# Stock Assessment of Golden Tilefish off the Southeastern United States 2016 SEDAR Update Assessment 



Southeast Fisheries Science Center
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## Contents

2 Executive Summary ..... 9
3 Data Review and Updates ..... 10
3.1 Life History ..... 10
3.2 Landings ..... 10
3.3 Discards ..... 11
3.4 Indices of abundance ..... 11
3.5 Length Composition ..... 11
3.6 Age Composition ..... 11
4 Stock Assessment Model and Results ..... 12
4.1 Model Methods ..... 12
4.1.1 Overview ..... 12
4.1.2 Data Sources ..... 12
4.1.3 Base Model Configuration ..... 12
4.1.4 Per Recruit and Equilibrium Analyses ..... 14
4.1.5 Biological Reference Points ..... 15
4.1.6 Sensitivity and Retrospective Analyses ..... 15
4.1.7 Uncertainty and Measures of Precision ..... 16
4.1.8 Projection Methods ..... 17
4.1.9 Acceptable Biological Catch ..... 18
4.2 Model Results ..... 18
4.2.1 Base Run Results ..... 18
4.2.2 Per Recruit and Equilibrium Analyses ..... 19
4.2.3 Benchmarks/Reference Points ..... 20
4.2.4 Status of the Stock and Fishery ..... 20
4.2.5 Sensitivity and Retrospective Analyses ..... 20
4.2.6 Comparison with Previous Assessment ..... 21
4.2.7 Projections ..... 21
5 Discussion ..... 21
5.1 Comments on Projections ..... 21
5.2 Recommendations for the Next Benchmark Assessment ..... 22
6 References ..... 23
7 Tables ..... 25
8 Figures ..... 59
Appendices ..... 109
A Abbreviations and Symbols ..... 109
B Terms of Reference ..... 110
C ADMB Parameter Estimates ..... 111

## List of Tables

7.1 Life history model inputs ..... 25
7.2 Time series of landings in weight by fleet ..... 26
7.2 (Continued) Time series of landings in weight by fleet ..... 27
7.3 Time series of estimated recreational landings in numbers and commerical landings in numbers by gear ..... 28
7.4 Observed time series of indices of abundance ..... 29
7.5 Sample sizes for length and age compositions by fleet and survey ..... 30
7.6 Length compositions by fleet ..... 31
7.6 (Continued) Length compositions by fleet ..... 32
7.7 Annual proportion at age from commercial samples ..... 33
7.7 (Continued) Annual proportion at age from commercial samples ..... 34
7.8 Annual proportion at age from the MARMAP survey ..... 35
7.8 (Continued) Annual proportion at age from the MARMAP survey ..... 36
7.9 Estimated total abundance at age (1000 fish) ..... 37
7.9 (Continued) Estimated total abundance at age (1000 fish) ..... 38
7.10 Estimated biomass at age ( 1000 lb ) ..... 39
7.10 (Continued) Estimated biomass at age (1000 lb) ..... 40
7.11 Estimated time series of status indicators, fishing mortality, and biomass ..... 41
7.12 Selectivity at age (terminal year) by fleet or survey ..... 42
7.13 Estimated instantaneous fishing mortality rate (per yr) at age ..... 43
7.13 (Continued) Estimated instantaneous fishing mortality rate (per yr) at age. ..... 44
7.14 Estimated time series of fully selected fishing mortality rates by fleet ..... 45
7.15 Estimated time series of landings in gutted weight ( 1000 lb ) ..... 46
7.16 Estimated time series of landings in numbers (1000 fish) ..... 47
7.17 Estimated landings at age in gutted weight ( 1000 lb ) ..... 48
7.17 (Continued) Estimated landings at age in gutted weight (1000 lb) ..... 49
7.18 Estimated total landings at age in numbers (1000 fish) ..... 50
7.18 (Continued) Estimated total landings at age in numbers (1000 fish) ..... 51
7.19 Estimated status indicators and benchmarks ..... 52
7.20 Results from sensitivity runs of the Beaufort catch-at-age model ..... 53
7.21 Projection results for $F=$ Fcurrent ..... 54
7.22 Projection results for $F=F_{\mathrm{MSY}}$ ..... 55
7.23 Projection results for $F=75 \% F_{\mathrm{MSY}}$ ..... 56
7.24 Projection results for $P *=0.35$ ..... 57
7.25 Projection results for $P *=0.5$ ..... 58
A. 1 Abbreviations and Symbols ..... 109

## List of Figures

8.1 Mean length at age (mm) and estimated $95 \%$ confidence interval of the population. ..... 60
8.2 Observed and estimated commercial handline landings ..... 61
8.3 Observed and estimated commercial longline landings ..... 62
8.4 Observed and estimated recreational landings ..... 63
8.5 Observed and estimated index of abundance from the commercial longline fishery ..... 64
8.6 Observed and estimated index of abundance from the MARMAP horizontal longline survey. ..... 65
8.7 Observed and estimated annual length and age compositions by fleet or survey. ..... 66
8.7 (Continued) Observed and estimated annual length and age compositions by fleet or survey. ..... 67
8.7 (Continued) Observed and estimated annual length and age compositions by fleet or survey. ..... 68
8.7 (Continued) Observed and estimated annual length and age compositions by fleet or survey ..... 69
8.7 (Continued) Observed and estimated annual length and age compositions by fleet or survey. ..... 70
8.8 Estimated abundance at age at start of year ..... 71
8.9 Estimated recruitment of age-1 fish. ..... 72
8.10 Estimated biomass at age at start of year. ..... 73
8.11 Estimated total and spawning stock biomass at start of year. ..... 74
8.12 Selectivities of commercial fleets ..... 75
8.13 Selectivities of the recreational fleet and MARMAP survey ..... 76
8.14 Average selectivity from the terminal assessment year ..... 77
8.15 Estimated fully selected fishing mortality rate by fishery ..... 78
8.16 Estimated landings in gutted weight ..... 79
8.17 Estimated landings in numbers by fishery ..... 80
8.18 Spawner-recruit curves ..... 81
8.19 Probability densities of spawner-recruit quantities from MCB analysis ..... 82
8.20 Estimated time series static spawners per recruit ..... 83
8.21 Yield per recruit and spawning potential ratio ..... 84
8.22 Equilibrium landings and SSB as a function of fishing mortality rate ..... 85
8.23 Equilibrium landings as a function of equilibrium biomass ..... 86
8.24 Probability densities of MSY-related benchmarks from MCB analysis ..... 87
8.25 Estimated time series relative to benchmarks ..... 88
8.26 Probability densities of terminal status estimates from MCB analysis ..... 89
8.27 Phase plot of terminal status estimates from MCB analysis ..... 90
8.28 Sensitivity of model to configuration changes ..... 91
8.29 Retrospective analyses - Fishing mortality rate ..... 92
8.30 Retrospective analyses - Biomass ..... 93
8.31 Retrospective analyses - Spawning stock biomass ..... 94
8.32 Retrospective analyses - Recruitment ..... 95
8.33 Retrospective analyses - Relative fishing mortality rate ..... 96
8.34 Retrospective analyses - Relative spawning stock biomass ..... 97
8.35 Comparison of base and continuity runs ..... 98
8.36 Projection results for $\operatorname{SSB} B, N$, and $R$ at $F=F_{\text {current }}$ ..... 99
8.37 Projection results for $F, L$, and probability $S S B>\operatorname{MSST}$ at $F=F_{\text {current }}$ ..... 100
8.38 Projection results for $\operatorname{SSB} B, N$, and $R$ at $F=F_{\mathrm{MSY}}$ ..... 101
8.39 Projection results for $F, L$, and probability $S S B>\mathrm{MSST}$ at $F=F_{\mathrm{MSY}}$ ..... 102
8.40 Projection results for $\operatorname{SSB} B, N$, and $R$ at $F=75 \% F_{\text {MSY }}$ ..... 103
8.41 Projection results for $F, L$, and probability $S S B>\mathrm{MSST}$ at $F=75 \% F_{\mathrm{MSY}}$ ..... 104
8.42 Projection results for $\operatorname{SSB} B, N$, and $R$ at $P *=0.35$ ..... 105
8.43 Projection results for $F, L$, and probability $S S B>\operatorname{MSST}$ at $P *=0.35$ ..... 106
8.44 Projection results for $\operatorname{SSB} B, N$, and $R$ at $P *=0.5$ ..... 107
8.45 Projection results for $F, L$, and probability $S S B>\operatorname{MSST}$ at $P *=0.5$ ..... 108

## 2 Executive Summary

This assessment provides an update to the SEDAR- $25^{1}$ benchmark assessment of tilefish (Lopholatilus chamaeleonticeps) off the southeastern United States. The primary objectives were to update the benchmark assessment of tilefish with recent data and conduct new stock projections as outlined in the Terms of Reference (Appendix B). For this assessment, data compilation and modeling methods were guided by the methodology of SEDAR-25 as well as more recent SEDAR assessments.

The assessment period was 1962-2014. Available data on this stock include indices of abundance, landings, and samples of annual length and age compositions from fishery dependent and independent sources. Two indices of abundance were fitted by the model: one from the commercial longline fleet, and one from a fishery-independent survey. Data on landings were available from commercial and recreational fleets.

The primary model used in SEDAR-25—updated here—was the Beaufort Assessment Model (BAM), a statistical catch-age formulation. A base run of BAM was configured to provide estimates of key management quantities, such as stock and fishery status. Uncertainty in estimates from the base run was evaluated through a mixed Monte Carlo/Bootstrap (MCB) procedure. Median values from the uncertainty analysis were also provided. Stock status was evaluated by measuring the 2014 spawning biomass against the minimum stock size threshold (MSST). The current definition of MSST is $\mathrm{MSST}=75 \% \mathrm{SSB}_{\mathrm{MSY}}$.

Spawning stock declined in the 1980s, remained low but stable from the mid-1990s to the mid-2000s, then increased over the last decade, rising above MSST in the last seven years of the assessment. The terminal (2014) base-run estimate of spawning stock biomass was above MSST (base: $\mathrm{SSB}_{2014} / \mathrm{MSST}=1.13$; median of MCBs: $\mathrm{SSB}_{2014} / \mathrm{MSST}=1.04$ ), but below $\mathrm{SSB}_{\mathrm{MSY}}$ (base: $\mathrm{SSB}_{2014} / \mathrm{SSB}_{\mathrm{MSY}}=0.85$; median of MCBs : $\mathrm{SSB}_{2014} / \mathrm{SSB}_{\mathrm{MSY}}=0.83$ ).

Estimated fishing mortality rates began increasing in the early 1980s, peaked in the early 1990s, displayed another smaller peak around 2000, then declined steadily until 2012 when rates began to increase again. The base-run estimate of fishing mortality $(F)$, represented by the geometric mean of the last three years (2012-2014), exceeded the MFMT $\left(F_{2012-2014} / F_{\mathrm{MSY}}=1.22\right)$ as did the median of the MCB estimates $\left(F_{2012-2014} / F_{\mathrm{MSY}}=1.43\right)$.

Thus, this assessment finds that the stock is experiencing overfishing, but is not overfished.
The MCB analysis indicated that estimates of stock and fishery status were robust, but also revealed some quantitative uncertainty in the results. Among all MCB runs, $53 \%$ are in qualitative agreement that the stock is not overfished $\left(\mathrm{SSB}_{2014} / \mathrm{MSST}>1.0\right)$, and $66 \%$ are in qualitative agreement that the stock is experiencing overfishing $\left(F_{2012-2014} / F_{\mathrm{MSY}}>1.0\right)$. Over $47 \%$ of MCB runs indicated that the stock was both overfished and that overfishing was occurring.

The estimated trends of this SEDAR-25-Update assessment are similar to those from the SEDAR-25 benchmark. However, the two assessments did show some differences in results, which was not surprising given several modifications made to both the data and model (described throughout the report). Of those modifications, the use of a robust multinomial likelihood function to fit age and length composition data had the greatest effect. Compared to SEDAR-25, this assessment suggests lower values of $\mathrm{SSB}_{\mathrm{MSY}}$ and MSY, and a higher value of $F_{\mathrm{MSY}}$.

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## 3 Data Review and Updates

In the SEDAR-25 benchmark, the assessment period was 1962-2010. In this update, the terminal year was extended to 2014. For some data sources, the data were simply updated with the additional four years. However, for other sources, data prior to 2011 were updated as necessary. The input data for this assessment are summarized in Tables $7.1-7.8$ and are described below with a focus on the data that required modification from SEDAR- 25.

In this update assessment, the Beaufort Assessment Model (BAM) was fitted to the same data sources as in SEDAR-25:

- Landings: commercial handline, commercial longline, and general recreational
- Indices of abundance: commercial longline fishery CPUE and MARMAP longline survey CPUE
- Length compositions of landings: commercial handline, commercial longline, and general recreational
- Age compositions of landings: commercial handline, commercial longline, and MARMAP longline survey.


### 3.1 Life History

Life-history inputs from SEDAR-25 remained the same in this assessment, including all growth parameters, natural mortality at age, female maturity and fecundity at age, the gutted weight to whole weight conversion, and other related inputs. Primary life-history information is summarized in Table 7.1.

### 3.2 Landings

Landings estimates were combined into three fleets: commercial handline, commercial longline, and general recreational. Estimates of commercial landings were updated for 1962-2014 using the methods outlined in SEDAR-25 (SEDAR 2011). An "other" category in the reported landings was distributed by year between handlines and longlines based on the yearly ratio of handline to longline landings as in SEDAR-25. Commercial handline and longline estimated landings were input as thousands of gutted pounds (Table 7.2) and converted to whole pounds in the model using the equation

$$
\text { Wholeweight }=1.059 * \text { Guttedweight }
$$

Commercial landings in numbers for longline and handline gears were estimated using updated landed pounds and average weights by stratum following the methodology outlined in SEDAR-25 (Table 7.3). Average weight for fish caught in the "other" gear category was assumed to be the same as SEDAR-25.

Inconsistent data reporting precluded reliable characterization of trends in tilefish landings by market category. Commercial landings reported by market category are not available prior to 1988 and are inconsistently available after that year. From 1989 to 2014, landings assigned "unknown market category" ranged from $8-99 \%$ of the data by pounds in a given year, with $46.7 \%$ of all pounds in that time period having an unknown market category.

In this assessment, recreational landings included estimates of the Southeast Region Headboat Survey headboat landings and the Marine Recreational Information Program (MRIP) private and charter landings in thousands of whole pounds (Table 7.2). Landings in numbers of fish are reported in Table 7.3. Using Marine Recreational Fisheries Statistics (MRFSS) landings estimates, SEDAR-25 considered the 2005 estimate to be unrealistically high and replaced it with the average of the 2003, 2004, 2006, and 2007 estimated landings. Revised MRIP estimates for 2005 were significantly lower and on par with adjacent years; therefore, this assessment used the reported MRIP estimate of recreational landings in 2005. In years with no recreational landings estimated, a small value of 0.02 ( 1000 lb ww ) was assumed as in SEDAR-25.

### 3.3 Discards

As in SEDAR-25, no discard estimates were included in the model as discards are assumed to be negligible in all sectors of the tilefish fishery.

### 3.4 Indices of abundance

The indices of abundance used in SEDAR-25 included the fishery-independent MARMAP horizontal longline index and the fishery dependent longline logbook index (Table 7.4). Both indices were updated for this assessment with the available data (2011-2014 for commercial logbook; 2011 for MARMAP given suspension of this data collection program). MARMAP CPUE from 2011 was pooled with 2009 and 2010 to generate the " 2010 " point estimate.

### 3.5 Length Composition

Length compositions were developed from the commercial handline, commercial longline, and recreational sampling data. Sample sizes (number of trips) are reported in Table 7.5. Following the methodology of SEDAR-25, the contribution of each length was weighted by the landings associated by state, gear, and year. In this assessment, length composition data for the commercial fishery was recalculated in some years because of small corrections in historical landings used as weights (Table 7.6). These corrections resulted in only slight changes in length compositions from SEDAR-25.

Recreational samples included a combination of headboat and MRIP samples. In SEDAR-25, the recreational headboat data query mistakenly included commercial and research samples. In this assessment, only recreational samples from the headboat survey were selected for use. As a result, a more appropriate set of samples were used to characterize the fishery in this assessment. However, the total number of recreational length samples was reduced. The number of trips (used to characterize effective sample size) fell from 85 to 75 , and the number of fish sampled fell from 578 to 382 . Proportions at length were re-calculated and included in this assessment if the number of trips in a given year was greater than two and the number of fish sampled was greater than ten (Table 7.6). The number of trips associated with recreational data are either vessel trips or angler trips depending on the year (Table 7.5).

### 3.6 Age Composition

Age data were available from the commercial handline, commercial longline, and MARMAP longline sampling programs. Ages greater than 25 were pooled to age 25 creating a plus group. The annual commercial age compositions developed for tilefish were updated with data from 2011 to 2014 for both fleets (Table 7.7). Sample sizes are reported in Tables 7.5 and 7.7.

All years of MARMAP age composition data were recalculated given an error identified in data treatment from SEDAR-25. Limited samples were available from the MARMAP longline index (Table 7.8); therefore, data were pooled across years as in SEDAR-25. Following the benchmark's methodology, 2011 samples were pooled with samples collected in 2009-2011 to generate an updated "2010" age composition. As in SEDAR-25, MARMAP longline standard deviation estimates provided were converted to CV (CPUE/SD) for input to the model.

## 4 Stock Assessment Model and Results

### 4.1 Model Methods

### 4.1.1 Overview

The Beaufort assessment model (BAM) developed for tilefish in SEDAR-25 was updated in this assessment. The BAM applies a statistical catch-age formulation (Williams and Shertzer 2015) and was implemented with the AD Model Builder software (ADMB Foundation 2011).

### 4.1.2 Data Sources

The catch-age model included data from a fishery dependent survey, a fishery independent survey, and three fleets that caught southeastern U.S. tilefish: commercial longline, commercial handline, and the recreational fishery. The model was fitted to annual landings, annual length compositions of landings, annual age compositions of landings, and two indices of abundance (MARMAP longline and the commercial longline logbook).

Data used in the model are described and tabulated in Section 3 of this report. One exception is that longline age composition data from 1996, 1998, and 1999 were removed from model inputs after data compilation in response to concerns raised in the SEDAR-25 DW report about how age samples appeared to be collected from the largest fish in the catch in those years (as opposed to random sample collection across all lengths of fish encountered by the port agents). This method of data collection generated biased age composition data. In this update assessment, 1996, 1998, and 1999 longline age composition data were removed from the model inputs.

### 4.1.3 Base Model Configuration

Model configuration was identical to SEDAR-25 with the following three exceptions discussed in detail below:

- Use of a robust multinomial likelihood
- Estimation of recreational fleet selectivity
- Removal of biased commercial longline age composition data from 1996, 1998, and 1999.

