ( $n$, in 1000s) or whole weight ( $w$, in 1000 lb ), pr.reb $=$ proportion of stochastic projection replicates with $\mathrm{SSB} \geq \mathrm{SSB}_{\mathrm{MSY}}$. The extension b indicates expected values (deterministic) from the base run; the extension med indicates median values from the stochastic projections. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Year | R.b | R.med | F.b | F.med | S.b(mt) | S.med(mt) | L.b(n) | L.med(n) | L.b(w) | L.med(w) | D.b(n) | D.med(n) | D.b(w) | D.med(w) | pr.reb |
|  | 2018 | 1019 | 965 | 0.31 | 0.30 | 723 | 768 | 140 | 142 | 230 | 235 | 42 | 39 | 52 | 49 | 0.000 |
|  | 2019 | 956 | 919 | 0.31 | 0.30 | 720 | 768 | 123 | 128 | 200 | 208 | 47 | 43 | 53 | 50 | 0.000 |
|  | 2020 | 951 | 911 | 0.31 | 0.30 | 751 | 803 | 121 | 125 | 187 | 195 | 51 | 47 | 56 | 53 | 0.000 |
|  | 2021 | 987 | 935 | 0.18 | 0.18 | 800 | 852 | 80 | 85 | 120 | 129 | 32 | 31 | 35 | 35 | 0.001 |
|  | 2022 | 1040 | 981 | 0.18 | 0.18 | 880 | 936 | 93 | 98 | 139 | 148 | 34 | 33 | 39 | 38 | 0.001 |
|  | 2023 | 1126 | 1046 | 0.18 | 0.18 | 956 | 1012 | 102 | 108 | 155 | 165 | 37 | 35 | 42 | 41 | 0.001 |
|  | 2024 | 1205 | 1119 | 0.18 | 0.18 | 1032 | 1086 | 111 | 116 | 169 | 180 | 40 | 38 | 45 | 44 | 0.002 |
|  | 2025 | 1282 | 1178 | 0.18 | 0.18 | 1110 | 1163 | 119 | 125 | 183 | 193 | 42 | 41 | 49 | 47 | 0.003 |
|  | 2026 | 1358 | 1239 | 0.18 | 0.18 | 1189 | 1240 | 128 | 133 | 197 | 207 | 45 | 43 | 52 | 50 | 0.005 |




Table 34. Projection results with fishing mortality rate fixed at $F=0$ starting in 2021 and projecting forward to 2032. From 2018 to 2020 the fishing mortality rate was fixed at $F_{\text {current }}=0.31 . R=$ number of age- 1 recruits (in 1000s), $F=$ fishing mortality rate (per year), $S=$ spawning stock ( $m t$ ), $L=$ landings expressed in numbers ( $n$, in 1000s) or whole weight ( $w$, in 1000 lb ), and $D=$ dead discards expressed in numbers ( $n$, in 1000s) or whole weight ( $w$, in 1000 lb ), pr.reb = proportion of stochastic projection replicates with $\mathrm{SSB} \geq \mathrm{SSB}_{\mathrm{MSy}}$. The extension b indicates expected values (deterministic) from the base run; the extension med indicates median values from the stochastic projections.

| Year | R.b | R.med | F.b | F.med | S.b(mt) | S.med(mt) | L.b(n) | L.med(n) | L.b(w) | L.med(w) | D.b(n) | D.med(n) | D.b(w) | D.med(w) | pr.reb |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2018 | 1019 | 972 | 0.31 | 0.3 | 723 | 770 | 140 | 142 | 230 | 235 | 42 | 39 | 52 | 49 | 0.000 |
| 2019 | 956 | 918 | 0.31 | 0.3 | 720 | 771 | 123 | 128 | 200 | 208 | 47 | 43 | 53 | 50 | 0.000 |
| 2020 | 951 | 917 | 0.31 | 0.3 | 751 | 806 | 121 | 125 | 187 | 196 | 51 | 47 | 56 | 53 | 0.000 |
| 2021 | 987 | 946 | 0.00 | 0.0 | 813 | 870 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 |
| 2022 | 1054 | 998 | 0.00 | 0.0 | 963 | 1029 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 |
| 2023 | 1213 | 1147 | 0.00 | 0.0 | 1117 | 1190 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.005 |
| 2024 | 1366 | 1266 | 0.00 | 0.0 | 1283 | 1361 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.014 |
| 2025 | 1521 | 1395 | 0.00 | 0.0 | 1463 | 1544 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.031 |
| 2026 | 1679 | 1517 | 0.00 | 0.0 | 1658 | 1739 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.062 |
| 2027 | 1838 | 1652 | 0.00 | 0.0 | 1865 | 1945 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.115 |
| 2028 | 1997 | 1757 | 0.00 | 0.0 | 2085 | 2164 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.185 |
| 2029 | 2154 | 1897 | 0.00 | 0.0 | 2314 | 2387 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.278 |
| 2030 | 2306 | 2014 | 0.00 | 0.0 | 2550 | 2616 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.382 |
| 2031 | 2452 | 2092 | 0.00 | 0.0 | 2792 | 2849 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.485 |
| 2032 | 2591 | 2235 | 0.00 | 0.0 | 3036 | 3090 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.583 |

## 9 Figures

Figure 3. Length, female maturity, and reproductive output at age. Top panel: Mean length at age (mm) and estimated $95 \%$ confidence interval of the population. Middle panel: Female maturity by age. Bottom panel: Reproductive output (mt) by age.


Figure 4. Observed (open circles) and estimated (solid line) annual and and length compositions by fleet. In panels indicating the data set: acomp = age compositions, lcomp = length compositions, $c H l=$ commercial handline, $r H b=$ recreational headboat, $s C T=$ Chevron trap survey, $c T w=$ commercial trawl. $N$ indicates the number of trips from which individual fish samples were taken. The four digit number in upper right corner of each panel indicates year of sampling (e.g. 1983, 1984).


Figure 4. (cont.) Observed (open circles) and estimated (solid line) annual age and length compositions by fleet.