A general description of the base run configuration follows.
Stock Dynamics In the assessment model, new biomass was acquired through growth and recruitment, while abundance of existing cohorts experienced exponential decay from fishing and natural mortality. The population was assumed closed to immigration and emigration. The model included age classes $1-25^{+}$, where the oldest age class $25^{+}$allowed for the accumulation of fish (i.e., plus group).

Initialization Initial (1962) abundance at age was assumed equal to the equilibrium age structure and computed as follows. First, the equilibrium age structure was computed for ages $1-25$ based on natural and fishing mortality ( $F$ ), where $F$ was set equal to a value that resulted in the 1962 biomass level equaling $90 \%$ of the unfished level. This was based on the assumption by the SEDAR-25 workshop panel that the stock was lightly exploited prior to the 1960's.

Natural Mortality Rate The natural mortality rate $(M)$ was assumed constant over time, but decreasing with age. The form of $M$ as a function of age was based on Lorenzen (1996). The Lorenzen estimates of $M_{a}$ were rescaled to provide the
same fraction of fish surviving from age- 1 through the oldest observed age ( 40 yr ) as would occur with constant $M=0.10$. This approach using cumulative mortality is consistent with the findings of Hoenig (1983) and Hewitt and Hoenig (2005).

Growth Mean size at age of the population (total length, TL) was modeled with the von Bertalanffy equation, and weight at age (whole weight, WW) was modeled as a function of total length (Figure 8.1, Table 7.1). Parameters of growth and conversions (TL-WW) were estimated during SEDAR-25 and were treated as input to the assessment model. The von Bertalanffy parameter estimates from the DW were $L_{\infty}=825.1, k=0.189$, and $t_{0}=-0.47$. For fitting length composition data, the distribution of size at age was assumed normal with coefficient of variation (CV) estimated by the assessment model. A constant CV, rather than constant standard deviation, was suggested by the size at age data.

Female maturity Females were modeled to be fully mature at age 4 and the proportion mature at ages 1, 2, and 3 were estimated to be 0.1, 0.25, and 0.5 respectively (Table 7.1).

Spawning Stock Spawning stock was modeled using mature female gonad weight measured at the time of peak spawning. For tilefish, peak spawning was considered to occur in May. In cases when reliable estimates of fecundity are unavailable, spawning biomass, and in this case, female gonad weight, is commonly used as a proxy for population fecundity.

Recruitment Expected recruitment of age-1 fish was predicted from spawning stock using the Beverton-Holt spawnerrecruit model. Annual variation in recruitment was assumed to occur with lognormal deviations for years 1976-2007 only. The start of recruitment residuals in 1976 was based on examination of a series of different starting years and the start of the age and length composition data that have information on year class strength. The ending year of estimated recruitment residuals (2007) is based on the age at full selection in the fisheries and the last year of age composition data. Because the age at full selection for the tilefish fisheries generally occurs at age 7 and the last year of composition data in the model is 2014, recruitment deviations during 2007-2014 could not be reliably estimated. Autocorrelation of the recruitment deviations was assumed to be zero.

Landings The model included time series of landings from three fleets: commercial longlines (1962-2014), commercial handlines (1962-2014), and general recreational (1981-2014). Landings were modeled with the Baranov catch equation (Baranov 1918) and were fitted in units of weight ( 1000 lb whole weight). A correction was made to the BAM code such that predicted recreational landings in whole pounds (vs. numbers in previous assessment) are now fit to observed landings in whole pounds.

Fishing Mortality For each time series of landings, the assessment model estimated a separate full fishing mortality rate $(F)$. Age-specific rates were then computed as the product of full $F$ and selectivity at age. Apical $F$ was computed as the maximum of $F$ at age summed across fleets.

Selectivities Selectivity curves applied to landings and CPUE series were estimated using a parametric approach. This approach applies plausible structure on the shape of the curves, and achieves greater parsimony than occurs with unique parameters for each age. As in SEDAR-25, selectivity of landings from all fleets were modeled as flat-topped, using a two parameter logistic function. Selectivity of the fishery-dependent longline logbook index was the same as that of the longline fleet. The MARMAP index was also modeled as a flat-topped, two parameter logistic function.

In SEDAR-25, a selectivity curve was not estimated for the recreational fleet due to concerns about low sample sizes and noisy composition data. Instead, the recreational selectivity was assumed to be equal to the commercial handline fishery because both sectors use vertical hook and line and their length compositions appeared to cover a similar range. However, once commercial and research SRHS samples were removed from the recreational data considered in this update assessment, it was determined that the recreational fleet selected for smaller fish than the handline fishery. Therefore, a separate selectivity curve was estimated for the recreational fishery in this update.

Indices of Abundance The model was fit to two indices of relative abundance: MARMAP longline (binned years between 1985 and 2011) and commercial longline (1993-2014). Predicted indices were conditional on selectivity of the corresponding fleet or survey and were computed from abundance or biomass (as appropriate) at the midpoint of the year.

In this assessment, commercial CPUE units within the model code were converted from gutted to whole weight to better match the population units of whole weight. This correction affected only the relative value of the index's catchability coefficient between assessments, but does not affect model results in any other way because catchability is a scaling parameter.

Catchability In the BAM, catchability scales indices of relative abundance to estimated exploitable abundance at large. Following the methodology used to generate the SEDAR-25 base run, this update assessment assumed time-invariant catchability.

Fitting Criterion The fitting criterion was a penalized likelihood approach in which observed landings were fit closely, and observed composition data and abundance indices were fit to the degree that they were compatible. Landings and index data were fitted using lognormal likelihoods. Length and age composition data were fitted using multinomial likelihoods. In this update, the multinomial likelihood was upgraded from the function described in SEDAR-25 to the robust multinomial likelihood (Francis 2011) to ensure best practices were used and to emulate other recent SEDAR assessments including the 2014 gag update assessment (SEDAR 2014). For multinomial likelihoods, the number of trips sampled was used as the measure of effective sample size.

The model includes the capability for each component of the likelihood to be weighted by user-supplied values (for instance, to give more influence to stronger data sources). For data components, these weights were applied by either adjusting CVs (lognormal components) or adjusting effective sample sizes (multinomial components). In this application to tilefish, CVs of landings (in arithmetic space) were assumed equal to 0.05 , to achieve a close fit to these time series yet allowing some imprecision. Weights on other data components (indices, age and length compositions) were adjusted iteratively, starting from initial weights as follows. The CVs of indices were set equal to the estimated values. Effective sample sizes of the multinomial components were assumed equal to the number of trips sampled annually, rather than the number of fish measured, reflecting the belief that the basic sampling unit occurs at the level of trip. These initial weights were then adjusted until standard deviations of normalized residuals were near 1.0 (SEDAR25-RW04, SEDAR25-RW06). The weight on the commercial longline index was then adjusted upward to a value of 3 (SEDAR25-RW06), in accordance with the principle that abundance data should be given primacy (Francis 2011), and consistent with SEDAR-25 procedures.

In addition, the compound objective function included several prior distributions, applied to the CV of growth (based on the empirical estimate), slope parameters for each selectivity function, and recruitment standard deviation based on Beddington and Cooke (1983) and Mertz and Myers (1996). Priors were applied to maintain parameter estimates near reasonable values, and to prevent the optimization routine from drifting into parameter space with negligible gradient in the likelihood which can result in a non-positive definite Hessian matrix.

Parameters Estimated The model estimated annual fishing mortality rates of each fishery, selectivity parameters, catchability coefficients associated with indices, parameters of the spawner-recruit model, annual recruitment deviations, and CV of size at age. Estimated parameters are described mathematically in the document (Williams and Shertzer 2015).

### 4.1.4 Per Recruit and Equilibrium Analyses

Static spawning potential ratio (static SPR) of each year was computed as the asymptotic spawners per recruit given that year's fishery-specific $F$ s and selectivities, divided by spawners per recruit that would be obtained in an unexploited stock. In this form, static SPR ranges between zero and one, and it represents SPR that would be achieved under an equilibrium age structure given the year-specific $F$.

Yield per recruit and spawning potential ratio were computed as functions of $F$, as were equilibrium landings and spawning biomass. Equilibrium landings were also computed as functions of biomass $B$, which itself is a function of $F$. As in computation of MSY-related benchmarks (described in the next section), per recruit and equilibrium analyses applied the most recent selectivity patterns averaged across fisheries, weighted by each fleet's $F$ from the last three years (2012-2014).

### 4.1.5 Biological Reference Points

Biological reference points (benchmarks) were calculated based on maximum sustainable yield (MSY) estimates in gutted pounds from the Beverton-Holt spawner-recruit model with bias correction (expected values in arithmetic space). In this assessment of tilefish, the quantities $F_{\mathrm{MSY}}, \mathrm{SSB}_{\mathrm{MSY}}, B_{\mathrm{MSY}}$, and MSY were estimated by the method of Shepherd (1982). Here, spawning stock measures total gonad weight of mature females. These benchmarks are conditional on the estimated selectivity functions and the relative contributions of each fleet's fishing mortality. The selectivity pattern used here was the effort-weighted selectivities at age, with effort from each fishery estimated as the full $F$ averaged over the last three years of the assessment (2012-2014).

The maximum fishing mortality threshold (MFMT) is defined by the SAFMC as $F_{\mathrm{MSY}}$, and the minimum stock size threshold (MSST) as MSST $=0.75 * \mathrm{SSB}_{\mathrm{MSY}}$ (Restrepo et al. 1998). Overfishing is defined as $F>$ MFMT and overfished as $\mathrm{SSB}<\mathrm{MSST}$. Current status of the stock is represented by SSB in the latest assessment year (2014), and current status of the fishery is represented by the geometric mean of $F$ from the latest three years (2012-2014).

### 4.1.6 Sensitivity and Retrospective Analyses

In order to portray uncertainty in point estimates produced by the base run described above, sensitivity analyses and a Monte-Carlo/bootstrap procedure were conducted. Steepness was not estimated for tilefish, but rather the parameter was fixed at 0.84 , the mode of the prior from the meta-analysis described in Shertzer and Conn (2012).

Sensitivity of results to the three changes made in this update were explored through sensitivity analyses. These model runs, as well as retrospective analyses, vary from the base run as follows:

- S1: multinomial likelihood as parameterized in SEDAR-25
- S2: recreational selectivity set equal to commercial handline selectivity
- S3: biased commercial longline age composition data from 1996, 1998, and 1999 included
- S4: Retrospective run with data through 2013
- S5: Retrospective run with data through 2012
- S6: Retrospective run with data through 2011
- S7: Retrospective run with data through 2010
- S8: Retrospective run with data through 2009
- S9: Retrospective run with data through 2008
- S10: Retrospective run with data through 2007
- S11: Retrospective run with data through 2006.

Retrospective analyses should be interpreted with caution because several data sources appear only near the end of the full time series. Also, some data are not continuous across years which removes information in larger intervals than a single year. Commercial handline age composition data and MARMAP index age composition data are not continuous by year from 2006 to 2011. The final year of recruitment deviations in each retrospective run was set to the terminal year minus seven years to mirror the base run model configuration.

### 4.1.7 Uncertainty and Measures of Precision

Uncertainty was in part examined through sensitivity runs. For the base run, uncertainty in results and precision of estimates was computed more thoroughly through a mixed Monte Carlo and bootstrap (MCB) approach (Efron and Tibshirani 1993; Restrepo et al. 1992; Legault et al. 2001; Manly 1997; SEDAR 2004; 2009; 2010). This approach is among those recommended for use in SEDAR assessments (SEDAR Procedural Guidance 2010).

In this assessment, the BAM was successively re-fit to $n=5000$ trials that differed from the original inputs by bootstrapping on data sources, and by Monte Carlo sampling of several key input parameters. Runs with $F>6$ were trimmed from the final uncertainty characterization. Runs with a maximum gradient $>1000$ and a recruitment standard deviation $>1.0$ were considered unrealistic and trimmed as well.

The MCB analysis should be interpreted as providing an approximation to the uncertainty associated with each output. The results are approximate for two related reasons. First, not all combinations of Monte Carlo parameter inputs are equally likely, as biological parameters might be correlated. Second, all runs are given equal weight in the results, yet some might provide better fits to data than others.

Bootstrap of Observed Data To include uncertainty in time series of observed landings and indices of abundance, multiplicative lognormal errors were applied through a parametric bootstrap. To implement this approach in the MCB trials, random variables $\left(x_{s, y}\right)$ were drawn for each year $y$ of time series $s$ from a normal distribution with mean 0 and variance $\sigma_{s, y}^{2}$ [that is, $x_{s, y} \sim N\left(0, \sigma_{s, y}^{2}\right)$ ]. Annual observations were then perturbed from their original values $\left(\hat{O}_{s, y}\right)$,

$$
\begin{equation*}
O_{s, y}=\hat{O}_{s, y}\left[\exp \left(x_{s, y}\right)-\sigma_{s, y}^{2} / 2\right] \tag{1}
\end{equation*}
$$

The term $\sigma_{s, y}^{2} / 2$ is a bias correction that centers the multiplicative error on the value of 1.0. Standard deviations in log space were computed from CVs in arithmetic space, $\sigma_{s, y}=\sqrt{\log \left(1.0+C V_{s, y}^{2}\right)}$. As used for fitting the base run, CVs of landings and discards were assumed to be 0.05 , and CVs of indices of abundance were those listed in Table 7.4.

Uncertainty in age and length compositions were included by drawing new distributions for each year of each data source, following a multinomial sampling process. Ages (or lengths) of individual fish were drawn at random with replacement using the cell probabilities of the original data. For each year of each data source, the number of individuals sampled was the same as in the original data (number of fish), and the effective sample sizes used for fitting (number of trips) was unmodified.

Monte Carlo Sampling In each successive fit of the model, several parameters were fixed (i.e., not estimated) at values drawn at random from distributions described below as in SEDAR-25.

Steepness: The steepness stock-recruit parameter was fixed at 0.84 in the base run based on a meta-analysis (Shertzer and Conn 2012). Uncertainty in this parameter was characterized by drawing random values from a truncated beta distribution [ $0.32,0.99$ ] with parameters $\alpha=5.94$ and $\beta=1.97$ estimated by Shertzer and Conn (2012).

Natural Mortality Point estimates of natural mortality $(M=0.10)$ were provided by the SEDAR-25 data workshop, but with some uncertainty. To carry forward this source of uncertainty, Monte Carlo sampling was used to generate deviations from the point estimate. A new $M$ value was drawn for each $M C B$ trial from a truncated normal distribution (range [0.03, $0.21])$ with mean equal to the point estimate $(M=0.10)$ and standard deviation set to provide a lower $95 \%$ confidence limit at 0.03 (the low end of the DW range). Each realized value of $M$ was used to scale the age-specific Lorenzen $M$, as in the base run.

Weighting of Indices In the base run, external weights applied to the commercial longline index was adjusted upward to a value of $\omega=3.0$. In MCB trials, that weight was drawn from a uniform distribution with bounds at $\pm 25 \%$ of 3.0.

### 4.1.8 Projection Methods

Projections were run to predict stock status in years after the assessment, 2015-2024. The structure of the projection model was the same as that of the assessment model, and parameter estimates were those from the assessment. Fully selected $F$ was apportioned between landings according to the selectivity curves averaged across fisheries, using geometric mean $F$ from the last three years of the assessment period (2012-2014).

Central tendencies of SSB (time of peak spawning), $F$, recruits, abundance, biomass, and landings were represented by deterministic projections using parameter estimates from the base run. These projections were built on the estimated spawner-recruit relationship with bias correction, and were thus consistent with estimated benchmarks in the sense that long-term fishing at $F_{\text {MSY }}$ would yield MSY from a stock size at $\mathrm{SSB}_{\text {MSY }}$. Uncertainty in future time series was quantified through projections that extended the Monte Carlo/Bootstrap (MCB) fits of the stock assessment model.

Initialization of projections Point estimates of initial abundance at age in the projection (start of 2015), other than at age 1, were taken to be the 2014 estimates from the assessment, discounted by 2014 natural and fishing mortalities. The initial abundance at age 1 was computed using the estimated spawner-recruit model and a 2014 estimate of SSB.

Fishing rates or catch levels that define the projections were assumed to start in 2017. Because the assessment period ended in 2014, the projections required an initialization period. Fishing mortality in 2015 was assumed equal to the geometric mean $F$ from the last three years of the assessment period. As requested in the TORs, three constant- $F$ projection scenarios were considered:

- Scenario 1: $F=$ Fcurrent
- Scenario 2: $F=F_{\mathrm{MSY}}$
- Scenario 3: $F=75 \% F_{\mathrm{MSY}}$.

Uncertainty of projections To characterize uncertainty in future stock dynamics, stochasticity was included in replicate projections, each an extension of a single MCB assessment model fit. Thus, projections carried forward uncertainties in natural mortality, as well as in estimated quantities such as spawner-recruit parameters, selectivity curves, and in initial (start of 2015) abundance at age. Initial and subsequent recruitment values were generated with stochasticity using a Monte Carlo procedure, in which the estimated Beverton-Holt model of each MCB fit was used to compute median unbiased annual recruitment values $\left(\bar{R}_{y}\right)$. Variability was added to the median values by choosing multiplicative deviations at random from the recruitment deviations estimated for that chosen MCB run. Because the base run model assumed no recruitment deviation for years 2007-2014, the initial projection year (start of 2015) ages 2-7 included additional variability in recruitment following the same method for subsequent years at age-1.

The procedure generated 20,000 replicate projections of MCB model fits drawn at random (with replacement) from the MCB runs. In cases where the same MCB run was drawn, projections would still differ as a result of stochasticity in projected recruitment streams. Precision of projections was represented graphically by the $5^{t h}$ and $95^{t h}$ percentiles of the replicate projections.

Rebuilding time frame Based on results from previous SEDAR assessments, tilefish was not overfished and no rebuilding plan is necessary.

### 4.1.9 Acceptable Biological Catch

Acceptable biological catch (ABC) was computed using the probability-based approach of Shertzer et al. (2008). In short, this approach solves for annual levels of projected landings that are consistent with a preset, acceptable probability of overfishing $\left(P^{\star}\right)$ in each year. The method considers uncertainty in $F_{\text {MSY }}$, computed as in Section 4.1.7, and described by the probability density function, $\phi_{F_{\mathrm{MSY}}}$. It also considers uncertainty in annual fishing mortality, described by the probability density function, $\phi_{F_{t}}$. Here, $\phi_{F_{t}}$ is computed by scaling $\phi_{F_{\mathrm{MSY}}}$ toward smaller values of F using a scalar multiplier $\lambda \leq 1.0$. Given the distributions $\phi_{F_{\mathrm{MSY}}}$ and $\phi_{F_{t}}$, the probability of overfishing associated with catch $C$ can be computed as,

$$
\begin{equation*}
\operatorname{Pr}\left(F_{t}>F_{\mathrm{MSY}}\right)=\int_{0}^{\infty}\left[\int_{F}^{\infty} \phi_{F_{t}}(\theta) d \theta\right] \phi_{F_{\mathrm{MSY}}}(F) d F \tag{2}
\end{equation*}
$$

where $\theta$ is a dummy integration variable. This equation was solved for $\lambda$ and the value of $\lambda$ was used in the projections such that in each iteration, the applied fishing rate was $F=\lambda F_{\mathrm{MSY}}$, where $F_{\mathrm{MSY}}$ was specific to that particular run (i.e., MCB draw). The ABC at the desired $P^{\star}$ was defined as the median catch from the projections.

In this application, projections were run for 10 years past the end of the assessment. The first year of the projection was 2015 and management was expected to begin in 2017. Given recent landings were greater than the ACL, interim landings were assumed to be the average of the last three years. No implementation uncertainty was included. The following projections (continued from list above) were conducted as requested in the Terms of Reference:

- Scenario 4: $P^{\star}=0.35$
- Scenario 5: $P^{\star}=0.50$.


### 4.2 Model Results

### 4.2.1 Base Run Results

Measures of Overall Model Fit Generally, the Beaufort Assessment Model (BAM) fit the available data well. The model was configured to fit observed commercial and recreational landings and the commerical longline index closely (Figures 8.2-8.5). Since the mid-2000s, the general trend in the commercial longline index had been increasing; however, CPUE declined in the last two years of this update assessment. As in SEDAR-25, fits to the fishery-independent MARMAP index of abundance captured recent, but not early, trends in relative abundance (Figure 8.6).

Predicted length compositions from each fishery were reasonably close to observed data in most years, as were predicted age compositions (Figure 8.7). As in SEDAR-25, fit of length composition data for the commercial longline fishery was poor, but could not be remedied during this assessment update.

Parameter Estimates Estimates of all parameters from the catch-age model are shown in Appendix C. Estimates of management quantities and some key parameters are reported in sections below.

Stock Abundance and Recruitment Estimated abundance declined in the early 1980s, exhibited a smaller peak in the 1990s, then stabilized at moderate levels in the 2000s (Figure 8.8). Older ages appear to have been significantly truncated by the late 1980s (Table 7.9). Moderate expansion of population age structure began again in the mid-2000s; however fish ages $17+$ are still relatively rare in the population.

Annual estimated number of recruits is shown in Table 7.9 (age-1 column) and in Figure 8.9. The model had identified the 2001 year class (age-1 fish) as being exceptionally strong in SEDAR-25; however, with the inclusion of additional years of age compositions data to the updated model, the 2001 year class shows no evidence of having been particularly strong.

Total and Spawning Biomass Estimated biomass and biomass at age exhibited a largely similar pattern to that of abundance (Figure 8.10; Table 7.10). Total biomass declined in the early 1980s, remained relatively low until the early 2000s when biomass climbed again and stabilized at moderate levels. Total and spawning biomass showed similar trends (Figure 8.11; Table 7.11).

Selectivity Selectivity estimates among all fisheries and surveys were similar with the exception of the recreational fleet which appeared to target smaller fish than the commercial fisheries or the MARMAP longline survey (Figures $8.12-8.13$ ). Fish were estimated to be near fully selected by age 4 for the recreational fleet, age 8 for MARMAP, and by age 9 for the commercial fleets.

Average selectivities of landings were computed from $F$-weighted selectivities in the most recent period (Figure 8.14). These average selectivities were used to compute benchmarks and central-tendency projections. All selectivities from the most recent period, including average selectivities, are tabulated in Table 7.12.

Fishing Mortality Estimated fishing mortality rates $(F)$ began increasing in the early 1980s, peaked in the early 1990s, displayed another smaller peak around 2000, then declined steadily until 2012 when rates began to increase again (Figure 8.15, Table 7.13). The commercial longline fleet dominates total $F$ (Table 7.14). In any given year, the maximum $F$ at age (i.e., apical $F$ ) may be less than that year's sum of fully selected $F$ s across fleets. This inequality is due to full selection occuring at different ages among gears in the estimated selectivities.

Table 7.15 shows total predicted landings in weight and Table 7.16 shows total predicted landings in numbers by fleet. Estimated landings at age in weight and numbers are provided in Tables 7.17, 7.18. Commercial harvest has increased in both fleets since the last assessment. In general, the majority of estimated landings were from the commercial longline sector (Figures 8.17, 8.16; Tables 7.15). During the same time period, recreational harvest increased for two years and then declined.

Spawner-Recruitment Parameters The estimated Beverton-Holt spawner-recruit curve is shown in Figure 8.18, along with the effect of density dependence on recruitment, depicted graphically by recruits per spawner as a function of spawners. Values of recruitment-related parameters were as follows: assumed steepness $h=0.84$, unfished age- 1 recruitment $\widehat{R_{0}}=362$, 411, unfished spawning biomass per recruit $\phi_{0}=0.00028$, and assumed standard deviation of recruitment residuals in log space $\sigma=0.4$ (which resulted in bias correction $\varsigma=1.06$ ). The empirical standard deviation of recruitment residuals in log space was $\widehat{\sigma}=0.34$. Uncertainty in these quantities was estimated through the Monte Carlo/bootstrap (MCB) analysis (Figure 8.19).

### 4.2.2 Per Recruit and Equilibrium Analyses

Static spawning potential ratio (static SPR) shows a general trend of decline during the 1970s and early 1980s, followed by a relatively stable period, an increasing trend between 2000 and 2011, then a decline in the last three years of the assessment (Figure 8.20, Table 7.11). Values lower than the MSY level, such as those seen in recent years, imply that population equilibria would be lower than desirable (as defined by MSY) given estimated fishing rates.

Yield per recruit and spawning potential ratio were computed as functions of $F$ (Figure 8.21). As in computation of MSYrelated benchmarks, per recruit analyses applied the most recent selectivity patterns averaged across fisheries, weighted by $F$ from the last three years (2012-2014). The $F$ s that provide $30 \%, 40 \%$, and $50 \%$ SPR are $0.17,0.11$, and 0.08 , respectively (Table 7.19).

As in per recruit analyses, equilibrium landings and spawning biomass were computed as functions of $F$ (Figures 8.22). By definition, the $F$ that maximizes equilibrium landings is $F_{\mathrm{MSY}}$, and the corresponding landings and spawning biomass are MSY and $\mathrm{SSB}_{\text {MSY }}$. Equilibrium landings and discards could also be viewed as functions of biomass $B$, which itself is a function of $F$ (Figure 8.23).

### 4.2.3 Benchmarks/Reference Points

Biological reference points (benchmarks) were derived analytically assuming equilibrium dynamics, corresponding to the expected spawner-recruit curve (Figure 8.18). Reference points estimated were $F_{\mathrm{MSY}}, \mathrm{MSY}, F / F_{\mathrm{MSY}}, B_{\mathrm{MSY}}, \mathrm{SSB}_{\mathrm{MSY}}$, and SSB/MSST. Based on $F_{\mathrm{MSY}}$, three possible values of $F$ at optimum yield (OY) were considered- $F_{\mathrm{OY}}=65 \% F_{\mathrm{MSY}}$, $F_{\mathrm{OY}}=75 \% F_{\mathrm{MSY}}$, and $F_{\mathrm{OY}}=85 \% F_{\mathrm{MSY}}$ —and for each, the corresponding yield was computed. Standard errors of benchmarks were approximated as those from Monte Carlo/bootstrap analysis (Section 4.1.7).

Estimates of benchmarks are summarized in Table 7.19. Point estimates of MSY-related quantities were $F_{\mathrm{MSY}}=0.24 \mathrm{y}^{-1}$, $\mathrm{MSY}=560 \mathrm{klb}, B_{\mathrm{MSY}}=2574 \mathrm{mt}, \mathrm{SSB}_{\mathrm{MSY}}=21.93 \mathrm{mt}$, and $\mathrm{SSB} / \mathrm{MSST}=1.13$. Distributions of these benchmarks are shown in Figure 8.24.

### 4.2.4 Status of the Stock and Fishery

Estimated time series of stock status (SSB/MSST) shows decline in the early 1980s, and then increase since the mid2000s, (Figure 8.25, Table 7.11). Base-run estimates of spawning biomass remained below MSST throughout the 1990s and most of the 2000s, then rose above MSST during 2009 to 2014. Current stock status in the base run was estimated to be $\mathrm{SSB}_{2014} / \mathrm{MSST}=1.13$ (Table 7.19). MCB analysis suggests that the stock status determination of being not overfished (i.e., SSB > MSST) has a high degree of uncertainty (Figures 8.26, 8.27). Over 47\% of MCB runs were below MSST in the terminal year.

The estimated time series of $F / F_{\mathrm{MSY}}$ suggests that overfishing has occurred throughout a large potion of the assessment period (Figure 8.25, Table 7.11). Spikes in the early 1980s through 2004 are due primarily to the longline fleet (Figure 8.15). Current fishery status in the terminal year, with current $F$ represented by the geometric mean from 2012-2014, is estimated in the base run to be $F_{2012-2014} / F_{\mathrm{MSY}}=1.22$ (Table 7.19). This estimate indicates that overfishing is occurring and appears robust across MCB trials (Figures $8.26,8.27$ ). Across all MCB runs, $66 \%$ were above MFMT.

### 4.2.5 Sensitivity and Retrospective Analyses

Sensitivity runs, described in Section 4.1.6, may be useful for evaluating implications of changes made to the base assessment model between this update and SEDAR-25. The effect of each change on stock status determinations is provided in Figure 8.28 and Table 7.20. The estimation of recreational selectivity (not set equal to commerical handline selectivity) and the removal of biased age composition data in the 1990s had little to no effect on stock status estimates. The adoption of a robust multinomial likelihood allowed the model to place less emphasis on outlier data points which resulted in higher $F$ and lower SSB relative to benchmarks.

Retrospective analyses revealed no apparent pattern in $F, B$, SSB, recruits, $F / F_{\mathrm{MSY}}$, or $\mathrm{SSB} / \mathrm{MSST}$ and seemed to indicate little retrospective error (Figures 8.29-8.34).

### 4.2.6 Comparison with Previous Assessment

To place this update assessment in context with SEDAR-25, the base run from this update assessment was compared with results from two other model configurations similar to that used in SEDAR-25. The first comparison run, called "SEDAR25 rev," was the run resulting from implementation of the SEDAR-25 code after an error in the likelihood evaluation function was corrected. The second comparison run, called "Continuity," was simply the "SEDAR25 rev" run implemented using data updated through 2014.

Stock and fishery status estimated by this assessment show trends similar to those from SEDAR-25 (Figure 8.35; Table 7.20). This update assessment estimates higher F and lower SSB in recent years than runs configured similarly to SEDAR25. The main cause of this difference is improvements in fitting procedures through the adoption of a robust multinomial likelihood function as described in Section 4.1.6. Given the advances in BAM and the additional years of data, estimates of stock and fishery status from this assessment are expected to be improvements over those from SEDAR- 25.

### 4.2.7 Projections

There are only slight differences in the $F_{\text {current }}, F_{\text {MSY }}$, and $75 \% F_{M S Y}$ projection scenarios (Tables $7.21-7.23$ and Figures $8.37-8.41$ ). Under the $F_{\text {current }}$ projection, $\mathrm{SSB}_{\mathrm{MSY}}$ could not be achieved nor could overfishing be prevented (Table 7.21 and Figure 8.37). Under the $F_{\text {MSY }}$ projection, SSB approached $\mathrm{SSB}_{\mathrm{MSY}}$ and $F$ remained steady near or at $F_{\text {MSY }}$ (Table 7.22 and Figure 8.39). Under the $75 \% F_{M S Y}$ projection, SSB exceeded $\mathrm{SSB}_{\mathrm{MSY}}$ and overfishing was prevented (Table 7.23 and Figure 8.41).

The $P *=0.35$ and $P *=0.5$ projections demonstrated similar trends to the $75 \% F_{M S Y}$ and $F_{\text {MSY }}$ projections, respectively (Figures $8.43-8.45$ ). Annual landings (in numbers and 1000 lb gutted weight) associated with each $P *$ are listed in Tables 7.24-7.25.

## 5 Discussion

### 5.1 Comments on Projections

Projections should be interpreted in light of model assumptions and key aspects of the data. Some major considerations are the following:

- In general, projections of fish stocks are highly uncertain, particularly in the long term (e.g., beyond 5-10 years).
- Although projections included many major sources of uncertainty, they did not include structural (model) uncertainty. That is, projection results are conditional on one set of functional forms used to describe population dynamics, selectivity, recruitment, etc.
- Fisheries were assumed to continue fishing at their estimated current proportions of total effort, using the estimated current selectivity patterns. New management regulations that alter those proportions or selectivities would likely affect projection results.
- The projections assumed that the estimated spawner-recruit relationship applies in the future and that past residuals represent future uncertainty in recruitment. If future recruitment is characterized by runs of large or small year classes, possibly due to environmental or ecological conditions, stock trajectories may be affected.
- Recruitment in the last 7 years is not informed by data and could be higher or lower than expected recruitment which would affect projection results.
- Projections were based on the calendar year, because they are extensions of the assessment model. A shift in the fishing year relative to calendar year may introduce some unquantified disconnect between projection results and management implementation. However, if quotas are reached each year prior to December 31, as might be expected, all fishing mortality within a fishing year would also occur within the same calendar year.


### 5.2 Recommendations for the Next Benchmark Assessment

The following recommendations are offered for consideration during the next benchmark assessment:

- Re-examine the quantity and quality of biological samples collected by "Other" commercial gears. If adequate, consider methods for inclusion.
- Monitor the quantity of commercial and recreational discards and consider methods for inclusion if deemed necessary.
- More closely examine historical length composition data used in the assessment and consider alternate methods for incorporating this information in the model.


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## 7 Tables

Table 7.1. Life-history characteristics at age of tilefish, including average body size in total length (TL) and weight (mid-year), gonad weight (GW), proportion females mature (F.mat), and natural mortality (M).

| Age | TL (mm) | TL (in) | CV length | Whole weight (kg) | Whole weight (lb) | GW (kg) | F.mat | M |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 256.5 | 10.1 | 0.15 | 0.16 | 0.36 | 0.00 | 0.10 | 0.2968 |
| 2 | 354.4 | 14.0 | 0.15 | 0.45 | 0.99 | 0.00 | 0.25 | 0.2174 |
| 3 | 435.5 | 17.1 | 0.15 | 0.86 | 1.89 | 0.01 | 0.50 | 0.1783 |
| 4 | 502.6 | 19.8 | 0.15 | 1.34 | 2.96 | 0.01 | 1.00 | 0.1554 |
| 5 | 558.1 | 22.0 | 0.15 | 1.87 | 4.13 | 0.02 | 1.00 | 0.1405 |
| 6 | 604.1 | 23.8 | 0.15 | 2.40 | 5.30 | 0.04 | 1.00 | 0.1302 |
| 7 | 642.2 | 25.3 | 0.15 | 2.91 | 6.42 | 0.05 | 1.00 | 0.1227 |
| 8 | 673.7 | 26.5 | 0.15 | 3.39 | 7.47 | 0.07 | 1.00 | 0.1172 |
| 9 | 699.7 | 27.5 | 0.15 | 3.82 | 8.42 | 0.09 | 1.00 | 0.1130 |
| 10 | 721.3 | 28.4 | 0.15 | 4.21 | 9.27 | 0.11 | 1.00 | 0.1097 |
| 11 | 739.2 | 29.1 | 0.15 | 4.54 | 10.02 | 0.13 | 1.00 | 0.1072 |
| 12 | 754.0 | 29.7 | 0.15 | 4.84 | 10.66 | 0.15 | 1.00 | 0.1052 |
| 13 | 766.2 | 30.2 | 0.15 | 5.09 | 11.22 | 0.17 | 1.00 | 0.1035 |
| 14 | 776.4 | 30.6 | 0.15 | 5.30 | 11.69 | 0.18 | 1.00 | 0.1022 |
| 15 | 784.8 | 30.9 | 0.15 | 5.49 | 12.10 | 0.19 | 1.00 | 0.1012 |
| 16 | 791.7 | 31.2 | 0.15 | 5.64 | 12.44 | 0.21 | 1.00 | 0.1003 |
| 17 | 797.5 | 31.4 | 0.15 | 5.77 | 12.72 | 0.22 | 1.00 | 0.0996 |
| 18 | 802.2 | 31.6 | 0.15 | 5.88 | 12.96 | 0.23 | 1.00 | 0.0991 |
| 19 | 806.2 | 31.7 | 0.15 | 5.97 | 13.17 | 0.24 | 1.00 | 0.0986 |
| 20 | 809.4 | 31.9 | 0.15 | 6.05 | 13.34 | 0.24 | 1.00 | 0.0982 |
| 21 | 812.1 | 32.0 | 0.15 | 6.11 | 13.48 | 0.25 | 1.00 | 0.0979 |
| 22 | 814.4 | 32.1 | 0.15 | 6.17 | 13.59 | 0.25 | 1.00 | 0.0976 |
| 23 | 816.2 | 32.1 | 0.15 | 6.21 | 13.69 | 0.26 | 1.00 | 0.0974 |
| 24 | 817.7 | 32.2 | 0.15 | 6.25 | 13.77 | 0.26 | 1.00 | 0.0973 |
| 25 | 819.0 | 32.2 | 0.15 | 6.28 | 13.84 | 0.27 | 1.00 | 0.0971 |

Table 7.2. Time series of estimated landings in weight for commercial handline, commercial longline, and recreational fleets. Confidential data indicated by an asterisk (*).

| Year | Thousand Pounds |  |  |
| :---: | :---: | :---: | :---: |
|  | Handline (Gutted) | Longline (Gutted) | Recreational (Whole) |
| 1962 | 0.468 | 2.93 |  |
| 1963 | 0.443 | 2.78 |  |
| 1964 | 0.138 | 0.86 |  |
| 1965 | 3.208 | 20.10 |  |
| 1966 | 0.602 | 3.77 |  |
| 1967 | 1.426 | 8.93 |  |
| 1968 | 0.873 | 5.47 |  |
| 1969 | 0.713 | 4.47 |  |
| 1970 | 1.413 | 8.85 |  |
| 1971 | 2.618 | 16.40 |  |
| 1972 | 1.561 | 9.78 |  |
| 1973 | 5.469 | 34.26 |  |
| 1974 | 12.425 | 77.84 |  |
| 1975 | 21.571 | 133.97 |  |
| 1976 | 21.928 | 129.79 |  |
| 1977 | 25.734 | 62.76 |  |
| 1978 | 91.554 | 92.14 |  |
| 1979 | 55.857 | 114.23 |  |
| 1980 | 148.605 | 177.80 |  |
| 1981 | 334.407 | 783.69 | 0.412 |
| 1982 | 596.732 | 2774.40 | 0.018 |
| 1983 | 263.259 | 1630.17 | 0.592 |
| 1984 | 202.687 | 1108.28 | 4.445 |
| 1985 | 142.764 | 985.93 | 58.259 |
| 1986 | 120.538 | 981.22 | 0.167 |
| 1987 | 23.827 | 233.46 | 0.228 |
| 1988 | 50.123 | 452.68 | 2.427 |
| 1989 | 92.569 | 743.82 | 0.014 |
| 1990 | 86.011 | 757.83 | 0.350 |
| 1991 | 82.231 | 821.40 | 0.407 |
| 1992 | 81.444 | 878.69 | 5.004 |
| 1993 | 170.882 | 864.03 | 0.020 |
| 1994 | 105.336 | 701.69 | 7.564 |
| 1995 | 82.706 | 591.46 | 0.020 |
| 1996 | * | * | 3.205 |
| 1997 | 33.830 | 328.21 | 20.354 |
| 1998 | 28.527 | 334.57 | 0.746 |
| 1999 | 37.705 | 473.60 | 5.781 |
| 2000 | 54.117 | 666.86 | 9.789 |
| 2001 | 38.459 | 389.57 | 11.488 |
| 2002 | * | * | 10.110 |
| 2003 | 18.408 | 222.24 | 29.438 |
| 2004 | 29.051 | 231.88 | 61.915 |
| 2005 | * | * | 97.346 |
| 2006 | 26.498 | 379.48 | 58.926 |
| 2007 | 49.660 | 260.57 | 16.665 |
| 2008 | * | * | 0.020 |
| 2009 | * | * | 18.206 |