(2):










Figure 4. (cont.) Observed (open circles) and estimated (solid line) annual age and length compositions by fleet.


Figure 4. (cont.) Observed (open circles) and estimated (solid line) annual age and length compositions by fleet.


Figure 4. (cont.) Observed (open circles) and estimated (solid line) annual age and length compositions by fleet.


Figure 4. (cont.) Observed (open circles) and estimated (solid line) annual age and length compositions by fleet.


Figure 5. Observed (open circles) and estimated (line, solid circles) commercial handline landings (1000 lb whole weight).


Figure 6. Observed (open circles) and estimated (line, solid circles) commercial trawl landings (1000 lb whole weight).


Figure 7. Observed (open circles) and estimated (line, solid circles) recreational MRIP landings (1000 fish).


Figure 8. Observed (open circles) and estimated (line, solid circles) recreational headboat landings (1000 fish).


Figure 9. Observed (open circles) and estimated (line, solid circles) commercial handline discards (1000 fish).


Figure 10. Observed (open circles) and estimated (line, solid circles) recreational MRIP discards (1000 fish).


Figure 11. Observed (open circles) and estimated (line, solid circles) recreational headboat discards (1000 fish).


Figure 12. Observed (open circles) and estimated (line, solid circles) recreational headboat index of abundance .


Figure 13. Observed (open circles) and estimated (line, solid circles) Chevron trap/video survey index of abundance


Figure 14. Estimated abundance at age at start of year


Figure 15. Estimated biomass at age at start of year.


Figure 16. Estimated recruitment time series. Top panel: Estimated recruitment of age-1 fish. Horizontal dashed line indicates $R_{\mathrm{MSY}}$. Bottom panel: log recruitment residuals.



Figure 17. Estimated total biomass and spawning stock time series. Top panel: Estimated total biomass (metric tons) at start of year. Horizontal dashed line indicates $B_{\mathrm{MSY}}$. Bottom panel: Estimated spawning stock (mt) at time of peak spawning.



Figure 18. Selectivities of SERFS Chevron trap/video index. Different colored lines indicate different selectivity blocks. The first year of each selectivity block is indicated in the legend. In this case, there was only one selectivity block.


Figure 19. Selectivities of commercial handline landings (top) and trawl landings (bottom). Different colored lines indicate different selectivity blocks. The first year of each selectivity block is indicated in the legend.


Figure 20. Selectivities of recreational MRIP (top) and headboat landings (bottom). Different colored lines indicate different selectivity blocks. The first year of each selectivity block is indicated in the legend. In this assessment, selectivities of these two fleets mirrored each other


Figure 21. Selectivities of commercial handline (top), recreational MRIP (middle), and headboat discards (bottom) .




Figure 22. Average selectivity from the terminal assessment year weighted by geometric mean Fs from the last three assessment years, for landings (top), discards (middle), and total removals (bottom). These selectivities are used in computation of benchmarks and central-tendency projections.


Figure 23. Estimated fully selected fishing mortality rate (per year) by fleet. rGe. $D=$ recreational MRIP discards, $r H b . D=$ recreational headboat discards, cHl. $D=$ commercial handline discards, rGe $=$ recreational MRIP landings, $r H b=$ recreational headboat landings, $c T w=$ commercial trawl landings, $c H l=$ commercial handline landings.


Figure 24. Estimated landings in absolute numbers (top) and proportion of total numbers (bottom) by fleet from the catch-at-age model. $r G e=$ recreational MRIP landings, $r H b=$ recreational headboat landings, $c T w=$ commercial trawl landings, $\mathrm{cHl}=$ commercial handline landings.


Figure 25. Estimated landings in absolute weight (top) and proportion of total weight (bottom) by fleet from the catch-at-age model. $r G e=$ recreational MRIP landings, $r H b=$ recreational headboat landings, $c T w=$ commercial trawl landings, $\mathrm{cHl}=$ commercial handline landings.



| Fishery |  |
| :--- | :--- |
| $\square$ | rGe |
| $\square$ | rHb |
| $\square$ | cTw |
| $\square$ | cHI |

Figure 26. Estimated discards in absolute numbers (top) and proportion of total numbers (bottom) by fleet from the catch-at-age model. rGe = recreational MRIP discards, rHb $=$ recreational headboat discards, cHl $=$ commercial handline discards.


Figure 27. Estimated discards in absolute weight (top) and proportion of total weight (bottom) by fleet from the catch-at-age model. rGe = recreational MRIP discards, rHb $=$ recreational headboat discards, cHl $=$ commercial handline discards.


Figure 28. Beverton-Holt spawner-recruit curve (top) with and without lognormal bias correction. The expected (upper) curve was used for computing management benchmarks. Years within panel indicate year of recruitment generated from spawning biomass. Natural log of recruits (number of age-1 fish) per spawner is also plotted as function of the spawning stock (lower).



Figure 29. Probability densities of spawner-recruit quantities R0 (unfished recruitment of age-1 fish), steepness, unfished spawners per recruit, and standard deviation of recruitment residuals in log space. Solid vertical lines represent point estimates or values from the BAM base run; dashed vertical lines represent medians from the MCB runs ( $n=3350$ ).


Figure 30. Yield per recruit (top) and spawning potential ratio (bottom; spawning biomass per recruit relative to that at the unfished level) over a range of $F$. Both curves are based on average selectivity from the end of the assessment period.


Fishing mortality rate


Fishing mortality rate

Figure 31. The top panels shows equilibrium landings at $F$. The peak occurs where fishing rate is $F_{\mathrm{MSY}}=0.18$ and equilibrium landings are MSY $=531$ (1000 lb). The bottom panel shows equilibrium spawning biomass at $F$. Both curves are based on average selectivity from the end of the assessment period.


Figure 32. Probability densities of MSY-related benchmarks from MCB analysis ( $n=3350$ ). Vertical lines represent point estimates from the BAM base run; dashed vertical lines represent medians from the MCB runs.