Table 7.2. (Continued) Time series of estimated landings in weight for commercial handline, commercial longline, and recreational fleets. Confidential data indicated by an asterisk (*).

|  | Thousand Pounds |  |  |
| :---: | ---: | ---: | ---: |
| Year | Handline <br> (Gutted) | Longline <br> (Gutted) | Recreational <br> (Whole) |
| 2010 | $*$ | $*$ | 13.935 |
| 2011 | 22.897 | 350.68 | 25.470 |
| 2012 | $*$ | $*$ | 28.382 |
| 2013 | $*$ | $*$ | 16.039 |
| 2014 | 175.722 | 523.99 | 7.013 |

Table 7.3. Time series of estimated recreational landings and commerical landings in numbers (1000 fish) by gear since sampling began. Confidential data indicated by an asterisk (*)

| Year | Handline | Longline | Other | Recreational |
| ---: | ---: | ---: | ---: | ---: |
| 1981 | $\cdot$ | $\cdot$ | $\cdot$ | 0.094 |
| 1982 | $\cdot$ | $\cdot$ | . | 0.012 |
| 1983 | 152.237 | 23.862 | 3.144 | 0.134 |
| 1984 | 81.664 | 18.442 | 1.827 | 1.004 |
| 1985 | 76.315 | 11.404 | 2.133 | 13.153 |
| 1986 | 86.233 | 10.929 | 1.751 | 0.038 |
| 1987 | 18.799 | 1.936 | 0.369 | 0.044 |
| 1988 | 43.842 | 4.555 | 0.851 | 0.548 |
| 1989 | 62.729 | 8.373 | 1.790 | 0.010 |
| 1990 | 60.164 | 7.816 | 1.442 | 0.092 |
| 1991 | 79.847 | 6.501 | 4.322 | 0.092 |
| 1992 | 75.837 | 6.925 | 7.426 | 1.144 |
| 1993 | 83.636 | 8.612 | 3.382 |  |
| 1994 | 90.347 | 15.437 | 0.166 | 1.713 |
| 1995 | 68.081 | 9.319 | 0.034 |  |
| 1996 | $*$ | $*$ | $*$ | 0.724 |
| 1997 | 33.268 | 3.255 | 0.448 | 4.567 |
| 1998 | 41.905 | 2.343 | 0.096 | 0.168 |
| 1999 | 49.906 | 1.962 | 0.552 | 1.308 |
| 2000 | 73.078 | 6.138 | 0.460 | 2.210 |
| 2001 | 40.876 | 2.165 | 0.122 | 2.594 |
| 2002 | $*$ | $*$ | $*$ | 2.283 |
| 2003 | 31.404 | 1.702 | 0.001 | 8.050 |
| 2004 | 17.116 | 2.684 | 0.022 | 13.979 |
| 2005 | $*$ | $*$ | $*$ | 27.706 |
| 2006 | 30.249 | 3.029 | 0.022 | 13.304 |
| 2007 | 23.365 | 4.592 | 0.001 | 3.763 |
| 2008 | $*$ | $*$ | $*$ |  |
| 2009 | $*$ | $*$ | $*$ | 4.110 |
| 2010 | $*$ | $*$ | $*$ | 3.146 |
| 2011 | 25.574 | 1.788 | 1.565 | 5.750 |
| 2012 | $*$ | $*$ | $*$ | 6.408 |
| 2013 | $*$ | $*$ | $*$ | 3.533 |
| 2014 | 30.977 | 12.754 | 0.025 | 1.132 |
|  |  |  |  |  |
|  |  | $*$ | $*$ |  |

Table 7.4. Observed indices of abundance and coefficient of variation (CV) from commercial longline logbooks and the MARMAP horizontal longline survey. MARMAP values are combined across years: 1983-1986 for 1985, 1996-1999 for 1998, 2000-2003 for 2002, 2004-2007 for 2006, and 2009-2011 for 2010.

| Year | C.Logbook | C.Logbook CV | MARMAP | MARMAP CV |
| :---: | :---: | :---: | :---: | :---: |
| 1985 | . | . | 1.03 | 1.63 |
| 1986 | . | . | . |  |
| 1987 |  | . | . |  |
| 1988 | . | . | . |  |
| 1989 |  | . | . |  |
| 1990 |  | . | . |  |
| 1991 | . | . | . |  |
| 1992 | . | . | . |  |
| 1993 | 0.62 | 0.21 | . |  |
| 1994 | 0.60 | 0.20 | . |  |
| 1995 | 0.74 | 0.22 | . |  |
| 1996 | 0.45 | 0.26 | . |  |
| 1997 | 0.62 | 0.22 | . | . |
| 1998 | 0.69 | 0.27 | 2.43 | 1.60 |
| 1999 | 0.70 | 0.28 | . |  |
| 2000 | 0.66 | 0.28 | . |  |
| 2001 | 0.62 | 0.24 | . | . |
| 2002 | 0.52 | 0.27 | 0.95 | 1.81 |
| 2003 | 0.68 | 0.24 | . | . |
| 2004 | 0.63 | 0.25 | . |  |
| 2005 | 0.96 | 0.33 | . | . |
| 2006 | 1.31 | 0.22 | 0.89 | 1.92 |
| 2007 | 1.94 | 0.23 | . |  |
| 2008 | 2.03 | 0.28 | . | . |
| 2009 | 1.84 | 0.30 | . | . |
| 2010 | 2.39 | 0.19 | 2.60 | 1.49 |
| 2011 | 2.82 | 0.21 | . | . |
| 2012 | 2.96 | 0.18 | . | . |
| 2013 | 2.27 | 0.19 | . |  |
| 2014 | 2.42 | 0.19 | . | . |

Table 7.5. Sample sizes (number of trips or sets) for length (len) and age (age) compositions by fleet and survey. Data sources are commercial handlines $(c H)$, commercial longlines ( $c L$ ), general recreational ( $r A$ ), and the MARMAP horizontal longline survey ( mm ).

| Year | len.cH | age.cH | len.cL | age.cL | len.rA | age.mm |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 2 | . | . | . | . |  |
| 1985 | 7 | . | . | . | . | 155 |
| 1986 | 3 | . | . |  | . | . |
| 1987 | 2 | . |  | 7 | . |  |
| 1988 | . | . | 8 | . | . |  |
| 1989 | . | . | 7 | . | 10 |  |
| 1990 |  | . | 7 | . | 7 |  |
| 1991 | 8 | . | 51 |  | . | . |
| 1992 | . | . | . | 100 | . |  |
| 1993 | 3 | . | . | 141 | . | . |
| 1994 | 3 | . | 77 |  | . |  |
| 1995 | 6 | . | . | 64 | . | . |
| 1996 | 3 | . | . |  | . | . |
| 1997 | . | 5 | . | 19 | 3 | . |
| 1998 | . | 2 | . | . | . | 74 |
| 1999 | . | 8 | . | . | . | . |
| 2000 | . | 8 | . | 13 | . | . |
| 2001 | . | 7 | . | 23 | 7 | . |
| 2002 | . | 13 | . | 19 | 6 | 59 |
| 2003 | . | . | . | 10 | 7 | . |
| 2004 | . | . | . | 15 | 4 | . |
| 2005 | . | 5 | . | 16 | 9 | . |
| 2006 | . | 2 | . | 36 | 4 | 52 |
| 2007 | . | . | . | 35 | . | . |
| 2008 | . | . | . | 20 | . | . |
| 2009 | . | . | . | 25 | 5 | . |
| 2010 | . | 2 | . | 24 | 5 | 93 |
| 2011 | . | 3 | . | 22 | . | . |
| 2012 | . | 21 | . | 48 | . | . |
| 2013 | . | 19 | . | 29 | 8 | . |
| 2014 | . | 38 | . | 18 | . | . |

Table 7.6. Annual proportion at length input to the tilefish model. The sample sizes are: n.fish-number of fish measured, n.trips-number of trips from which fish were measured.

| Year | n.fish | n.trips | 340 | 370 | Commercial Handline |  |  |  | 520 | 550 | 580 | 610 | 640 | 670 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | 400 | 430 | 460 | 490 |  |  |  |  |  |  |
| 1984 | 13 | 2 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0598 | 0.0000 | 0.0000 | 0.0000 | 0.1340 | 0.1794 | 0.4186 |
| 1985 | 32 | 7 | 0.0000 | 0.0000 | 0.0000 | 0.0070 | 0.0000 | 0.0070 | 0.0000 | 0.0018 | 0.0007 | 0.0470 | 0.1176 | 0.3115 |
| 1986 | 40 | 3 | 0.0000 | 0.0128 | 0.0256 | 0.0000 | 0.0897 | 0.0897 | 0.0641 | 0.0384 | 0.0897 | 0.0513 | 0.0769 | 0.1031 |
| 1987 | 17 | 2 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0345 | 0.0690 | 0.0690 | 0.0000 | 0.0000 | 0.0345 |
| 1991 | 55 | 8 | 0.0000 | 0.0000 | 0.0089 | 0.0268 | 0.0268 | 0.0537 | 0.0268 | 0.0805 | 0.0716 | 0.0805 | 0.0537 | 0.0910 |
| 1993 | 37 | 3 | 0.0000 | 0.0186 | 0.0186 | 0.0186 | 0.0186 | 0.0186 | 0.0929 | 0.0372 | 0.0372 | 0.1487 | 0.0929 | 0.0743 |
| 1994 | 30 | 3 | 0.0000 | 0.0000 | 0.0000 | 0.0121 | 0.0284 | 0.2195 | 0.0774 | 0.2038 | 0.1264 | 0.0732 | 0.1094 | 0.0526 |
| 1995 | 43 | 6 | 0.0000 | 0.0000 | 0.0000 | 0.0119 | 0.0460 | 0.0657 | 0.0642 | 0.1465 | 0.1216 | 0.1426 | 0.0753 | 0.0763 |
| 1996 | 13 | 3 | 0.0000 | 0.0215 | 0.0000 | 0.0108 | 0.0108 | 0.0108 | 0.0000 | 0.0000 | 0.1541 | 0.0108 | 0.1541 | 0.0000 |

Commercial Longline

| 1988 | 163 | 8 | 0.0019 | 0.0045 | 0.0108 | 0.0347 | 0.0596 | 0.0671 | 0.1192 | 0.1492 | 0.1074 | 0.0910 | 0.0625 |
| :--- | ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1989 | 62 | 7 | 0.0074 | 0.0037 | 0.0323 | 0.0414 | 0.0635 | 0.0561 | 0.1488 | 0.1207 | 0.1062 | 0.0799 | 0.0541 |
| 1990 | 129 | 7 | 0.0000 | 0.0021 | 0.0045 | 0.0046 | 0.0288 | 0.0499 | 0.0396 | 0.0879 | 0.1231 | 0.1191 | 0.0798 |
| 1991 | 225 | 51 | 0.0066 | 0.0110 | 0.0362 | 0.1062 | 0.1144 | 0.0850 | 0.0478 | 0.0727 | 0.0783 | 0.0945 | 0.0739 |
| 1994 | 183 | 77 | 0.0011 | 0.0013 | 0.0068 | 0.0601 | 0.1496 | 0.1764 | 0.1131 | 0.1263 | 0.0970 | 0.0628 | 0.0352 | Recreational


| 1989 | 17 | 10 | 0.3529 | 0.1765 | 0.1176 | 0.1176 | 0.0000 | 0.0588 | 0.0000 | 0.0000 | 0.1176 | 0.0000 | 0.0000 |
| :--- | ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1990 | 14 | 7 | 0.8571 | 0.0714 | 0.0000 | 0.0000 | 0.0000 | 0.0714 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 0.0000 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1997 | 14 | 3 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0714 | 0.1429 | 0.2143 | 0.1429 | 0.0714 | 0.0714 |
| 2001 | 20 | 7 | 0.1000 | 0.0000 | 0.0500 | 0.0500 | 0.0000 | 0.0000 | 0.4500 | 0.0500 | 0.1000 | 0.0500 | 0.0500 |
| 2002 | 28 | 6 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0357 | 0.1071 | 0.1071 | 0.0357 | 0.1786 | 0.1429 | 0.1071 |
| 2003 | 64 | 7 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0938 | 0.2656 | 0.2344 | 0.0781 | 0.1094 | 0.0938 | 0.0938 |
| 0.0312 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2004 | 28 | 4 | 0.0714 | 0.2857 | 0.2143 | 0.0357 | 0.0000 | 0.0000 | 0.1786 | 0.1786 | 0.0000 | 0.0000 | 0.0357 |
| 2005 | 130 | 9 | 0.0000 | 0.0000 | 0.0000 | 0.0077 | 0.0692 | 0.1538 | 0.3154 | 0.1692 | 0.1308 | 0.0692 | 0.0462 |
| 2006 | 17 | 4 | 0.0588 | 0.0588 | 0.2353 | 0.0000 | 0.1176 | 0.0588 | 0.1176 | 0.1765 | 0.0000 | 0.0000 | 0.0588 |
| 2009 | 12 | 5 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0833 | 0.0833 | 0.0833 | 0.1667 | 0.1667 | 0.1667 |
| 2010 | 11 | 5 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0909 | 0.0000 | 0.1818 | 0.0000 |
| 2013 | 27 | 8 | 0.0000 | 0.0000 | 0.0741 | 0.0000 | 0.0370 | 0.0370 | 0.1481 | 0.0741 | 0.0370 | 0.0370 | 0.1111 |

Table 7.6. (Continued) Annual proportion at length input to the tilefish model. The sample sizes are: n.fish-number of fish measured, n.tripsnumber of trips from which fish were measured.

| Commercial Handline |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | n.fish | n.trips | 700 | 770 | 760 | 790 | 820 | 850 | 880 | 910 | 940 | 970 | 1000 |
| 1984 | 13 | 2 | 0.0598 | 0.0598 | 0.0598 | 0.0000 | 0.0144 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0144 |
| 1985 | 32 | 7 | 0.1176 | 0.0710 | 0.0660 | 0.0235 | 0.0882 | 0.1176 | 0.0000 | 0.0235 | 0.0000 | 0.0000 | 0.0000 |
| 1986 | 40 | 3 | 0.1025 | 0.0513 | 0.0769 | 0.0256 | 0.0128 | 0.0384 | 0.0384 | 0.0128 | 0.0000 | 0.0000 | 0.0000 |
| 1987 | 17 | 2 | 0.1035 | 0.2414 | 0.3104 | 0.0690 | 0.0000 | 0.0000 | 0.0345 | 0.0345 | 0.0000 | 0.0000 | 0.0000 |
| 1991 | 55 | 8 | 0.1081 | 0.0753 | 0.0447 | 0.0455 | 0.0805 | 0.0805 | 0.0358 | 0.0000 | 0.0089 | 0.0000 | 0.0000 |
| 1993 | 37 | 3 | 0.0372 | 0.0000 | 0.0743 | 0.0186 | 0.0106 | 0.0186 | 0.0558 | 0.0186 | 0.0417 | 0.0558 | 0.0929 |
| 1994 | 30 | 3 | 0.0405 | 0.0163 | 0.0121 | 0.0121 | 0.0000 | 0.0163 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1995 | 43 | 6 | 0.0557 | 0.0542 | 0.0573 | 0.0411 | 0.0152 | 0.0191 | 0.0073 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1996 | 13 | 3 | 0.1649 | 0.0000 | 0.1541 | 0.0000 | 0.0000 | 0.1541 | 0.0000 | 0.0000 | 0.0000 | 0.1541 | 0.0000 |
| Commercial Longline |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1988 | 163 | 8 | 0.0509 | 0.0458 | 0.0362 | 0.0248 | 0.0119 | 0.0143 | 0.0162 | 0.0187 | 0.0075 | 0.0053 | 0.0013 |
| 1989 | 62 | 7 | 0.0366 | 0.0332 | 0.0410 | 0.0154 | 0.0095 | 0.0210 | 0.0191 | 0.0369 | 0.0185 | 0.0022 | 0.0030 |
| 1990 | 129 | 7 | 0.0937 | 0.0652 | 0.0602 | 0.0324 | 0.0362 | 0.0145 | 0.0193 | 0.0140 | 0.0242 | 0.0137 | 0.0068 |
| 1991 | 225 | 51 | 0.0506 | 0.0440 | 0.0348 | 0.0234 | 0.0206 | 0.0140 | 0.0110 | 0.0063 | 0.0031 | 0.0024 | 0.0044 |
| 1994 | 183 | 77 | 0.0394 | 0.0243 | 0.0164 | 0.0087 | 0.0109 | 0.0077 | 0.0062 | 0.0033 | 0.0031 | 0.0022 | 0.0006 |
| Recreational |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1989 | 17 | 10 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 1990 | 14 | 7 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 1997 | 14 | 3 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 2001 | 20 | 7 | 0.0000 | 0.0000 | 0.0000 | 0.0500 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.050 | 0.000 | 0.000 |
| 2002 | 28 | 6 | 0.0714 | 0.0357 | 0.0000 | 0.0000 | 0.0357 | 0.0357 | 0.0000 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 2003 | 64 | 7 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 2004 | 28 | 4 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 2005 | 130 | 9 | 0.0154 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 2006 | 17 | 4 | 0.0000 | 0.0588 | 0.0000 | 0.0000 | 0.0000 | 0.0588 | 0.0000 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 2009 | 12 | 5 | 0.0000 | 0.0833 | 0.0833 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0833 | 0.000 | 0.000 | 0.000 |
| 2010 | 11 | 5 | 0.0909 | 0.0000 | 0.0909 | 0.0909 | 0.1818 | 0.0909 | 0.0909 | 0.0000 | 0.000 | 0.000 | 0.000 |
| 2013 | 27 | 8 | 0.1111 | 0.0000 | 0.0000 | 0.0000 | 0.0370 | 0.0000 | 0.0000 | 0.1111 | 0.037 | 0.037 | 0.000 |

Table 7.7. Annual proportion at age from commercial samples input to the tilefish model. The sample sizes are: n.fish-number of fish measured, n.trips-number of trips from which fish were measured.

| Commercial Handline |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | n.fish | n.trips | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| 1997 | 84 | 5 | 0.0000 | 0.0000 | 0.0000 | 0.0083 | 0.0880 | 0.1334 | 0.1586 | 0.1231 | 0.2620 | 0.1549 | 0.0261 | 0.0118 | 0.0233 |
| 1998 | 43 | 2 | 0.0000 | 0.0000 | 0.0000 | 0.0175 | 0.0700 | 0.1581 | 0.0530 | 0.1751 | 0.2106 | 0.1406 | 0.1401 | 0.0175 | 0.0175 |
| 1999 | 35 | 8 | 0.0000 | 0.0000 | 0.0000 | 0.0167 | 0.0000 | 0.3357 | 0.2188 | 0.0772 | 0.1032 | 0.2185 | 0.0225 | 0.0000 | 0.0000 |
| 2000 | 222 | 8 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0407 | 0.1354 | 0.3657 | 0.2680 | 0.0541 | 0.0379 | 0.0460 | 0.0249 | 0.0139 |
| 2001 | 46 | 7 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0016 | 0.0013 | 0.1082 | 0.4005 | 0.2144 | 0.1132 | 0.0032 | 0.0000 | 0.1037 |
| 2002 | 160 | 13 | 0.0000 | 0.0000 | 0.0012 | 0.0298 | 0.1292 | 0.1598 | 0.0908 | 0.1033 | 0.1258 | 0.1111 | 0.0412 | 0.0476 | 0.0704 |
| 2005 | 103 | 5 | 0.0000 | 0.0000 | 0.0000 | 0.0037 | 0.0460 | 0.0934 | 0.1922 | 0.2220 | 0.0770 | 0.0503 | 0.0947 | 0.0658 | 0.0204 |
| 2006 | 59 | 2 | 0.0000 | 0.0000 | 0.0000 | 0.0058 | 0.1081 | 0.3178 | 0.2682 | 0.1807 | 0.0640 | 0.0291 | 0.0029 | 0.0058 | 0.0087 |
| 2010 | 13 | 2 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.1808 | 0.2417 | 0.0591 | 0.0627 | 0.0627 | 0.1495 | 0.1513 | 0.0313 |
| 2011 | 71 | 3 | 0.0000 | 0.0000 | 0.0141 | 0.0282 | 0.0563 | 0.2817 | 0.2817 | 0.1972 | 0.0704 | 0.0423 | 0.0141 | 0.0000 | 0.0141 |
| 2012 | 454 | 21 | 0.0000 | 0.0000 | 0.0220 | 0.1454 | 0.2599 | 0.0815 | 0.0727 | 0.1211 | 0.1035 | 0.0815 | 0.0441 | 0.0352 | 0.0176 |
| 2013 | 293 | 19 | 0.0000 | 0.0034 | 0.0273 | 0.0819 | 0.1297 | 0.0717 | 0.1024 | 0.0819 | 0.1229 | 0.1024 | 0.0648 | 0.0819 | 0.0375 |
| 2014 | 478 | 38 | 0.0000 | 0.0000 | 0.0021 | 0.0335 | 0.0941 | 0.0858 | 0.0649 | 0.0900 | 0.1360 | 0.0983 | 0.0795 | 0.1025 | 0.0586 | Commercial Longline


| 1987 | 28 | 7 | 0.0000 | 0.0000 | 0.1993 | 0.1313 | 0.3979 | 0.2715 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| :--- | ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1992 | 124 | 100 | 0.0000 | 0.0000 | 0.0000 | 0.0045 | 0.0754 | 0.2587 | 0.1848 | 0.1208 | 0.1536 | 0.0426 | 0.0445 | 0.0157 |
| 1993 | 209 | 141 | 0.0000 | 0.0000 | 0.0029 | 0.1446 | 0.2937 | 0.2695 | 0.1373 | 0.0795 | 0.0097 | 0.0078 | 0.0148 | 0.0085 |
| 1995 | 373 | 64 | 0.0000 | 0.0000 | 0.0000 | 0.0137 | 0.1159 | 0.2277 | 0.2885 | 0.2057 | 0.1044 | 0.0149 | 0.0078 | 0.0026 |
| 1997 | 782 | 19 | 0.0000 | 0.0000 | 0.0021 | 0.0266 | 0.0714 | 0.1371 | 0.2186 | 0.1904 | 0.1533 | 0.0883 | 0.0367 | 0.0287 |
| 0.0146 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2000 | 312 | 13 | 0.0000 | 0.0000 | 0.0026 | 0.0043 | 0.0864 | 0.1708 | 0.3229 | 0.1264 | 0.0730 | 0.0383 | 0.0432 | 0.0523 |
| 2001 | 234 | 23 | 0.0000 | 0.0000 | 0.0000 | 0.0180 | 0.0696 | 0.1177 | 0.2461 | 0.1965 | 0.0990 | 0.0664 | 0.0593 | 0.0360 |
| 2002 | 32 | 19 | 0.0000 | 0.0000 | 0.0000 | 0.0232 | 0.0596 | 0.0585 | 0.2390 | 0.1820 | 0.1160 | 0.0000 | 0.0216 | 0.0689 |
| 2003 | 167 | 10 | 0.0000 | 0.0000 | 0.0002 | 0.0006 | 0.0800 | 0.1507 | 0.3706 | 0.1719 | 0.1158 | 0.0577 | 0.0136 | 0.0110 |
| 2004 | 264 | 15 | 0.0000 | 0.0000 | 0.0065 | 0.0141 | 0.0387 | 0.0969 | 0.2925 | 0.1724 | 0.1415 | 0.0613 | 0.0452 | 0.0470 |
| 2005 | 368 | 16 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0036 | 0.0387 | 0.1377 | 0.2071 | 0.1496 | 0.1137 | 0.1110 | 0.0763 |
| 20.0183 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2006 | 820 | 36 | 0.0000 | 0.0000 | 0.0001 | 0.0124 | 0.0474 | 0.1204 | 0.1510 | 0.2279 | 0.1648 | 0.0904 | 0.0744 | 0.0255 |
| 2007 | 945 | 35 | 0.0000 | 0.0000 | 0.0000 | 0.0047 | 0.0386 | 0.1516 | 0.1229 | 0.1844 | 0.2009 | 0.0939 | 0.0667 | 0.0442 |
| 2008 | 554 | 20 | 0.0000 | 0.0000 | 0.0000 | 0.0013 | 0.0224 | 0.0810 | 0.1732 | 0.1862 | 0.1871 | 0.1641 | 0.0627 | 0.0342 |
| 2009 | 880 | 25 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0114 | 0.0679 | 0.1407 | 0.1580 | 0.1336 | 0.1423 | 0.1148 | 0.0753 |
| 0.0450 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2010 | 703 | 24 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0089 | 0.0564 | 0.1087 | 0.1745 | 0.1705 | 0.1407 | 0.1014 | 0.0744 |
| 2011 | 528 | 22 | 0.0000 | 0.0000 | 0.0000 | 0.0019 | 0.0246 | 0.0758 | 0.1761 | 0.2386 | 0.1477 | 0.1193 | 0.0606 | 0.0606 |
| 2012 | 1264 | 48 | 0.0000 | 0.0000 | 0.0063 | 0.0229 | 0.0372 | 0.0728 | 0.1187 | 0.1598 | 0.1709 | 0.1646 | 0.0910 | 0.0601 |
| 2013 | 694 | 29 | 0.0000 | 0.0000 | 0.0000 | 0.0202 | 0.0303 | 0.0346 | 0.0865 | 0.1081 | 0.1398 | 0.1657 | 0.1124 | 0.0821 |
| 2014 | 427 | 18 | 0.0000 | 0.0000 | 0.0000 | 0.0117 | 0.0398 | 0.0211 | 0.0632 | 0.0984 | 0.1101 | 0.1522 | 0.1382 | 0.0867 |
| 0.0562 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 7.7. (Continued) Annual proportion at age from commercial samples input to the tilefish model. The sample sizes are: n.fish-number of

|  |  | 18 | Commercial Handline |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | n.fish | n.trips | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| 1997 | 84 | 5 | 0.0039 | 0.0052 | 0.0013 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1998 | 43 | 2 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1999 | 35 | 8 | 0.0000 | 0.0000 | 0.0074 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2000 | 222 | 8 | 0.0069 | 0.0000 | 0.0060 | 0.0000 | 0.0000 | 0.0003 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2001 | 46 | 7 | 0.0539 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2002 | 160 | 13 | 0.0304 | 0.0003 | 0.0086 | 0.0202 | 0.0000 | 0.0000 | 0.0202 | 0.0000 | 0.0000 | 0.0099 | 0.0000 | 0.0000 |
| 2005 | 103 | 5 | 0.0638 | 0.0354 | 0.0068 | 0.0136 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0149 | 0.0000 | 0.0000 | 0.0000 |
| 2006 | 59 | 2 | 0.0058 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0029 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2010 | 13 | 2 | 0.0295 | 0.0000 | 0.0000 | 0.0313 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2011 | 71 | 3 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2012 | 454 | 21 | 0.0044 | 0.0022 | 0.0022 | 0.0022 | 0.0000 | 0.0022 | 0.0022 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2013 | 293 | 19 | 0.0171 | 0.0307 | 0.0137 | 0.0068 | 0.0102 | 0.0034 | 0.0000 | 0.0034 | 0.0000 | 0.0000 | 0.0000 | 0.0068 |
| 2014 | 478 | 38 | 0.0397 | 0.0272 | 0.0272 | 0.0084 | 0.0042 | 0.0084 | 0.0105 | 0.0084 | 0.0000 | 0.0021 | 0.0105 | 0.0084 |

Commercial Longline

| 1987 | 28 | 7 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| :--- | ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.0000 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 124 | 100 | 0.0135 | 0.0257 | 0.0194 | 0.0000 | 0.0000 | 0.0073 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1993 | 209 | 141 | 0.0075 | 0.0085 | 0.0082 | 0.0001 | 0.0000 | 0.0000 | 0.0000 | 0.0035 | 0.0000 | 0.0000 | 0.0000 |
| 1995 | 373 | 64 | 0.0024 | 0.0051 | 0.0006 | 0.0015 | 0.0003 | 0.0001 | 0.0002 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1997 | 782 | 19 | 0.0082 | 0.0016 | 0.0045 | 0.0006 | 0.0045 | 0.0031 | 0.0019 | 0.0029 | 0.0025 | 0.0006 | 0.0006 |
| 0.0011 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2000 | 312 | 13 | 0.0144 | 0.0042 | 0.0024 | 0.0000 | 0.0032 | 0.0044 | 0.0045 | 0.0028 | 0.0023 | 0.0006 | 0.0000 |
| 2001 | 234 | 23 | 0.0051 | 0.0041 | 0.0091 | 0.0098 | 0.0057 | 0.0015 | 0.0009 | 0.0005 | 0.0042 | 0.0053 | 0.0032 |
| 0.0121 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2002 | 32 | 19 | 0.0476 | 0.0000 | 0.0000 | 0.0241 | 0.0381 | 0.0000 | 0.0000 | 0.0260 | 0.0000 | 0.0000 | 0.0000 |
| 2003 | 167 | 10 | 0.0016 | 0.0132 | 0.0000 | 0.0005 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2004 | 264 | 15 | 0.0103 | 0.0000 | 0.0033 | 0.0063 | 0.0086 | 0.0011 | 0.0017 | 0.0000 | 0.0104 | 0.0197 | 0.0032 |
| 2005 | 368 | 16 | 0.0179 | 0.0344 | 0.0055 | 0.0281 | 0.0097 | 0.0111 | 0.0064 | 0.0000 | 0.0017 | 0.0026 | 0.0069 |
| 0.0060 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2006 | 820 | 36 | 0.0107 | 0.0071 | 0.0099 | 0.0091 | 0.0000 | 0.0020 | 0.0077 | 0.0030 | 0.0000 | 0.0011 | 0.0015 |
| 2007 | 945 | 35 | 0.0128 | 0.0141 | 0.0034 | 0.0046 | 0.0072 | 0.0036 | 0.0030 | 0.0004 | 0.0019 | 0.0016 | 0.0014 |
| 0.0056 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2008 | 554 | 20 | 0.0205 | 0.0132 | 0.0096 | 0.0029 | 0.0041 | 0.0022 | 0.0041 | 0.0025 | 0.0014 | 0.0006 | 0.0016 |
| 2009 | 880 | 25 | 0.0253 | 0.0200 | 0.0090 | 0.0100 | 0.0171 | 0.0094 | 0.0013 | 0.0040 | 0.0055 | 0.0044 | 0.0010 |
| 2010 | 703 | 24 | 0.0341 | 0.0226 | 0.0129 | 0.0163 | 0.0057 | 0.0025 | 0.0020 | 0.0036 | 0.0016 | 0.0039 | 0.0009 |
| 2011 | 528 | 22 | 0.0170 | 0.0114 | 0.0057 | 0.0019 | 0.0000 | 0.0000 | 0.0000 | 0.0038 | 0.0000 | 0.0000 | 0.0000 |
| 20.0053 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2012 | 1264 | 48 | 0.0111 | 0.0142 | 0.0040 | 0.0079 | 0.0040 | 0.0008 | 0.0032 | 0.0000 | 0.0008 | 0.0008 | 0.0000 |
| 2013 | 694 | 29 | 0.0490 | 0.0331 | 0.0245 | 0.0159 | 0.0058 | 0.0029 | 0.0058 | 0.0058 | 0.0014 | 0.0029 | 0.0014 |
| 2014 | 427 | 18 | 0.0515 | 0.0398 | 0.0398 | 0.0141 | 0.0234 | 0.0141 | 0.0187 | 0.0023 | 0.0070 | 0.0000 | 0.0000 |
| 0.0072 |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 7.8. Annual proportion at age from the MARMAP survey input to the tilefish model. The sample sizes are: n.fish-number of fish measured, n.sets-number of sets from which fish were measured. MARMAP values are combined across years: 1983-1986 for 1985, 1996-1999 for 1998, 2000-2003 for 2002, 2004-2007 for 2006, and 2009-2011 for 2010.

| MARMAP longline survey |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | n.fish | n.sets | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| 1985 | 269 | 155 | 0.0000 | 0.0223 | 0.0892 | 0.0818 | 0.1041 | 0.1190 | 0.1413 | 0.1375 | 0.1190 | 0.0446 | 0.0260 | 0.0186 | 0.0074 |
| 1998 | 344 | 74 | 0.0000 | 0.0029 | 0.0087 | 0.0494 | 0.1134 | 0.1715 | 0.2064 | 0.1831 | 0.1134 | 0.0552 | 0.0640 | 0.0203 | 0.0116 |
| 2002 | 94 | 59 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0213 | 0.0213 | 0.1277 | 0.2447 | 0.1596 | 0.1809 | 0.1170 | 0.0745 | 0.0426 |
| 2006 | 80 | 52 | 0.0000 | 0.0000 | 0.0000 | 0.0125 | 0.0625 | 0.1250 | 0.1750 | 0.1375 | 0.2250 | 0.1000 | 0.1000 | 0.0375 | 0.0125 |
| 2010 | 461 | 93 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0260 | 0.0889 | 0.1302 | 0.1952 | 0.1562 | 0.1605 | 0.0803 | 0.0607 | 0.0325 |

Table 7.8. (Continued) Annual proportion at age from the MARMAP survey input to the tilefish model. The sample sizes are: n.fish-number of fish measured, n.sets-number of sets from which fish were measured. MARMAP values are combined across years: 1983-1986 for 1985, 1996-1999 for 1998, 2000-2003 for 2002, 2004-2007 for 2006, and 2009-2011 for 2010.

| MARMAP longline survey |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | n.fish | n.trips | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| 1985 | 269 | 155 | 0.0112 | 0.0000 | 0.0223 | 0.0074 | 0.0074 | 0.0112 | 0.0000 | 0.0000 | 0.0000 | 0.0037 | 0.0074 | 0.0186 |
| 1998 | 344 | 74 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2002 | 94 | 59 | 0.0000 | 0.0000 | 0.0106 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2006 | 80 | 52 | 0.0000 | 0.0125 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2010 | 461 | 93 | 0.0239 | 0.0130 | 0.0174 | 0.0087 | 0.0022 | 0.0043 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |

Table 7.9. Estimated total abundance at age (1000 fish) at start of year.