Figure 33. Estimated time series of SSB and $F$ relative to benchmarks: (top) spawning biomass relative to the minimum stock size threshold (MSST), (middle) spawning biomass relative to $\mathrm{SSB}_{\mathrm{MSY}}$, and (bottom) F relative to $F_{\mathrm{MSY}}$. Shaded region represents $95 \%$ confidence bands from the MCB runs $(n=3350)$.


Figure 34. Probability densities of terminal status estimates from MCB analysis of the Beaufort Assessment Model ( $n=3350$ ). Vertical lines represent point estimates from the BAM base run; dashed vertical lines represent medians from the MCB runs.




Figure 35. Phase plot of terminal status estimates from MCB analysis of the Beaufort Assessment Model ( $n=3350$ ). The intersection of crosshairs indicates estimates from the BAM base run; lengths of crosshairs defined by $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of $M C B$ runs.


Figure 36. Age structure relative to the equilibrium expected at $F_{\mathrm{MSY}}$.


Figure 37. Sensitivity to low and high fixed values of natural mortality: sensitivity runs S1-S2. Estimated time series of $F$ and SSB relative to benchmarks, as well as biomass (B) and number of recruits. Solid line and solid circles indicate estimates from the BAM base run. Sensitivity runs are indicated by colored broken lines, represented in the legend. (top left) $F$ relative to $F_{\mathrm{MSY}}$. (top right) spawning stock biomass (SSB) relative to MSST. (bottom left) biomass ( $B$ ). (bottom right) number of recruits.


Figure 38. Sensitivity to low and high fixed values of steepness: sensitivity runs S3-S4. Estimated time series of $F$ and SSB relative to benchmarks, as well as biomass $(B)$ and number of recruits. Solid line and solid circles indicate estimates from the BAM base run. Sensitivity runs are indicated by colored broken lines, represented in the legend. (top left) $F$ relative to $F_{\mathrm{MSY}}$. (top right) spawning stock biomass (SSB) relative to MSST. (bottom left) biomass (B). (bottom right) number of recruits.





Figure 39. Sensitivity to a low value of R0 associated with the minimum likelihood of the age composition data likelihood component in likelihood profiling: sensitivity run S5. Estimated time series of $F$ and $S S B$ relative to benchmarks, as well as biomass ( $B$ ) and number of recruits. Solid line and solid circles indicate estimates from the $B A M$ base run. Sensitivity runs are indicated by colored broken lines, represented in the legend. (top left) $F$ relative to $F_{\mathrm{MSY}}$. (top right) spawning stock biomass (SSB) relative to MSST. (bottom left) biomass (B). (bottom right) number of recruits.


Figure 40. Sensitivity to including the MARMAP Florida Snapper Trap Index and associated age composition data: sensitivity run S6. Estimated time series of $F$ and $S S B$ relative to benchmarks, as well as biomass ( $B$ ) and number of recruits. Solid line and solid circles indicate estimates from the BAM base run. Sensitivity runs are indicated by colored broken lines, represented in the legend. (top left) $F$ relative to $F_{\mathrm{MSY}}$. (top right) spawning stock biomass (SSB) relative to MSST. (bottom left) biomass (B). (bottom right) number of recruits.


Figure 41. Sensitivity to including time-varying female maturity, similar to the previous assessment: sensitivity run S7. Estimated time series of $F$ and $S S B$ relative to benchmarks, as well as biomass (B) and number of recruits. Solid line and solid circles indicate estimates from the BAM base run. Sensitivity runs are indicated by colored broken lines, represented in the legend. (top left) $F$ relative to $F_{\mathrm{MSY}}$. (top right) spawning stock biomass (SSB) relative to MSST. (bottom left) biomass ( $B$ ). (bottom right) number of recruits.


Figure 42. Sensitivity to upweighting the headboat index by a factor of 2 or 3: sensitivity runs S8-S9. Estimated time series of $F$ and SSB relative to benchmarks, as well as biomass (B) and number of recruits. Solid line and solid circles indicate estimates from the BAM base run. Sensitivity runs are indicated by colored broken lines, represented in the legend. (top left) $F$ relative to $F_{\mathrm{MSY}}$. (top right) spawning stock biomass (SSB) relative to MSST. (bottom left) biomass ( $B$ ). (bottom right) number of recruits.


Figure 43. Sensitivity to 2016 MRIP landings and discards, investigated by replacing 2016 estimates with average values from 2015 and 2017: sensitivity run S10. Estimated time series of $F$ and $S S B$ relative to benchmarks, as well as biomass (B) and number of recruits. Solid line and solid circles indicate estimates from the BAM base run. Sensitivity runs are indicated by colored broken lines, represented in the legend. (top left) $F$ relative to $F_{\mathrm{MSY}}$. (top right) spawning stock biomass (SSB) relative to MSST. (bottom left) biomass (B). (bottom right) number of recruits.


Figure 44. Phase plots of terminal status estimates from BAM sensitivity runs. Point colors and shapes are indicated in the legend. The number of each sensitivity run is also plotted in black text over each point. The base run is represented by a black point labeled "base".


Figure 45. Retrospective analysis reducing the terminal year of the assessment from 2017 to values over a range from 2011 to 2016: . Estimated time series of $F$ and $S S B$ relative to benchmarks, as well as biomass ( $B$ ) and number of recruits. Solid line and solid circles indicate estimates from the BAM base run. Retrospective runs are indicated by colored broken lines, represented in the legend. (top left) $F$ relative to $F_{\text {MSY }}$. (top right) spawning stock biomass (SSB) relative to MSST. (bottom left) biomass (B). (bottom right) number of recruits.


Figure 46. Plots of SSB, landings, recruits, dead discards, F and the probability that SSB $>$ MSST for projections with fishing mortality rate at fixed $F$ that provides $P^{*}=0.50$. In all panels except the bottom right, expected values (base run) represented by solid lines with solid circles, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal blue lines mark MSY-related quantities; dashed horizontal green lines represent corresponding medians. Spawning stock (SSB) is at time of peak spawning. In the bottom right panel, the curve represents the proportion of projection replicates for which SSB has reached the replicate-specific MSST.


Figure 47. Plots of SSB, landings, recruits, dead discards, F and the probability that SSB $>$ MSST for projections with fishing mortality rate fixed at $F=F_{\mathrm{MSY}}$. In all panels except the bottom right, expected values (base run) represented by solid lines with solid circles, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{t h}$ and $95^{t h}$ percentiles of replicate projections. Solid horizontal blue lines mark MSY-related quantities; dashed horizontal green lines represent corresponding medians. Spawning stock (SSB) is at time of peak spawning. In the bottom right panel, the curve represents the proportion of projection replicates for which SSB has reached the replicate-specific MSST.


Figure 48. Plots of SSB, landings, recruits, dead discards, F and the probability that SSB $>$ MSST for projections with fishing mortality rate fixed at $F=75 \% F_{\mathrm{MSY}}$. In all panels except the bottom right, expected values (base run) represented by solid lines with solid circles, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal blue lines mark MSY-related quantities; dashed horizontal green lines represent corresponding medians. Spawning stock (SSB) is at time of peak spawning. In the bottom right panel, the curve represents the proportion of projection replicates for which SSB has reached the replicate-specific MSST.


Figure 49. Plots of SSB , landings, recruits, dead discards, $F$ and the probability that $\mathrm{SSB}>\mathrm{MSST}$ for projections with fishing mortality rate fixed at $F=0$. In all panels except the bottom right, expected values (base run) represented by solid lines with solid circles, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal blue lines mark MSYrelated quantities; dashed horizontal green lines represent corresponding medians. Spawning stock (SSB) is at time of peak spawning. In the bottom right panel, the curve represents the proportion of projection replicates for which SSB has reached the replicate-specific MSST.


## Appendix A Abbreviations and Symbols

Table 35. Acronyms and abbreviations used in this report

| Symbol | Meaning |
| :---: | :---: |
| ABC | Acceptable Biological Catch |
| AW | Assessment Workshop (here, for red porgy) |
| ASY | Average Sustainable Yield |
| $B$ | Total biomass of stock, conventionally on January 1 |
| BAM | Beaufort Assessment Model (a statistical catch-age formulation) |
| CPUE | Catch per unit effort; used after adjustment as an index of abundance |
| CV | Coefficient of variation |
| DW | Data Workshop (here, for red porgy) |
| $F$ | Instantaneous rate of fishing mortality |
| $F_{\text {MSY }}$ | Fishing mortality rate at which MSY can be attained |
| FL | State of Florida |
| GA | State of Georgia |
| GLM | Generalized linear model |
| K | Average size of stock when not exploited by man; carrying capacity |
| kg | Kilogram(s); 1 kg is about 2.2 lb . |
| klb | Thousand pounds; thousands of pounds |
| lb | Pound(s); 1 lb is about 0.454 kg |
| m | Meter(s); 1 m is about 3.28 feet. |
| M | Instantaneous rate of natural (non-fishing) mortality |
| MARMAP | Marine Resources Monitoring, Assessment, and Prediction Program, a fishery-independent data collection program of SCDNR |
| MCB | Monte Carlo/Bootstrap, an approach to quantifying uncertainty in model results |
| MFMT | Maximum fishing-mortality threshold; a limit reference point used in U.S. fishery management; often based on $F_{\mathrm{MSY}}$ |
| mm | Millimeter(s); 1 inch $=25.4 \mathrm{~mm}$ |
| MRFSS | Marine Recreational Fisheries Statistics Survey, a data-collection program of NMFS, predecessor of MRIP |
| MRIP | Marine Recreational Information Program, a data-collection program of NMFS, descended from MRFSS |
| MSST | Minimum stock-size threshold; a limit reference point used in U.S. fishery management. The SAFMC has defined MSST for red porgy as $(1-M) \mathrm{SSB}_{\mathrm{MSY}}=0.7 \mathrm{SSB}_{\mathrm{MSY}}$. |
| MSY | Maximum sustainable yield (per year) |
| mt | Metric ton(s). One mt is 1000 kg , or about 2205 lb . |
| $N$ | Number of fish in a stock, conventionally on January 1 |
| NC | State of North Carolina |
| NMFS | National Marine Fisheries Service, same as "NOAA Fisheries Service" |
| NOAA | National Oceanic and Atmospheric Administration; parent agency of NMFS |
| OY | Optimum yield; SFA specifies that OY $\leq$ MSY. |
| PSE | Proportional standard error |
| $R$ | Recruitment |
| SAFMC | South Atlantic Fishery Management Council (also, Council) |
| SC | State of South Carolina |
| SCDNR | Department of Natural Resources of SC |
| SDNR | Standard deviation of normalized residuals |
| SEDAR | SouthEast Data Assessment and Review process |
| SEFIS | SouthEast Fishery-Independent Survey |
| SERFS | SouthEast Reef Fish Survey |
| SFA | Sustainable Fisheries Act; the Magnuson-Stevens Act, as amended |
| SL | Standard length (of a fish) |
| SPR | Spawning potential ratio |
| SSB | Spawning stock biomass; mature biomass of males and females |
| $\mathrm{SSB}_{\text {MSY }}$ | Level of SSB at which MSY can be attained |
| TIP | Trip Interview Program, a fishery-dependent biodata collection program of NMFS |
| TL | Total length (of a fish), as opposed to FL (fork length) or SL (standard length) |
| VPA | Virtual population analysis, an age-structured assessment |
| WW | Whole weight, as opposed to GW (gutted weight) |
| yr | Year(s) |

## Appendix B Parameter estimates from the Beaufort Assessment Model

Table 36. Names and estimated values of parameters estimated in the base run of the Beaufort Assessment Model.