| Table 7.9. Estimated total abundance at age (1000 fish) at start of year. |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| 1962 | 381.45 | 283.49 | 228.09 | 190.83 | 163.34 | 141.86 | 124.32 | 109.47 | 96.63 | 85.51 | 75.88 | 67.50 | 60.15 |
| 1963 | 381.45 | 283.49 | 228.09 | 190.83 | 163.37 | 141.93 | 124.54 | 109.95 | 97.34 | 86.28 | 76.60 | 68.15 | 60.74 |
| 1964 | 381.69 | 283.49 | 228.09 | 190.84 | 163.37 | 141.96 | 124.61 | 110.14 | 97.77 | 86.91 | 77.29 | 68.80 | 61.33 |
| 1965 | 381.84 | 283.67 | 228.09 | 190.84 | 163.38 | 141.96 | 124.63 | 110.21 | 97.95 | 87.31 | 77.87 | 69.43 | 61.92 |
| 1966 | 381.97 | 283.78 | 228.23 | 190.84 | 163.37 | 141.95 | 124.59 | 110.13 | 97.86 | 87.31 | 78.07 | 69.81 | 62.36 |
| 1967 | 382.09 | 283.88 | 228.33 | 190.95 | 163.37 | 141.96 | 124.62 | 110.18 | 97.92 | 87.37 | 78.21 | 70.11 | 62.81 |
| 1968 | 382.21 | 283.97 | 228.41 | 191.03 | 163.47 | 141.96 | 124.61 | 110.18 | 97.92 | 87.38 | 78.22 | 70.19 | 63.05 |
| 1969 | 382.32 | 284.06 | 228.48 | 191.10 | 163.54 | 142.05 | 124.62 | 110.19 | 97.95 | 87.41 | 78.25 | 70.23 | 63.15 |
| 1970 | 382.43 | 284.14 | 228.55 | 191.16 | 163.60 | 142.11 | 124.70 | 110.20 | 97.97 | 87.44 | 78.29 | 70.27 | 63.19 |
| 1971 | 382.52 | 284.22 | 228.61 | 191.22 | 163.65 | 142.15 | 124.74 | 110.25 | 97.95 | 87.43 | 78.29 | 70.27 | 63.19 |
| 1972 | 382.60 | 284.29 | 228.68 | 191.27 | 163.70 | 142.19 | 124.76 | 110.25 | 97.93 | 87.34 | 78.21 | 70.21 | 63.15 |
| 1973 | 382.67 | 284.35 | 228.74 | 191.32 | 163.74 | 142.24 | 124.81 | 110.30 | 97.98 | 87.38 | 78.19 | 70.19 | 63.14 |
| 1974 | 382.72 | 284.40 | 228.78 | 191.37 | 163.78 | 142.26 | 124.80 | 110.22 | 97.84 | 87.22 | 78.03 | 69.99 | 62.96 |
| 1975 | 382.71 | 284.44 | 228.82 | 191.41 | 163.81 | 142.26 | 124.71 | 109.97 | 97.42 | 86.73 | 77.54 | 69.54 | 62.50 |
| 1976 | 300.14 | 284.43 | 228.85 | 191.43 | 163.83 | 142.23 | 124.57 | 109.60 | 96.77 | 85.89 | 76.66 | 68.69 | 61.72 |
| 1977 | 310.16 | 223.06 | 228.84 | 191.46 | 163.85 | 142.25 | 124.56 | 109.49 | 96.45 | 85.33 | 75.93 | 67.92 | 60.98 |
| 1978 | 303.46 | 230.51 | 179.47 | 191.46 | 163.89 | 142.31 | 124.70 | 109.76 | 96.78 | 85.50 | 75.86 | 67.66 | 60.65 |
| 1979 | 260.03 | 225.53 | 185.46 | 150.15 | 163.86 | 142.27 | 124.54 | 109.41 | 96.34 | 85.09 | 75.36 | 67.01 | 59.89 |
| 1980 | 233.15 | 193.25 | 181.45 | 155.16 | 128.51 | 142.26 | 124.54 | 109.34 | 96.13 | 84.80 | 75.07 | 66.64 | 59.38 |
| 1981 | 238.26 | 173.27 | 155.48 | 151.79 | 132.76 | 111.47 | 124.17 | 108.55 | 94.95 | 83.43 | 73.72 | 65.40 | 58.17 |
| 1982 | 307.12 | 177.07 | 139.40 | 130.02 | 129.72 | 114.62 | 95.82 | 104.19 | 88.52 | 76.38 | 66.93 | 59.19 | 52.59 |
| 1983 | 262.68 | 228.22 | 142.40 | 116.41 | 110.56 | 109.96 | 92.80 | 69.18 | 66.18 | 52.60 | 44.45 | 38.79 | 34.31 |
| 1984 | 250.94 | 195.20 | 183.55 | 118.96 | 99.12 | 94.19 | 90.51 | 69.83 | 47.05 | 42.69 | 33.39 | 28.13 | 24.56 |
| 1985 | 193.41 | 186.48 | 157.00 | 153.33 | 101.27 | 84.54 | 77.93 | 69.08 | 48.65 | 31.26 | 27.97 | 21.82 | 18.40 |
| 1986 | 256.15 | 143.72 | 149.93 | 130.62 | 129.10 | 85.22 | 68.72 | 57.91 | 46.38 | 30.98 | 19.58 | 17.47 | 13.63 |
| 1987 | 339.96 | 190.34 | 115.58 | 125.20 | 111.04 | 109.36 | 68.85 | 49.35 | 36.42 | 27.20 | 17.77 | 11.18 | 9.98 |
| 1988 | 653.82 | 252.65 | 153.13 | 96.65 | 106.98 | 95.83 | 93.92 | 57.62 | 40.03 | 29.08 | 21.64 | 14.15 | 8.92 |
| 1989 | 425.09 | 485.88 | 203.22 | 127.96 | 82.39 | 91.68 | 80.53 | 74.50 | 42.77 | 28.71 | 20.65 | 15.35 | 10.05 |
| 1990 | 219.92 | 315.88 | 390.75 | 169.72 | 108.81 | 69.85 | 74.26 | 58.20 | 47.34 | 25.40 | 16.69 | 11.95 | 8.89 |
| 1991 | 222.31 | 163.42 | 254.00 | 326.16 | 144.04 | 91.64 | 55.39 | 50.90 | 33.85 | 25.25 | 13.18 | 8.60 | 6.16 |
| 1992 | 232.62 | 165.18 | 131.37 | 211.78 | 275.69 | 119.55 | 69.27 | 33.66 | 24.22 | 14.15 | 10.11 | 5.22 | 3.40 |
| 1993 | 250.94 | 172.82 | 132.71 | 109.26 | 177.43 | 222.41 | 82.65 | 33.73 | 11.08 | 6.47 | 3.52 | 2.47 | 1.27 |
| 1994 | 388.64 | 186.40 | 138.76 | 110.18 | 90.89 | 138.95 | 138.88 | 31.14 | 7.28 | 1.78 | 0.94 | 0.50 | 0.35 |
| 1995 | 241.68 | 288.71 | 149.71 | 115.24 | 91.87 | 72.41 | 92.57 | 61.87 | 8.88 | 1.64 | 0.37 | 0.19 | 0.10 |
| 1996 | 165.36 | 179.54 | 231.95 | 124.57 | 96.64 | 74.07 | 49.73 | 44.25 | 19.77 | 2.29 | 0.39 | 0.09 | 0.04 |
| 1997 | 199.31 | 122.87 | 144.34 | 193.37 | 105.29 | 80.54 | 56.96 | 31.65 | 22.77 | 9.09 | 1.01 | 0.17 | 0.04 |
| 1998 | 234.72 | 148.09 | 98.75 | 119.99 | 162.23 | 87.13 | 61.74 | 36.44 | 16.53 | 10.68 | 4.11 | 0.46 | 0.08 |
| 1999 | 394.38 | 174.41 | 119.07 | 82.40 | 101.71 | 136.12 | 68.27 | 41.07 | 20.16 | 8.29 | 5.19 | 1.98 | 0.22 |
| 2000 | 395.65 | 293.03 | 140.19 | 99.19 | 69.49 | 84.25 | 102.87 | 41.64 | 19.69 | 8.50 | 3.35 | 2.08 | 0.79 |
| 2001 | 358.35 | 293.94 | 235.42 | 116.48 | 82.92 | 56.10 | 58.91 | 51.84 | 14.55 | 5.66 | 2.29 | 0.89 | 0.55 |
| 2002 | 328.14 | 266.26 | 236.24 | 195.83 | 97.84 | 68.34 | 42.11 | 35.54 | 24.49 | 6.03 | 2.25 | 0.90 | 0.35 |
| 2003 | ${ }^{318.04}$ | 243.81 | 213.98 | 196.56 | 164.54 | 80.46 | 50.73 | 24.66 | 15.99 | 9.58 | 2.26 | 0.83 | 0.33 |
| 2004 | 343.65 | 236.33 | 195.99 | 177.86 | 164.98 | 137.34 | 63.92 | 35.66 | 15.09 | 9.10 | 5.33 | 1.25 | 0.46 |
| 2005 | 318.30 | 255.36 | 189.93 | 162.36 | 147.98 | 136.79 | 109.45 | 46.06 | 22.93 | 9.14 | 5.41 | 3.15 | 0.74 |
| 2006 | 245.14 | 236.52 | 205.18 | 156.83 | 133.94 | 121.78 | 108.87 | 79.85 | 30.45 | 14.40 | 5.65 | 3.33 | 1.95 |
| 2007 | 226.56 | 182.17 | 190.12 | 170.28 | 131.14 | 111.83 | 98.19 | 80.16 | 53.01 | 19.14 | 8.90 | 3.48 | 2.05 |
| 2008 | 294.87 | 168.36 | 146.51 | 158.64 | 144.57 | 111.92 | 93.66 | 77.80 | 59.60 | 38.15 | 13.65 | 6.34 | 2.48 |
| 2009 | 303.65 | 219.13 | 135.43 | 122.47 | 135.38 | 124.17 | 94.59 | 75.30 | 59.02 | 43.89 | 27.86 | 9.96 | 4.63 |
| 2010 | 310.25 | 225.66 | 176.25 | 113.03 | 104.08 | 115.88 | 104.99 | 76.71 | 58.15 | 44.45 | 32.83 | 20.83 | 7.46 |
| 2011 | 315.08 | 230.56 | 181.51 | 147.17 | 96.17 | 89.21 | 98.11 | 85.27 | 59.33 | 43.87 | 33.31 | 24.59 | 15.63 |
| 2012 | 318.74 | 234.15 | 185.44 | 151.41 | 124.90 | 82.23 | 75.44 | 79.80 | 66.24 | 45.02 | 33.08 | 25.11 | 18.57 |
| 2013 | 320.10 | 236.87 | 188.31 | 154.61 | 128.28 | 106.29 | 68.56 | 59.30 | 58.66 | 47.06 | 31.67 | 23.25 | 17.67 |
| 2014 | 319.03 | 237.88 | 190.50 | 157.15 | 131.31 | 109.35 | 88.49 | 53.43 | 42.85 | 40.77 | 32.33 | 21.73 | 15.96 |
| 2015 | 316.15 | 237.08 | 191.31 | 159.05 | 133.56 | 111.60 | 89.66 | 66.17 | 36.06 | 27.43 | 25.69 | 20.31 | 13.66 |

Table 7.9. (Continued) Estimated total abundance at age (1000 fish) at start of year.








N


$\stackrel{\rightharpoonup}{\mathrm{N}}$


















Table 7.10. Estimated biomass at age (1000 lb) at start of year

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1962 | 135.4 | 279.3 | 430.3 | 565.7 | 674.2 | 751.6 | 798.7 | 817.9 | 813.9 | 792.8 | 759.9 | 719.6 | 674.8 |
| 1963 | 135.4 | 279.3 | 430.3 | 565.7 | 674.2 | 752.0 | 800.1 | 821.7 | 819.9 | 799.8 | 767.2 | 726.6 | 681.2 |
| 1964 | 135.6 | 279.3 | 430.3 | 565.7 | 674.2 | 752.0 | 800.5 | 823.0 | 823.6 | 805.8 | 774.0 | 733.5 | 687.8 |
| 1965 | 135.6 | 279.3 | 430.3 | 565.7 | 674.2 | 752.2 | 800.7 | 823.4 | 825.2 | 809.5 | 780.0 | 740.1 | 694.7 |
| 1966 | 135.6 | 279.5 | 430.6 | 565.7 | 674.2 | 752.0 | 800.3 | 823.0 | 824.3 | 809.5 | 782.0 | 744.3 | 699.5 |
| 1967 | 135.6 | 279.5 | 430.8 | 566.1 | 674.2 | 752.0 | 800.5 | 823.2 | 824.7 | 810.0 | 783.3 | 747.4 | 704.6 |
| 1968 | 135.8 | 279.8 | 430.8 | 566.4 | 674.6 | 752.0 | 800.5 | 823.2 | 824.7 | 810.0 | 783.3 | 748.2 | 707.2 |
| 1969 | 135.8 | 279.8 | 431.0 | 566.6 | 674.8 | 752.7 | 800.5 | 823.4 | 825.0 | 810.4 | 783.7 | 748.7 | 708.3 |
| 1970 | 135.8 | 280.0 | 431.2 | 566.8 | 675.3 | 752.9 | 801.2 | 823.4 | 825.2 | 810.6 | 784.2 | 749.1 | 708.8 |
| 1971 | 135.8 | 280.0 | 431.2 | 567.0 | 675.3 | 753.1 | 801.4 | 823.9 | 825.0 | 810.6 | 784.0 | 749.1 | 708.8 |
| 1972 | 135.8 | 280.0 | 431.4 | 567.0 | 675.5 | 753.3 | 801.6 | 823.9 | 825.0 | 809.8 | 783.3 | 748.5 | 708.3 |
| 1973 | 135.8 | 280.0 | 431.4 | 567.2 | 675.7 | 753.5 | 801.8 | 824.3 | 825.4 | 810.2 | 783.1 | 748.2 | 708.1 |
| 1974 | 136.0 | 280.2 | 431.7 | 567.5 | 675.9 | 753.8 | 801.8 | 823.6 | 824.1 | 808.7 | 781.5 | 746.3 | 706.1 |
| 1975 | 135.8 | 280.2 | 431.7 | 567.5 | 676.2 | 753.8 | 801.2 | 821.7 | 820.6 | 804.0 | 776.7 | 741.4 | 701.1 |
| 1976 | 106.5 | 280.2 | 431.7 | 567.7 | 676.2 | 753.5 | 800.3 | 819.0 | 815.0 | 796.3 | 767.6 | 732.4 | 692.3 |
| 1977 | 110.2 | 219.8 | 431.7 | 567.7 | 676.2 | 753.5 | 800.3 | 818.1 | 812.4 | 791.0 | 760.4 | 724.2 | 684.1 |
| 1978 | 107.8 | 227.1 | 338.6 | 567.7 | 676.4 | 754.0 | 801.2 | 820.1 | 815.3 | 792.8 | 759.7 | 721.4 | 680.3 |
| 1979 | 92.4 | 222.2 | 349.9 | 445.1 | 676.4 | 753.8 | 800.1 | 817.5 | 811.5 | 788.8 | 754.6 | 714.5 | 671.7 |
| 1980 | 82.9 | 190.3 | 342.4 | 460.1 | 530.4 | 753.8 | 800.1 | 817.0 | 809.8 | 786.2 | 751.8 | 710.5 | 666.0 |
| 1981 | 84.7 | 170.6 | 293.2 | 450.0 | 547.8 | 590.6 | 797.6 | 811.1 | 799.8 | 773.6 | 738.3 | 697.3 | 652.6 |
| 1982 | 109.1 | 174.4 | 263.0 | 385.6 | 535.3 | 607.4 | 615.5 | 778.5 | 745.6 | 708.1 | 670.4 | 631.0 | 590.0 |
| 1983 | 93.3 | 224.9 | 268.7 | 345.2 | 456.4 | 582.7 | 596.1 | 517.0 | 557.5 | 487.7 | 445.1 | 413.6 | 384.9 |
| 1984 | 89.1 | 192.2 | 346.3 | 352.7 | 409.0 | 498.9 | 581.4 | 521.8 | 396.4 | 395.7 | 334.4 | 299.8 | 275.6 |
| 1985 | 68.8 | 183.6 | 296.1 | 454.6 | 418.0 | 447.8 | 500.7 | 516.1 | 409.8 | 289.9 | 280.2 | 232.6 | 206.4 |
| 1986 | 91.1 | 141.5 | 282.9 | 387.4 | 532.9 | 451.5 | 441.4 | 432.8 | 390.7 | 287.3 | 196.2 | 186.3 | 153.0 |
| 1987 | 120.8 | 187.4 | 218.0 | 371.3 | 458.3 | 579.4 | 442.2 | 368.8 | 306.9 | 252.2 | 177.9 | 119.3 | 112.0 |
| 1988 | 232.1 | 248.9 | 288.8 | 286.6 | 441.6 | 507.7 | 603.4 | 430.6 | 337.1 | 269.6 | 216.7 | 150.8 | 100.1 |
| 1989 | 151.0 | 478.6 | 383.4 | 379.4 | 340.0 | 485.7 | 517.4 | 556.7 | 360.2 | 266.1 | 206.8 | 163.6 | 112.7 |
| 1990 | 78.0 | 311.1 | 737.2 | 503.1 | 449.1 | 370.2 | 477.1 | 435.0 | 398.8 | 235.5 | 167.1 | 127.4 | 99.6 |
| 1991 | 78.9 | 160.9 | 479.1 | 966.9 | 594.4 | 485.5 | 355.8 | 380.3 | 285.3 | 234.1 | 132.1 | 91.7 | 69.0 |
| 1992 | 82.7 | 162.7 | 247.8 | 627.9 | 1137.8 | 633.4 | 445.1 | 251.5 | 203.9 | 131.2 | 101.2 | 55.6 | 38.1 |
| 1993 | 89.1 | 170.2 | 250.4 | 323.9 | 732.2 | 1178.4 | 530.9 | 252.0 | 93.3 | 60.0 | 35.3 | 26.5 | 14.3 |
| 1994 | 138.0 | 183.6 | 261.7 | 326.7 | 375.2 | 736.1 | 892.2 | 232.8 | 61.3 | 16.5 | 9.5 | 5.3 | 4.0 |
| 1995 | 85.8 | 284.4 | 282.4 | 341.7 | 379.2 | 383.6 | 594.8 | 462.3 | 74.7 | 15.2 | 3.7 | 2.0 | 1.1 |
| 1996 | 58.6 | 176.8 | 437.6 | 369.3 | 398.8 | 392.4 | 319.4 | 330.7 | 166.4 | 21.2 | 4.0 | 0.9 | 0.4 |
| 1997 | 70.8 | 121.0 | 272.3 | 573.4 | 434.5 | 426.6 | 366.0 | 236.6 | 191.8 | 84.2 | 10.1 | 1.8 | 0.4 |
| 1998 | 83.3 | 145.9 | 186.3 | 355.8 | 669.5 | 461.6 | 396.6 | 272.3 | 139.1 | 99.0 | 41.2 | 4.9 | 0.9 |
| 1999 | 140.0 | 171.7 | 224.7 | 244.3 | 419.8 | 721.1 | 438.5 | 306.9 | 169.8 | 76.9 | 52.0 | 21.2 | 2.4 |
| 2000 | 140.4 | 288.6 | 264.6 | 294.1 | 286.8 | 446.4 | 660.9 | 311.1 | 165.8 | 78.7 | 33.5 | 22.0 | 8.8 |
| 2001 | 127.2 | 289.5 | 444.0 | 345.5 | 342.2 | 297.2 | 378.5 | 387.4 | 122.6 | 52.5 | 22.9 | 9.5 | 6.2 |
| 2002 | 116.6 | 262.3 | 445.6 | 580.7 | 403.9 | 362.0 | 270.5 | 265.7 | 206.4 | 56.0 | 22.5 | 9.7 | 4.0 |
| 2003 | 112.9 | 240.1 | 403.7 | 582.7 | 679.0 | 426.2 | 325.8 | 184.3 | 134.7 | 88.8 | 22.5 | 8.8 | 3.7 |
| 2004 | 122.1 | 232.8 | 369.7 | 527.3 | 680.8 | 727.7 | 410.7 | 266.5 | 127.2 | 84.4 | 53.4 | 13.2 | 5.1 |
| 2005 | 113.1 | 251.5 | 358.3 | 481.3 | 610.7 | 724.7 | 703.1 | 344.1 | 193.1 | 84.7 | 54.2 | 33.7 | 8.4 |
| 2006 | 87.1 | 233.0 | 387.1 | 465.0 | 552.7 | 645.3 | 699.3 | 596.6 | 256.4 | 133.4 | 56.7 | 35.5 | 21.8 |
| 2007 | 80.5 | 179.5 | 358.7 | 504.9 | 541.2 | 592.4 | 630.7 | 599.0 | 446.4 | 177.5 | 89.1 | 37.0 | 23.1 |
| 2008 | 104.7 | 165.8 | 276.5 | 470.2 | 596.6 | 593.0 | 601.6 | 581.4 | 502.0 | 353.6 | 136.7 | 67.7 | 27.8 |
| 2009 | 107.8 | 215.8 | 255.5 | 363.1 | 558.7 | 657.9 | 607.6 | 562.6 | 497.1 | 407.0 | 279.1 | 106.3 | 52.0 |
| 2010 | 110.2 | 222.2 | 332.5 | 335.1 | 429.5 | 614.0 | 674.4 | 573.2 | 489.9 | 412.0 | 328.7 | 222.0 | 83.6 |
| 2011 | 112.0 | 227.1 | 342.4 | 436.3 | 396.8 | 472.7 | 630.3 | 637.1 | 499.8 | 406.8 | 333.6 | 262.1 | 175.3 |
| 2012 | 113.1 | 230.6 | 349.9 | 448.9 | 515.4 | 435.6 | 484.6 | 596.3 | 558.0 | 417.3 | 331.4 | 267.6 | 208.3 |
| 2013 | 113.8 | 233.2 | 355.2 | 458.3 | 529.3 | 563.1 | 440.5 | 443.1 | 494.1 | 436.3 | 317.2 | 247.8 | 198.2 |
| 2014 | 113.3 | 234.4 | 359.4 | 465.8 | 541.9 | 579.4 | 568.6 | 399.3 | 360.9 | 378.1 | 323.9 | 231.7 | 179.0 |
| 2015 | 112.2 | 233.5 | 360.9 | 471.6 | 551.2 | 591.3 | 576.1 | 494.5 | 303.8 | 254.4 | 257.3 | 216.5 | 153.2 |