| ID | Parameter | Value |
| :---: | :---: | :---: |
| 1 | log.Nage.dev.age02 | -0.1178400 |
| 2 | log.Nage.dev.age03 | -0.1281600 |
| 3 | log.Nage.dev.age04 | -0.1178800 |
| 4 | log.Nage.dev.age05 | -0.1018900 |
| 5 | log.Nage.dev.age06 | -0.0851260 |
| 6 | log.Nage.dev.age07 | -0.0695820 |
| 7 | log.Nage.dev.age08 | -0.0559360 |
| 8 | log.Nage.dev.age09 | -0.0444580 |
| 9 | log.Nage.dev.age10 | -0.0350730 |
| 10 | log.Nage.dev.age11 | -0.0275950 |
| 11 | log.Nage.dev.age12 | -0.0215640 |
| 12 | log.Nage.dev.age13 | -0.0167420 |
| 13 | log.Nage.dev.age14 | -0.0533490 |
| 14 | log.R0 | 15.0480000 |
| 15 | steep | 0.3784400 |
| 16 | rec.sigma | 0.4547200 |
| 17 | log.rec.dev. 1975 | -0.4838600 |
| 18 | log.rec.dev. 1976 | 0.0113270 |
| 19 | log.rec.dev. 1977 | 1.3032000 |
| 20 | log.rec.dev. 1978 | 0.2069300 |
| 21 | log.rec.dev. 1979 | -0.4402700 |
| 22 | log.rec.dev. 1980 | 0.1365900 |
| 23 | log.rec.dev. 1981 | 0.0063075 |
| 24 | log.rec.dev. 1982 | -0.3472300 |
| 25 | log.rec.dev. 1983 | 0.1530400 |
| 26 | log.rec.dev. 1984 | 0.1887700 |
| 27 | log.rec.dev. 1985 | 0.3195300 |
| 28 | log.rec.dev. 1986 | 0.3060000 |
| 29 | log.rec.dev. 1987 | 0.2302100 |
| 30 | log.rec.dev. 1988 | 0.8133900 |
| 31 | log.rec.dev. 1989 | 0.5251900 |
| 32 | log.rec.dev. 1990 | 0.4351100 |
| 33 | log.rec.dev. 1991 | 0.3468100 |
| 34 | log.rec.dev. 1992 | 0.0046042 |
| 35 | log.rec.dev. 1993 | 0.0607640 |
| 36 | log.rec.dev. 1994 | 0.2412300 |
| 37 | log.rec.dev. 1995 | 0.6085900 |
| 38 | log.rec.dev. 1996 | -0.0943350 |
| 39 | log.rec.dev. 1997 | 0.0671870 |
| 40 | log.rec.dev. 1998 | 0.2144600 |

Table 36. (continued)

| ID | Parameter | Value |
| :---: | :---: | :---: |
| 41 | log.rec.dev. 1999 | -0.1440900 |
| 42 | log.rec.dev. 2000 | -0.7319700 |
| 43 | log.rec.dev. 2001 | -0.0027692 |
| 44 | log.rec.dev. 2002 | 0.3732900 |
| 45 | log.rec.dev. 2003 | 0.0373090 |
| 46 | log.rec.dev. 2004 | 0.0159200 |
| 47 | log.rec.dev. 2005 | 0.1598300 |
| 48 | log.rec.dev. 2006 | 0.2830400 |
| 49 | log.rec.dev. 2007 | -0.1894700 |
| 50 | log.rec.dev. 2008 | -0.6122300 |
| 51 | log.rec.dev. 2009 | -0.1780100 |
| 52 | log.rec.dev. 2010 | 0.3406900 |
| 53 | log.rec.dev. 2011 | 0.2702700 |
| 54 | log.rec.dev. 2012 | -0.2676300 |
| 55 | log.rec.dev. 2013 | -0.9199500 |
| 56 | log.rec.dev. 2014 | -0.5984200 |
| 57 | log.rec.dev. 2015 | -0.8643600 |
| 58 | log.rec.dev. 2016 | -0.5478900 |
| 59 | log.rec.dev. 2017 | -1.2371000 |
| 60 | log.dm.lenc.cTw | 3.0350000 |
| 61 | log.dm.agec.cHl | 4.3894000 |
| 62 | log.dm.agec.rHb | 3.0351000 |
| 63 | log.dm.agec.sCT | 1.7548000 |
| 64 | A50.sel.cHl2 | 1.8675000 |
| 65 | slope.sel.cHl2 | 5.8927000 |
| 66 | A50.sel.cHl3 | 2.9444000 |
| 67 | slope.sel.cHl3 | 2.5971000 |
| 68 | A50.sel.cTw | 0.8780300 |
| 69 | slope.sel.cTw | 4.1436000 |
| 70 | A50.sel.rHb1 | 1.1089000 |
| 71 | slope.sel.rHb1 | 7.9389000 |
| 72 | A50.sel.rHb2 | 2.0408000 |
| 73 | slope.sel.rHb2 | 3.5149000 |
| 74 | A50.sel.rHb3 | 3.6992000 |
| 75 | slope.sel.rHb3 | 1.9141000 |
| 76 | A50.sel.sCT | 2.3405000 |
| 77 | slope.sel.sCT | 1.3823000 |
| 78 | log.q.cpue.rHb | -15.1340000 |
| 79 | log.q.cpue.sCT | -14.1490000 |
| 80 | log.avg.F.L.cHl | -2.3228000 |

Table 36. (continued)