Table 7.10. (Continued) Estimated biomass at age (1000 lb) at start of year

| Year | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1962 | 627.9 | 580.5 | 534.2 | 489.4 | 446.9 | 407.0 | 369.7 | 335.3 | 303.6 | 274.7 | 248.0 | 2204.8 | 15035.9 |
| 1963 | 634.0 | 586.2 | 539.3 | 494.1 | 451.3 | 410.9 | 373.2 | 338.6 | 306.7 | 277.3 | 250.4 | 2226.2 | 15141.6 |
| 1964 | 640.0 | 591.9 | 544.5 | 498.9 | 455.7 | 414.9 | 377.0 | 341.9 | 309.5 | 280.0 | 25.9 | 2247.8 | 15240.5 |
| 1965 | 646.4 | 597.7 | 550.1 | 504.0 | 460.1 | 419.1 | 380.7 | ${ }^{345.2}$ | ${ }^{312.6}$ | 282.9 | 255.5 | 2270.3 | 15335.3 |
| 1966 | 651.5 | 602.5 | 554.2 | 507.9 | 463.6 | 422.4 | 383.6 | 347.9 | 315.0 | 285.1 | 257.5 | 2288.2 | 15399.9 |
| 1967 | ${ }_{6}^{657.2}$ | ${ }^{608.0}$ | 559.5 | 512.8 | 468.3 | 426.4 | 387.4 | 351.4 | 318.1 | 287.7 | 259.9 | 2310.2 | 15479.5 |
| 1968 | 661.6 | 613.1 | 564.6 | 517.4 | 472.5 | ${ }^{430.3}$ | 390.9 | 354.5 | 321.0 | 290.3 | 262.3 | 2331.2 | 15546.8 |
| 1969 | 664.3 | 617.5 | 569.5 | 522.1 | 476.9 | 434.3 | 394.6 | 357.8 | 324.1 | 293.0 | 264.8 | 2353.2 | 15613.1 |
| 1970 | 665.4 | ${ }^{620.2}$ | 573.6 | 526.7 | 481.3 | 438.5 | 398.4 | 361.3 | 327.2 | 295.9 | 267.4 | 2375.7 | 15675.7 |
| 1971 | 665.6 | 620.8 | 575.6 | 530.2 | 485.5 | 442.2 | 401.9 | 364.6 | 330.3 | 298.5 | 269.8 | 2397.3 | ${ }^{15728.0}$ |
| 1972 | ${ }^{665.1}$ | ${ }^{620.6}$ | 575.8 | ${ }^{532.0}$ | 488.3 | 445.8 | 405.2 | 367.5 | ${ }^{332.9}$ | 301.2 | 272.1 | 2417.4 | 15766.8 |
| 1973 | ${ }_{6}^{665.1}$ | ${ }^{620.6}$ | 576.1 | 532.4 | 490.1 | 448.6 | 408.5 | 370.8 | 336.0 | 303.8 | 274.5 | 2439.2 | 15810.4 |
| 1974 | 663.1 | 618.8 | 574.5 | 531.1 | 489.4 | 449.1 | 410.3 | 373.0 | 338.0 | 305.8 | 276.2 | 2454.8 | 15821.0 |
| 1975 | 658.3 | 614.4 | 570.6 | 527.3 | 485.9 | 446.4 | 409.0 | 372.8 | 338.4 | 306.2 | 276.7 | 2459.5 | 15777.1 |
| 1976 | 649.7 | 606.3 | 563.1 | 520.5 | 479.7 | 440.7 | 404.1 | 369.3 | 336.2 | 304.7 | 275.6 | 2449.6 | 15637.8 |
| 1977 | 641.8 | 598.3 | 555.6 | 513.9 | 473.6 | ${ }_{435.2}$ | 398.8 | 365.1 | 333.1 | 302.9 | 274.3 | 2440.1 | 15481.9 |
| 1978 | 637.8 | 594.6 | 551.6 | 509.9 | 470.0 | ${ }^{432.1}$ | 396.2 | 362.4 | 331.1 | 301.8 | 274.3 | 2444.3 | 15367.7 |
| 1979 | ${ }^{628.8}$ | 585.5 | 543.2 | 501.8 | ${ }^{462.5}$ | 425.1 | 389.8 | 356.7 | 325.8 | 297.4 | 270.7 | 2426.6 | 15112.9 |
| 1980 | 621.5 | 577.8 | 535.7 | 494.7 | 455.7 | 418.7 | 384.0 | 351.4 | 321.2 | 293.0 | 267.2 | 2410.5 | 14832.5 |
| 1981 | 606.9 | 562.8 | 520.7 | 480.6 | 442.5 | 406.3 | 372.6 | 341.1 | 311.7 | 284.4 | 259.3 | 2357.2 | 14353.6 |
| 1982 | 547.8 | 506.4 | 467.2 | 430.3 | 395.9 | 363.5 | 333.1 | 304.9 | 278.7 | 254.4 | 231.9 | 2121.7 | 13049.4 |
| 1983 | 356.7 | 329.1 | 302.7 | 278.2 | 255.5 | 234.4 | 214.7 | 196.2 | 179.5 | 163.8 | 149.3 | 1374.8 | 9407.6 |
| 1984 | 254.2 | 234.4 | 215.0 | 196.9 | 180.3 | 165.1 | 151.2 | 138.2 | 126.3 | 115.3 | 105.2 | 972.9 | 7548.6 |
| 1985 | 188.1 | 172.4 | 158.1 | 144.4 | 131.8 | 120.4 | 110.0 | 100.5 | 91.7 | 83.8 | 76.3 | 710.5 | ${ }^{6393.0}$ |
| 1986 | 134.5 | 121.9 | 111.1 | 101.4 | 92.4 | 84.2 | 76.7 | 69.9 | 63.7 | 58.2 | 52.9 | 495.6 | 5437.3 |
| 1987 | 91.1 | 79.6 | 71.9 | 65.3 | 59.3 | 54.0 | 48.9 | 44.5 | ${ }^{40.6}$ | 37.0 | 33.7 | 315.7 | 4655.7 |
| 1988 | 93.0 | 75.4 | 65.5 | 58.9 | 53.4 | 48.3 | 43.9 | 39.7 | 36.2 | 32.8 | 29.8 | 280.9 | 4971.9 |
| 1989 | 74.1 | 68.6 | 55.3 | 47.8 | 42.8 | 38.6 | 35.1 | 31.7 | 28.7 | 26.0 | 23.6 | 222.4 | 5096.2 |
| 1990 | 68.1 | 44.5 | 41.0 | 32.8 | 28.4 | 25.4 | 22.7 | 20.5 | 18.5 | 16.8 | 15.2 | 143.5 | 4866.9 |
| 1991 | 53.6 | 36.4 | 23.6 | 21.6 | 17.4 | 15.0 | 13.2 | 11.9 | 10.8 | 9.7 | 8.8 | 82.5 | 4618.9 |
| 1992 | 28.4 | 22.0 | 14.8 | 9.7 | 8.8 | 7.1 | 6.0 | 5.3 | 4.9 | 4.4 | 4.0 | 36.4 | ${ }^{4270.3}$ |
| 1993 | 9.7 | 7.3 | 5.5 | 3.7 | 2.4 | 2.2 | 1.8 | 1.5 | 1.3 | 1.1 | 1.1 | 9.9 | ${ }_{3}^{3803.9}$ |
| 1994 | 2.2 | 1.3 | 1.1 | 0.9 | 0.4 | 0.4 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 1.5 | ${ }^{3251.6}$ |
| 1995 | 0.9 | 0.4 | 0.2 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 2913.6 |
| 1996 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2678.0 |
| 1997 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2790.2 |
| 1998 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2857.0 |
| 1999 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2989.9 |
| 2000 | 1.1 | ${ }^{0.2}$ | ${ }^{0.0}$ | ${ }^{0.0}$ | ${ }^{0.0}$ | ${ }^{0.0}$ | ${ }^{0.0}$ | 0.0 | 0.0 | 0.0 | ${ }^{0.0}$ | 0.0 | 3003.4 |
| ${ }_{2002}^{2001}$ | ${ }_{2.4}^{2.4}$ | 0.2 1.1 | ${ }_{0}^{0.0}$ | 0.0 0.0 | 0.0 0.0 | ${ }_{0}^{0.0}$ | 0.0 | 0.0 0.0 | 0.0 0.0 | 0.0 0.0 | 0.0 0.0 | 0.0 | 2827.9 30089 |
| 2003 | 1.5 | 0.9 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3216.8 |
| 2004 | 2.2 | 0.9 | 0.4 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3624.6 |
| 2005 | 3.1 | 1.3 | 0.4 | 0.4 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3966.6 |
| 2006 | 5.3 | 2.0 | 0.9 | 0.4 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 4178.9 |
| 2007 | 14.1 | 3.3 | 1.3 | 0.4 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 4279.8 |
| 2008 | 17.2 | 10.4 | 2.4 | 0.9 | 0.4 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 4509.3 |
| 2009 | 21.2 | 13.0 | 7.9 | 2.0 | 0.7 | 0.2 | 0.2 | ${ }^{0.0}$ | ${ }^{0.0}$ | 0.0 | 0.0 | 0.0 | 4715.7 |
| 2010 | 40.6 | 16.5 | 10.1 | 6.0 | 1.5 | 0.4 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | ${ }^{4903.3}$ |
| 2011 | 65.5 | 31.5 | 12.8 | 7.7 | 4.6 | 1.1 | 0.4 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 5056.3 |
| 2012 | 138.0 | 51.4 | 24.7 | 9.9 | 6.0 | 3.5 | 0.9 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 5192.3 |
| 2013 | 153.0 | 100.8 | 37.3 | 17.9 | 7.1 | 4.4 | 2.6 | 0.7 | 0.2 | 0.0 | 0.0 | 0.0 | 5154.2 |
| 2014 | 142.0 | 108.9 | 71.4 | 26.2 | 12.6 | 5.1 | 3.1 | 1.8 | 0.4 | 0.2 | 0.0 | 0.0 | 5106.8 |
| 2015 | 117.5 | 92.6 | 70.5 | 46.1 | 17.0 | 7.9 | 3.3 | 2.0 | 1.1 | 0.2 | 0.0 | 0.0 | 4934.6 |

Table 7.11. Estimated time series of status indicators, fishing mortality, biomass, and static SPR. Fishing mortality rate is apical $F$. Total biomass $(B, m t)$ is at the start of the year, and spawning biomass (SSB, female gonad weight $m t$ ) at the end of May (time of peak spawning). The MSST is defined by $M S S T=0.75 * S S B_{M S Y}$ with constant $M=0.1 . S P R$ is the static spawning potential ratio.

| Year | $F$ | $F / F_{M S Y}$ | $B$ | $B / B_{M S Y}$ | SSB | $S S B / S S B_{M S Y}$ | $S S B / M S S T$ | SPR |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1962 | 0.000 | 0.001 | 6820 | 0.914 | 95 | 4.314 | 5.752 | 0.996 |
| 1963 | 0.000 | 0.001 | 6868 | 0.920 | 96 | 4.367 | 5.822 | 0.996 |
| 1964 | 0.000 | 0.000 | 6913 | 0.926 | 97 | 4.403 | 5.870 | 0.999 |
| 1965 | 0.002 | 0.009 | 6956 | 0.932 | 97 | 4.434 | 5.911 | 0.974 |
| 1966 | 0.000 | 0.002 | 6985 | 0.936 | 98 | 4.461 | 5.948 | 0.995 |
| 1967 | 0.001 | 0.004 | 7021 | 0.941 | 98 | 4.490 | 5.987 | 0.988 |
| 1968 | 0.001 | 0.002 | 7052 | 0.945 | 99 | 4.517 | 6.022 | 0.993 |
| 1969 | 0.000 | 0.002 | 7082 | 0.949 | 100 | 4.543 | 6.057 | 0.994 |
| 1970 | 0.001 | 0.004 | 7110 | 0.953 | 100 | 4.566 | 6.089 | 0.989 |
| 1971 | 0.002 | 0.007 | 7134 | 0.956 | 101 | 4.586 | 6.115 | 0.979 |
| 1972 | 0.001 | 0.004 | 7152 | 0.958 | 101 | 4.603 | 6.138 | 0.988 |
| 1973 | 0.004 | 0.015 | 7172 | 0.961 | 101 | 4.616 | 6.155 | 0.958 |
| 1974 | 0.008 | 0.034 | 7176 | 0.962 | 101 | 4.614 | 6.152 | 0.908 |
| 1975 | 0.014 | 0.060 | 7156 | 0.959 | 101 | 4.590 | 6.120 | 0.850 |
| 1976 | 0.014 | 0.059 | 7093 | 0.951 | 100 | 4.553 | 6.071 | 0.852 |
| 1977 | 0.008 | 0.034 | 7022 | 0.941 | 99 | 4.528 | 6.037 | 0.908 |
| 1978 | 0.017 | 0.072 | 6971 | 0.934 | 99 | 4.499 | 5.999 | 0.824 |
| 1979 | 0.016 | 0.067 | 6855 | 0.919 | 97 | 4.445 | 5.927 | 0.834 |
| 1980 | 0.031 | 0.131 | 6728 | 0.902 | 96 | 4.365 | 5.821 | 0.717 |
| 1981 | 0.113 | 0.479 | 6511 | 0.872 | 90 | 4.116 | 5.487 | 0.400 |
| 1982 | 0.441 | 1.870 | 5919 | 0.793 | 72 | 3.277 | 4.369 | 0.157 |
| 1983 | 0.352 | 1.494 | 4267 | 0.572 | 51 | 2.313 | 3.085 | 0.183 |
| 1984 | 0.320 | 1.357 | 3424 | 0.459 | 39 | 1.774 | 2.365 | 0.195 |
| 1985 | 0.366 | 1.550 | 2900 | 0.389 | 31 | 1.400 | 1.866 | 0.172 |
| 1986 | 0.456 | 1.933 | 2466 | 0.331 | 24 | 1.086 | 1.448 | 0.153 |
| 1987 | 0.121 | 0.515 | 2112 | 0.283 | 21 | 0.948 | 1.265 | 0.383 |
| 1988 | 0.238 | 1.008 | 2255 | 0.302 | 20 | 0.933 | 1.244 | 0.242 |
| 1989 | 0.442 | 1.875 | 2312 | 0.310 | 18 | 0.837 | 1.116 | 0.157 |
| 1990 | 0.559 | 2.369 | 2208 | 0.296 | 16 | 0.709 | 0.946 | 0.134 |
| 1991 | 0.823 | 3.492 | 2095 | 0.281 | 13 | 0.610 | 0.813 | 0.106 |
| 1992 | 1.308 | 5.549 | 1937 | 0.260 | 11 | 0.512 | 0.683 | 0.082 |
| 1993 | 1.858 | 7.879 | 1725 | 0.231 | 9 | 0.423 | 0.563 | 0.068 |
| 1994 | 1.494 | 6.337 | 1475 | 0.198 | 8 | 0.359 | 0.478 | 0.076 |
| 1995 | 1.346 | 5.708 | 1322 | 0.177 | 7 | 0.308 | 0.411 | 0.081 |
| 1996 | 0.720 | 3.052 | 1215 | 0.163 | 7 | 0.301 | 0.402 | 0.114 |
| 1997 | 0.698 | 2.959 | 1266 | 0.170 | 7 | 0.334 | 0.446 | 0.114 |
| 1998 | 0.626 | 2.654 | 1296 | 0.174 | 8 | 0.362 | 0.483 | 0.125 |
| 1999 | 0.813 | 3.450 | 1356 | 0.182 | 8 | 0.369 | 0.492 | 0.106 |
| 2000 | 1.230 | 5.214 | 1362 | 0.183 | 7 | 0.328 | 0.437 | 0.084 |
| 2001 | 0.831 | 3.525 | 1283 | 0.172 | 7 | 0.298 | 0.397 | 0.103 |
| 2002 | 0.893 | 3.788 | 1365 | 0.183 | 7 | 0.312 | 0.416 | 0.099 |
| 2003 | 0.488 | 2.070 | 1459 | 0.196 | 8 | 0.361 | 0.482 | 0.141 |
| 2004 | 0.419 | 1.777 | 1644 | 0.220 | 10 | 0.439 | 0.585 | 0.151 |
| 2005 | 0.379 | 1.606 | 1799 | 0.241 | 11 | 0.512 | 0.683 | 0.156 |
| 2006 | 0.379 | 1.608 | 1895 | 0.254 | 13 | 0.571 | 0.761 | 0.164 |
| 2007 | 0.233 | 0.990 | 1941 | 0.260 | 14 | 0.635 | 0.847 | 0.241 |
| 2008 | 0.209 | 0.887 | 2045 | 0.274 | 16 | 0.713 | 0.950 | 0.266 |
| 2009 | 0.184 | 0.782 | 2139 | 0.287 | 17 | 0.781 | 1.041 | 0.285 |
| 2010 | 0.183 | 0.775 | 2224 | 0.298 | 18 | 0.837 | 1.116 | 0.288 |
| 2011 | 0.176 | 0.747 | 2294 | 0.307 | 19 | 0.884 | 1.179 | 0.293 |
| 2012 | 0.247 | 1.047 | 2355 | 0.316 | 20 | 0.903 | 1.204 | 0.230 |
| 2013 | 0.271 | 1.151 | 2338 | 0.313 | 19 | 0.888 | 1.184 | 0.217 |
| 2014 | 0.360 | 1.525 | 2316 | 0.310 | 19 | 0.850 | 1.134 | 0.179 |
| 2015 | . |  | 2238 | 0.300 | 20 | 0.933 | 1.244 |  |