|  | ID | Parameter |
| :--- | :--- | ---: |
| 81 | log.F.dev.L.cHl.1972 | -3.1281000 |
| 82 | log.F.dev.L.cHl.1973 | -3.9212000 |
| 83 | log.F.dev.L.cHl.1974 | -1.9232000 |
| 84 | log.F.dev.L.cHl.1975 | -1.3176000 |
| 85 | log.F.dev.L.cHl.1976 | -1.1810000 |
| 86 | log.F.dev.L.cHl.1977 | -0.8172300 |
| 87 | log.F.dev.L.cHl.1978 | -0.1423500 |
| 88 | log.F.dev.L.cHl.1979 | 0.1204800 |
| 89 | log.F.dev.L.cHl.1980 | 0.2185500 |
| 90 | log.F.dev.L.cHl.1981 | 0.6771300 |
| 91 | log.F.dev.L.cHl.1982 | 0.9677800 |
| 92 | log.F.dev.L.cHl.1983 | 1.0812000 |
| 93 | log.F.dev.L.cHl.1984 | 1.1681000 |
| 94 | log.F.dev.L.cHl.1985 | 1.0919000 |
| 95 | log.F.dev.L.cHl.1986 | 1.2444000 |
| 96 | log.F.dev.L.cHl.1987 | 1.1637000 |
| 97 | log.F.dev.L.cHl.1988 | 1.4154000 |
| 98 | log.F.dev.L.cHl.1989 | 1.4389000 |
| 99 | log.F.dev.L.cHl.1990 | 1.6805000 |
| 100 | log.F.dev.L.cHl.1991 | 1.4310000 |
| 101 | log.F.dev.L.cHl.1992 | 0.9873300 |
| 102 | log.F.dev.L.cHl.1993 | 0.9739500 |
| 103 | log.F.dev.L.cHl.1994 | 0.9837200 |
| 104 | log.F.dev.L.cHl.1995 | 1.0032000 |
| 105 | log.F.dev.L.cHl.1996 | 0.8978500 |
| 106 | log.F.dev.L.cHl.1997 | 0.9345900 |
| 107 | log.F.dev.L.cHl.1998 | 0.6808000 |
| 108 | log.F.dev.L.cHl.1999 | -0.0695520 |
| 109 | log.F.dev.L.cHl.2000 | -1.6131000 |
| 110 | log.F.dev.L.cHl.2001 | -0.7631700 |
| 111 | log.F.dev.L.cHl.2002 | -0.7757000 |
| 112 | log.F.dev.L.cHl.2003 | -0.8673600 |
| 113 | log.F.dev.L.cHl.2004 | -1.0043000 |
| 114 | log.F.dev.L.cHl.2005 | -1.1397000 |
| 115 | log.F.dev.L.cHl.2006 | -0.6729900 |
| 116 | log.F.dev.L.cHl.2007 | -0.1782200 |
| 117 | log.F.dev.L.cHl.2008 | -0.0281710 |
| 118 | log.F.dev.L.cHl.2009 | -0.0645230 |
| 119 | log.F.dev.L.cHl.2010 | -0.0615060 |
| 120 | log.F.dev.L.cHl.2011 | 0.2016100 |
|  |  |  |

Table 36. (continued)

|  |  |  |
| :--- | :--- | ---: |
| ID | Parameter | Value |
| 121 | log.F.dev.L.cHl. 2012 | -0.1008900 |
| 122 | log.F.dev.L.cHl. 2013 | -0.1679200 |
| 123 | log.F.dev.L.cHl. 2014 | -0.2856400 |
| 124 | log.F.dev.L.cHl. 2015 | -0.2143400 |
| 125 | log.F.dev.L.cHl.2016 | -0.1203500 |
| 126 | log.F.dev.L.cHl. 2017 | 0.1960400 |
| 127 | log.avg.F.L.cTw | -5.6266000 |
| 128 | log.F.dev.L.cTw. 1972 | -3.8410000 |
| 129 | log.F.dev.L.cTw. 1973 | -0.8929900 |
| 130 | log.F.dev.L.cTw. 1974 | -3.8491000 |
| 131 | log.F.dev.L.cTw. 1975 | -3.2645000 |
| 132 | log.F.dev.L.cTw. 1976 | 0.3096500 |
| 133 | log.F.dev.L.cTw. 1977 | 1.4549000 |
| 134 | log.F.dev.L.cTw. 1978 | -1.6443000 |
| 135 | log.F.dev.L.cTw. 1979 | 0.8452400 |
| 136 | log.F.dev.L.cTw. 1980 | 2.2138000 |
| 137 | log.F.dev.L.cTw. 1981 | 2.3811000 |
| 138 | log.F.dev.L.cTw. 1982 | 2.2993000 |
| 139 | log.F.dev.L.cTw. 1983 | 1.8288000 |
| 140 | log.F.dev.L.cTw. 1984 | 1.3734000 |
| 141 | log.F.dev.L.cTw. 1985 | 0.1257900 |
| 142 | log.F.dev.L.cTw. 1986 | 0.1507900 |
| 143 | log.F.dev.L.cTw. 1987 | -0.2133400 |
| 144 | log.F.dev.L.cTw. 1988 | 0.7225100 |
| 145 | log.avg.F.L.rHb | -3.3390000 |
| 146 | log.F.dev.L.rHb. 1972 | -0.0236160 |
| 147 | log.F.dev.L.rHb. 1973 | 0.2668400 |
| 148 | log.F.dev.L.rHb. 1974 | -0.0450960 |
| 149 | log.F.dev.L.rHb. 1975 | -0.0107150 |
| 150 | log.F.dev.L.rHb. 1976 | -0.0664090 |
| 151 | log.F.dev.L.rHb. 1977 | -0.0661990 |
| 152 | log.F.dev.L.rHb. 1978 | -0.3952600 |
| 153 | log.F.dev.L.rHb. 1979 | -0.5809500 |
| 154 | log.F.dev.L.rHb. 1980 | -0.3248000 |
| 155 | log.F.dev.L.rHb. 1981 | -0.2139000 |
| 156 | log.F.dev.L.rHb. 1982 | 0.4878000 |
| 157 | log.F.dev.L.rHb. 1983 | 0.1265900 |
| 158 | log.F.dev.L.rHb. 1984 | 0.0238200 |
| 159 | log.F.dev.L.rHb. 1985 | 0.4094500 |
| 160 | log.F.dev.L.rHb. 1986 | 0.3616000 |
|  |  |  |

Table 36. (continued)

| ID | Parameter | Value |
| :---: | :---: | :---: |
| 161 | log.F.dev.L.rHb. 1987 | 0.5092000 |
| 162 | log.F.dev.L.rHb. 1988 | 0.4755600 |
| 163 | log.F.dev.L.rHb. 1989 | 0.2612200 |
| 164 | log.F.dev.L.rHb. 1990 | 0.0226940 |
| 165 | log.F.dev.L.rHb. 1991 | 0.3218300 |
| 166 | log.F.dev.L.rHb. 1992 | 0.4708400 |
| 167 | log.F.dev.L.rHb. 1993 | 0.5111400 |
| 168 | log.F.dev.L.rHb. 1994 | 0.4644200 |
| 169 | log.F.dev.L.rHb. 1995 | 0.4915300 |
| 170 | log.F.dev.L.rHb. 1996 | 0.3042400 |
| 171 | log.F.dev.L.rHb. 1997 | 0.1552700 |
| 172 | log.F.dev.L.rHb. 1998 | 0.2279400 |
| 173 | log.F.dev.L.rHb. 1999 | 0.4616900 |
| 174 | log.F.dev.L.rHb. 2000 | -1.0465000 |
| 175 | log.F.dev.L.rHb. 2001 | 0.1252000 |
| 176 | log.F.dev.L.rHb. 2002 | -0.1184800 |
| 177 | log.F.dev.L.rHb. 2003 | -0.0643670 |
| 178 | log.F.dev.L.rHb. 2004 | 0.0147100 |
| 179 | log.F.dev.L.rHb. 2005 | -0.0771180 |
| 180 | log.F.dev.L.rHb. 2006 | 0.3059400 |
| 181 | log.F.dev.L.rHb. 2007 | 0.8941700 |
| 182 | log.F.dev.L.rHb. 2008 | 0.0514170 |
| 183 | log.F.dev.L.rHb. 2009 | -0.4739300 |
| 184 | log.F.dev.L.rHb. 2010 | -0.3482400 |
| 185 | log.F.dev.L.rHb. 2011 | -0.3449800 |
| 186 | log.F.dev.L.rHb. 2012 | -0.2881000 |
| 187 | log.F.dev.L.rHb. 2013 | -0.6852400 |
| 188 | log.F.dev.L.rHb. 2014 | -0.7904400 |
| 189 | log.F.dev.L.rHb. 2015 | -0.8239500 |
| 190 | log.F.dev.L.rHb. 2016 | -0.5285100 |
| 191 | log.F.dev.L.rHb. 2017 | -0.4282500 |
| 192 | log.avg.F.L.rGe | -3.3895000 |
| 193 | log.F.dev.L.rGe. 1972 | -0.9655600 |
| 194 | log.F.dev.L.rGe. 1973 | -0.9844900 |
| 195 | log.F.dev.L.rGe. 1974 | -0.9868900 |
| 196 | log.F.dev.L.rGe. 1975 | -0.9325900 |
| 197 | log.F.dev.L.rGe. 1976 | -0.8445100 |
| 198 | log.F.dev.L.rGe. 1977 | -1.1099000 |
| 199 | log.F.dev.L.rGe. 1978 | -1.3542000 |
| 200 | log.F.dev.L.rGe. 1979 | -1.1828000 |

Table 36. (continued)

|  |  | ID |
| :--- | :--- | ---: |
| Parameter | Value |  |
| 201 | log.F.dev.L.rGe. 1980 | -0.9999800 |
| 202 | log.F.dev.L.rGe. 1981 | -1.2612000 |
| 203 | log.F.dev.L.rGe. 1982 | -1.0663000 |
| 204 | log.F.dev.L.rGe. 1983 | -1.8138000 |
| 205 | log.F.dev.L.rGe. 1984 | 0.7608700 |
| 206 | log.F.dev.L.rGe. 1985 | 0.5626100 |
| 207 | log.F.dev.L.rGe. 1986 | -1.0777000 |
| 208 | log.F.dev.L.rGe. 1987 | -0.3409400 |
| 209 | log.F.dev.L.rGe. 1988 | 0.7393200 |
| 210 | log.F.dev.L.rGe. 1989 | 0.2556300 |
| 211 | log.F.dev.L.rGe. 1990 | 0.0983180 |
| 212 | log.F.dev.L.rGe. 1991 | -0.3901100 |
| 213 | log.F.dev.L.rGe. 1992 | 1.1341000 |
| 214 | log.F.dev.L.rGe. 1993 | 0.1583500 |
| 215 | log.F.dev.L.rGe. 1994 | 0.5139200 |
| 216 | log.F.dev.L.rGe. 1995 | 0.0889670 |
| 217 | log.F.dev.L.rGe. 1996 | 0.3715000 |
| 218 | log.F.dev.L.rGe. 1997 | -0.7485700 |
| 219 | log.F.dev.L.rGe. 1998 | -0.2707400 |
| 220 | log.F.dev.L.rGe. 1999 | 0.3939500 |
| 221 | log.F.dev.L.rGe. 2000 | -0.7232500 |
| 222 | log.F.dev.L.rGe. 2001 | 0.5111700 |
| 223 | log.F.dev.L.rGe. 2002 | 0.9455700 |
| 224 | log.F.dev.L.rGe. 2003 | 1.0028000 |
| 225 | log.F.dev.L.rGe. 2004 | 1.2214000 |
| 226 | log.F.dev.L.rGe. 2005 | 0.6782400 |
| 227 | log.F.dev.L.rGe. 2006 | 0.0613540 |
| 228 | log.F.dev.L.rGe. 2007 | 0.6258400 |
| 229 | log.F.dev.L.rGe. 2008 | 1.2936000 |
| 230 | log.F.dev.L.rGe. 2009 | 0.5746100 |
| 231 | log.F.dev.L.rGe. 2010 | -0.0098104 |
| 232 | log.F.dev.L.rGe. 2011 | 0.6118500 |
| 233 | log.F.dev.L.rGe. 2012 | 0.5237300 |
| 234 | log.F.dev.L.rGe. 2013 | 0.0678810 |
| 235 | log.F.dev.L.rGe. 2014 | -0.4171200 |
| 236 | log.F.dev.L.rGe. 2015 | 0.6704200 |
| 237 | log.F.dev.L.rGe. 2016 | 2.4057000 |
| 238 | log.F.dev.L.rGe. 2017 | 1.2088000 |
| 239 | log.F.dev.D.FHl.cHl 1999 | -4.4329000 |
| 240 | 0.8968800 |  |
|  |  |  |

Table 36. (continued)

| ID | Parameter | Value |
| :---: | :---: | :---: |
| 2 | log.F.dev.D.cHl. 2000 | 0.9572200 |
| 242 | log.F.dev.D.cHl. 2001 | 1.0157000 |
| 243 | log.F.dev.D.cHl. 2002 | 2.1717000 |
| 244 | log.F.dev.D.cHl. 2003 | 0.2927100 |
| 245 | log.F.dev.D.cHl. 2004 | 0.0668620 |
| 246 | log.F.dev.D.cHl. 2005 | -0.4085700 |
| 247 | log.F.dev.D.cHl. 2006 | -0.0227870 |
| 248 | log.F.dev.D.cHl. 2007 | -0.5820100 |
| 249 | log.F.dev.D.cHl. 2008 | -0.0626020 |
| 250 | log.F.dev.D.cHl. 2009 | -0.1028300 |
| 251 | log.F.dev.D.cHl. 2010 | -0.5642100 |
| 252 | log.F.dev.D.cHl. 2011 | -1.2876000 |
| 253 | log.F.dev.D.cHl. 2012 | -0.5606100 |
| 254 | log.F.dev.D.cHl. 2013 | -0.5834600 |
| 255 | log.F.dev.D.cHl. 2014 | -0.3657900 |
| 256 | log.F.dev.D.cHl. 2015 | -0.1600300 |
| 257 | log.F.dev.D.cHl. 2016 | -0.4867300 |
| 258 | log.F.dev.D.cHl. 2017 | -0.2138300 |
| 259 | log.avg.F.D.rHb | -5.4211000 |
| 260 | log.F.dev.D.rHb. 2001 | 0.8871300 |
| 261 | log.F.dev.D.rHb. 2002 | 0.4635100 |
| 262 | log.F.dev.D.rHb. 2003 | 0.3052400 |
| 263 | log.F.dev.D.rHb. 2004 | 1.0246000 |
| 264 | log.F.dev.D.rHb. 2005 | -0.2501100 |
| 265 | log.F.dev.D.rHb. 2006 | 0.4734800 |
| 266 | log.F.dev.D.rHb. 2007 | 0.4195500 |
| 267 | log.F.dev.D.rHb. 2008 | 0.1171100 |
| 268 | log.F.dev.D.rHb. 2009 | -0.4173300 |
| 269 | log.F.dev.D.rHb. 2010 | -0.6517000 |
| 270 | log.F.dev.D.rHb. 2011 | -0.6499100 |
| 271 | log.F.dev.D.rHb. 2012 | -0.5627800 |
| 272 | log.F.dev.D.rHb. 2013 | -0.6399200 |
| 273 | log.F.dev.D.rHb. 2014 | -0.2412500 |
| 274 | log.F.dev.D.rHb. 2015 | -0.0711900 |
| 275 | log.F.dev.D.rHb. 2016 | -0.0421180 |
| 276 | log.F.dev.D.rHb. 2017 | -0.1643100 |
| 277 | log.avg.F.D.rGe | -6.2704000 |
| 278 | log.F.dev.D.rGe. 1981 | -3.0696000 |
| 279 | log.F.dev.D.rGe. 1982 | -2.4047000 |
| 280 | log.F.dev.D.rGe. 1983 | -3.2250000 |

Table 36. (continued)

|  |  | ID |
| :--- | :--- | ---: |
| 281 | log.F.dev.D.rGe. 1984 | Value |
| 282 | log.F.dev.D.rGe. 1985 | -0.011848000 |
| 283 | log.F.dev.D.rGe. 1986 | -3.2783000 |
| 284 | log.F.dev.D.rGe. 1987 | 0.5098500 |
| 285 | log.F.dev.D.rGe. 1988 | -3.0139000 |
| 286 | log.F.dev.D.rGe.1989 | 0.3616600 |
| 287 | log.F.dev.D.rGe. 1990 | -4.1982000 |
| 288 | log.F.dev.D.rGe. 1991 | -4.1127000 |
| 289 | log.F.dev.D.rGe.1992 | 0.1688500 |
| 290 | log.F.dev.D.rGe. 1993 | -0.1613800 |
| 291 | log.F.dev.D.rGe. 1994 | -1.2591000 |
| 292 | log.F.dev.D.rGe. 1995 | 0.8215800 |
| 293 | log.F.dev.D.rGe. 1996 | -0.3566000 |
| 294 | log.F.dev.D.rGe. 1997 | -1.6018000 |
| 295 | log.F.dev.D.rGe. 1998 | -0.2882000 |
| 296 | log.F.dev.D.rGe. 1999 | 1.9218000 |
| 297 | log.F.dev.D.rGe. 2000 | 1.0056000 |
| 298 | log.F.dev.D.rGe. 2001 | 2.7250000 |
| 299 | log.F.dev.D.rGe. 2002 | 1.9803000 |
| 300 | log.F.dev.D.rGe. 2003 | 2.3069000 |
| 301 | log.F.dev.D.rGe. 2004 | 2.2386000 |
| 302 | log.F.dev.D.rGe. 2005 | 1.6084000 |
| 303 | log.F.dev.D.rGe. 2006 | -0.3401900 |
| 304 | log.F.dev.D.rGe. 2007 | 1.4707000 |
| 305 | log.F.dev.D.rGe. 2008 | 2.3649000 |
| 306 | log.F.dev.D.rGe. 2009 | 0.1392000 |
| 307 | log.F.dev.D.rGe. 2010 | 0.7838100 |
| 308 | log.F.dev.D.rGe. 2011 | 0.5734000 |
| 309 | log.F.dev.D.rGe. 2012 | -0.3743100 |
| 310 | log.F.dev.D.rGe. 2013 | 0.2469200 |
| 311 | log.F.dev.D.rGe. 2014 | 1.2979000 |
| 312 | log.F.dev.D.rGe. 2015 | 2.0510000 |
| 313 | log.F.dev.D.rGe. 2016 | 3.6758000 |
| 314 | log.F.dev.D.rGe. 2017 | 1.8486000 |
|  |  |  |