Table 7.12. Selectivity at age for commercial handline (cH) landings, commerical longlines (cL) landings, recreational ( $r A$ ) landings, and the MARMAP longline survey (mm). TL is total length.

| Age | $\mathrm{TL}(\mathrm{mm})$ | $\mathrm{TL}(\mathrm{in})$ | cH | cL | rA | mm |
| ---: | ---: | ---: | :---: | :---: | :---: | :---: |
| 1 | 256.5 | 10.1 | 0.000 | 0.000 | 0.002 | 0.000 |
| 2 | 354.4 | 14.0 | 0.001 | 0.001 | 0.028 | 0.000 |
| 3 | 435.5 | 17.1 | 0.004 | 0.004 | 0.330 | 0.000 |
| 4 | 502.6 | 19.8 | 0.016 | 0.015 | 0.895 | 0.002 |
| 5 | 558.1 | 22.0 | 0.062 | 0.055 | 0.993 | 0.023 |
| 6 | 604.1 | 23.8 | 0.205 | 0.180 | 1.000 | 0.235 |
| 7 | 642.2 | 25.3 | 0.503 | 0.451 | 1.000 | 0.804 |
| 8 | 673.7 | 26.5 | 0.799 | 0.756 | 1.000 | 0.982 |
| 9 | 699.7 | 27.5 | 0.940 | 0.921 | 1.000 | 0.999 |
| 10 | 721.3 | 28.4 | 0.984 | 0.978 | 1.000 | 1.000 |
| 11 | 739.2 | 29.1 | 0.996 | 0.994 | 1.000 | 1.000 |
| 12 | 754.0 | 29.7 | 0.999 | 0.998 | 1.000 | 1.000 |
| 13 | 766.2 | 30.2 | 1.000 | 1.000 | 1.000 | 1.000 |
| 14 | 776.4 | 30.6 | 1.000 | 1.000 | 1.000 | 1.000 |
| 15 | 784.8 | 30.9 | 1.000 | 1.000 | 1.000 | 1.000 |
| 16 | 791.7 | 31.2 | 1.000 | 1.000 | 1.000 | 1.000 |
| 17 | 797.5 | 31.4 | 1.000 | 1.000 | 1.000 | 1.000 |
| 18 | 802.2 | 31.6 | 1.000 | 1.000 | 1.000 | 1.000 |
| 19 | 806.2 | 31.7 | 1.000 | 1.000 | 1.000 | 1.000 |
| 20 | 809.4 | 31.9 | 1.000 | 1.000 | 1.000 | 1.000 |
| 21 | 812.1 | 32.0 | 1.000 | 1.000 | 1.000 | 1.000 |
| 22 | 814.4 | 32.1 | 1.000 | 1.000 | 1.000 | 1.000 |
| 23 | 816.2 | 32.1 | 1.000 | 1.000 | 1.000 | 1.000 |
| 24 | 817.7 | 32.2 | 1.000 | 1.000 | 1.000 | 1.000 |
| 25 | 819.0 | 32.2 | 1.000 | 1.000 | 1.000 | 1.000 |
|  |  |  |  |  |  |  |

Table 7.13. Estimated instantaneous fishing mortality rate (per yr) at age.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1962 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1963 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1964 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1965 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 |
| 1966 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1967 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| 1968 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| 1969 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1970 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| 1971 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 |
| 1972 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |
| 1973 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.002 | 0.003 | 0.003 | 0.003 | 0.004 | 0.004 | 0.004 |
| 1974 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.004 | 0.006 | 0.007 | 0.008 | 0.008 | 0.008 | 0.008 |
| 1975 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.006 | 0.011 | 0.013 | 0.014 | 0.014 | 0.014 | 0.014 |
| 1976 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.006 | 0.011 | 0.013 | 0.014 | 0.014 | 0.014 | 0.014 |
| 1977 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.004 | 0.006 | 0.008 | 0.008 | 0.008 | 0.008 | 0.008 |
| 1978 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.008 | 0.013 | 0.016 | 0.017 | 0.017 | 0.017 | 0.017 |
| 1979 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.003 | 0.007 | 0.012 | 0.015 | 0.016 | 0.016 | 0.016 | 0.016 |
| 1980 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.006 | 0.015 | 0.024 | 0.029 | 0.030 | 0.031 | 0.031 | 0.031 |
| 1981 | 0.000 | 0.000 | 0.000 | 0.002 | 0.006 | 0.021 | 0.053 | 0.087 | 0.105 | 0.111 | 0.112 | 0.113 | 0.113 |
| 1982 | 0.000 | 0.000 | 0.002 | 0.007 | 0.025 | 0.081 | 0.203 | 0.337 | 0.408 | 0.432 | 0.438 | 0.440 | 0.441 |
| 1983 | 0.000 | 0.000 | 0.001 | 0.005 | 0.020 | 0.065 | 0.162 | 0.268 | 0.325 | 0.345 | 0.350 | 0.352 | 0.352 |
| 1984 | 0.000 | 0.000 | 0.002 | 0.006 | 0.019 | 0.059 | 0.147 | 0.244 | 0.296 | 0.313 | 0.318 | 0.320 | 0.320 |
| 1985 | 0.000 | 0.001 | 0.006 | 0.017 | 0.032 | 0.077 | 0.174 | 0.281 | 0.338 | 0.358 | 0.364 | 0.365 | 0.365 |
| 1986 | 0.000 | 0.000 | 0.002 | 0.007 | 0.025 | 0.083 | 0.208 | 0.347 | 0.421 | 0.446 | 0.453 | 0.455 | 0.456 |
| 1987 | 0.000 | 0.000 | 0.001 | 0.002 | 0.007 | 0.022 | 0.055 | 0.092 | 0.112 | 0.119 | 0.121 | 0.121 | 0.121 |
| 1988 | 0.000 | 0.000 | 0.001 | 0.004 | 0.014 | 0.044 | 0.109 | 0.181 | 0.219 | 0.233 | 0.236 | 0.237 | 0.238 |
| 1989 | 0.000 | 0.000 | 0.002 | 0.007 | 0.025 | 0.081 | 0.202 | 0.336 | 0.408 | 0.433 | 0.440 | 0.441 | 0.442 |
| 1990 | 0.000 | 0.001 | 0.002 | 0.009 | 0.031 | 0.102 | 0.255 | 0.425 | 0.515 | 0.547 | 0.555 | 0.558 | 0.558 |
| 1991 | 0.000 | 0.001 | 0.003 | 0.013 | 0.046 | 0.150 | 0.375 | 0.625 | 0.760 | 0.805 | 0.819 | 0.822 | 0.823 |
| 1992 | 0.000 | 0.001 | 0.006 | 0.022 | 0.074 | 0.239 | 0.597 | 0.994 | 1.207 | 1.280 | 1.301 | 1.306 | 1.308 |
| 1993 | 0.001 | 0.002 | 0.008 | 0.029 | 0.104 | 0.341 | 0.853 | 1.416 | 1.716 | 1.818 | 1.847 | 1.855 | 1.857 |
| 1994 | 0.000 | 0.002 | 0.007 | 0.026 | 0.087 | 0.276 | 0.686 | 1.138 | 1.380 | 1.462 | 1.486 | 1.492 | 1.494 |
| 1995 | 0.000 | 0.001 | 0.006 | 0.021 | 0.075 | 0.245 | 0.615 | 1.024 | 1.242 | 1.317 | 1.338 | 1.344 | 1.345 |
| 1996 | 0.000 | 0.001 | 0.004 | 0.013 | 0.042 | 0.132 | 0.329 | 0.547 | 0.664 | 0.704 | 0.716 | 0.719 | 0.719 |
| 1997 | 0.000 | 0.001 | 0.006 | 0.020 | 0.049 | 0.136 | 0.324 | 0.533 | 0.645 | 0.683 | 0.694 | 0.697 | 0.697 |
| 1998 | 0.000 | 0.001 | 0.003 | 0.010 | 0.035 | 0.114 | 0.285 | 0.475 | 0.577 | 0.612 | 0.622 | 0.625 | 0.625 |
| 1999 | 0.000 | 0.001 | 0.004 | 0.015 | 0.048 | 0.150 | 0.372 | 0.618 | 0.750 | 0.796 | 0.809 | 0.812 | 0.813 |
| 2000 | 0.000 | 0.001 | 0.007 | 0.024 | 0.074 | 0.228 | 0.563 | 0.934 | 1.134 | 1.203 | 1.222 | 1.228 | 1.229 |
| 2001 | 0.000 | 0.001 | 0.006 | 0.019 | 0.053 | 0.157 | 0.383 | 0.633 | 0.767 | 0.813 | 0.826 | 0.830 | 0.831 |
| 2002 | 0.000 | 0.001 | 0.006 | 0.019 | 0.055 | 0.168 | 0.412 | 0.681 | 0.825 | 0.874 | 0.888 | 0.892 | 0.893 |
| 2003 | 0.000 | 0.001 | 0.007 | 0.020 | 0.040 | 0.100 | 0.230 | 0.374 | 0.451 | 0.478 | 0.485 | 0.487 | 0.488 |
| 2004 | 0.000 | 0.001 | 0.010 | 0.029 | 0.047 | 0.097 | 0.205 | 0.325 | 0.389 | 0.410 | 0.417 | 0.418 | 0.419 |
| 2005 | 0.000 | 0.001 | 0.013 | 0.037 | 0.054 | 0.098 | 0.193 | 0.297 | 0.352 | 0.371 | 0.377 | 0.378 | 0.378 |
| 2006 | 0.000 | 0.001 | 0.008 | 0.024 | 0.040 | 0.085 | 0.183 | 0.293 | 0.351 | 0.371 | 0.377 | 0.379 | 0.379 |
| 2007 | 0.000 | 0.000 | 0.003 | 0.008 | 0.018 | 0.047 | 0.110 | 0.179 | 0.216 | 0.228 | 0.232 | 0.233 | 0.233 |
| 2008 | 0.000 | 0.000 | 0.001 | 0.003 | 0.012 | 0.038 | 0.095 | 0.159 | 0.193 | 0.205 | 0.208 | 0.209 | 0.209 |
| 2009 | 0.000 | 0.000 | 0.002 | 0.007 | 0.015 | 0.038 | 0.087 | 0.141 | 0.171 | 0.181 | 0.183 | 0.184 | 0.184 |
| 2010 | 0.000 | 0.000 | 0.002 | 0.006 | 0.014 | 0.036 | 0.085 | 0.140 | 0.169 | 0.179 | 0.182 | 0.182 | 0.183 |
| 2011 | 0.000 | 0.000 | 0.003 | 0.009 | 0.016 | 0.037 | 0.084 | 0.135 | 0.163 | 0.173 | 0.175 | 0.176 | 0.176 |
| 2012 | 0.000 | 0.000 | 0.003 | 0.010 | 0.021 | 0.052 | 0.118 | 0.191 | 0.229 | 0.242 | 0.246 | 0.247 | 0.247 |
| 2013 | 0.000 | 0.000 | 0.003 | 0.008 | 0.019 | 0.053 | 0.127 | 0.208 | 0.251 | 0.266 | 0.270 | 0.271 | 0.271 |
| 2014 | 0.000 | 0.000 | 0.002 | 0.007 | 0.022 | 0.068 | 0.168 | 0.276 | 0.333 | 0.352 | 0.358 | 0.359 | 0.360 |

Table 7.13. (Continued) Estimated instantaneous fishing mortality rate (per yr) at age.

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| \% |  |

Table 7.14. Estimated time series of fully selected fishing mortality rates for commercial handlines (F.cH), commercial longline (F.cL), recreational (F.rA) landings (L). Also shown is apical F, the maximum $F$ at age summed across fleets.

|  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | F.cH | F.cL | F.rA | Apical F |
| 1962 | 0.00004 | 0.00028 | 0.00000 | 0.00032 |
| 1963 | 0.00004 | 0.00026 | 0.00000 | 0.00031 |
| 1964 | 0.00001 | 0.00008 | 0.00000 | 0.00009 |
| 1965 | 0.00030 | 0.00188 | 0.00000 | 0.00217 |
| 1966 | 0.00006 | 0.00035 | 0.00000 | 0.00041 |
| 1967 | 0.00013 | 0.00082 | 0.00000 | 0.00095 |
| 1968 | 0.00008 | 0.00050 | 0.00000 | 0.00058 |
| 1969 | 0.00006 | 0.00041 | 0.00000 | 0.00047 |
| 1970 | 0.00013 | 0.00080 | 0.00000 | 0.00093 |
| 1971 | 0.00023 | 0.00148 | 0.00000 | 0.00172 |
| 1972 | 0.00014 | 0.00088 | 0.00000 | 0.00102 |
| 1973 | 0.00049 | 0.00308 | 0.00000 | 0.00357 |
| 1974 | 0.00111 | 0.00701 | 0.00000 | 0.00812 |
| 1975 | 0.00194 | 0.01214 | 0.00000 | 0.01408 |
| 1976 | 0.00198 | 0.01187 | 0.00000 | 0.01385 |
| 1977 | 0.00234 | 0.00577 | 0.00000 | 0.00811 |
| 1978 | 0.00839 | 0.00852 | 0.00000 | 0.01691 |
| 1979 | 0.00516 | 0.01067 | 0.00000 | 0.01584 |
| 1980 | 0.01397 | 0.01688 | 0.00000 | 0.03085 |
| 1981 | 0.03347 | 0.07943 | 0.00003 | 0.11294 |
| 1982 | 0.07663 | 0.36432 | 0.00000 | 0.44096 |
| 1983 | 0.04815 | 0.30411 | 0.00009 | 0.35234 |
| 1984 | 0.04855 | 0.27075 | 0.00080 | 0.32010 |
| 1985 | 0.04380 | 0.30921 | 0.01255 | 0.36556 |
| 1986 | 0.04876 | 0.40703 | 0.00004 | 0.45583 |
| 1987 | 0.01096 | 0.11034 | 0.00006 | 0.12137 |
| 1988 | 0.02297 | 0.21414 | 0.00068 | 0.23780 |
| 1989 | 0.04739 | 0.39476 | 0.00000 | 0.44216 |
| 1990 | 0.05499 | 0.50355 | 0.00011 | 0.55866 |
| 1991 | 0.07195 | 0.75134 | 0.00014 | 0.82343 |
| 1992 | 0.10465 | 1.20202 | 0.00178 | 1.30845 |
| 1993 | 0.28602 | 1.57184 | 0.00001 | 1.85787 |
| 1994 | 0.18077 | 1.30958 | 0.00384 | 1.49419 |
| 1995 | 0.15230 | 1.19361 | 0.00001 | 1.34592 |
| 1996 | 0.06513 | 0.65269 | 0.00191 | 0.71973 |
| 1997 | 0.06062 | 0.62632 | 0.01081 | 0.69775 |
| 1998 | 0.04624 | 0.57910 | 0.00037 | 0.62572 |
| 1999 | 0.05615 | 0.75434 | 0.00297 | 0.81347 |
| 2000 | 0.08660 | 1.13717 | 0.00574 | 1.22951 |
| 2001 | 0.06980 | 0.75421 | 0.00714 | 0.83115 |
| 2002 | 0.11372 | 0.77380 | 0.00569 | 0.89320 |
| 2003 | 0.03404 | 0.44005 | 0.01402 | 0.48812 |
| 2004 | 0.04121 | 0.35259 | 0.02513 | 0.41892 |
| 2005 | 0.04376 | 0.29936 | 0.03551 | 0.37863 |
| 2006 | 0.02248 | 0.33661 | 0.02017 | 0.37926 |
| 2007 | 0.03509 | 0.19296 | 0.00531 | 0.23335 |
| 2008 | 0.02035 | 0.18879 | 0.00001 | 0.20915 |
| 2009 | 0.01444 | 0.16487 | 0.00514 | 0.18445 |
| 2010 | 0.01435 | 0.16458 | 0.00382 | 0.18275 |
| 2011 | 0.01012 | 0.15934 | 0.00679 | 0.17625 |
| 2014 | 0.04755 | 0.19188 | 0.00751 | 0.24694 |
|  | 0.03463 | 0.23243 | 0.00431 | 0.27137 |
|  |  |  |  |  |

Table 7.15. Estimated time series of landings in gutted weight (1000 lb) for commercial handlines (cH), loglines (cL), and recreational ( $r A$ ).

| Year | cH | cL | rA | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1962 | 0.47 | 2.93 | 0.00 | 3.40 |
| 1963 | 0.44 | 2.78 | 0.00 | 3.22 |
| 1964 | 0.14 | 0.86 | 0.00 | 1.00 |
| 1965 | 3.21 | 20.10 | 0.00 | 23.31 |
| 1966 | 0.60 | 3.77 | 0.00 | 4.37 |
| 1967 | 1.43 | 8.93 | 0.00 | 10.36 |
| 1968 | 0.87 | 5.47 | 0.00 | 6.34 |
| 1969 | 0.71 | 4.47 | 0.00 | 5.18 |
| 1970 | 1.41 | 8.85 | 0.00 | 10.26 |
| 1971 | 2.62 | 16.40 | 0.00 | 19.02 |
| 1972 | 1.56 | 9.78 | 0.00 | 11.34 |
| 1973 | 5.47 | 34.26 | 0.00 | 39.73 |
| 1974 | 12.43 | 77.86 | 0.00 | 90.29 |
| 1975 | 21.57 | 134.04 | 0.00 | 155.61 |
| 1976 | 21.93 | 129.86 | 0.00 | 151.79 |
| 1977 | 25.74 | 62.78 | 0.00 | 88.51 |
| 1978 | 91.59 | 92.18 | 0.00 | 183.77 |
| 1979 | 55.87 | 114.29 | 0.00 | 170.16 |
| 1980 | 148.71 | 177.96 | 0.00 | 326.67 |
| 1981 | 334.99 | 786.89 | 0.41 | 1122.29 |
| 1982 | 598.61 | 2816.02 | 0.02 | 3414.65 |
| 1983 | 263.61 | 1644.04 | 0.59 | 1908.25 |
| 1984 | 202.87 | 1113.91 | 4.45 | 1321.23 |
| 1985 | 142.86 | 990.30 | 58.28 | 1191.43 |
| 1986 | 120.61 | 985.51 | 0.17 | 1106.28 |
| 1987 | 23.83 | 233.82 | 0.23 | 257.88 |
| 1988 | 50.14 | 454.29 | 2.43 | 506.86 |
| 1989 | 92.64 | 748.45 | 0.01 | 841.10 |
| 1990 | 86.07 | 762.77 | 0.35 | 849.20 |
| 1991 | 82.28 | 826.60 | 0.41 | 909.29 |
| 1992 | 81.48 | 883.42 | 5.00 | 969.91 |
| 1993 | 170.74 | 860.80 | 0.02 | 1031.56 |
| 1994 | 105.14 | 693.00 | 7.56 | 805.71 |
| 1995 | 82.85 | 598.72 | 0.02 | 681.59 |
| 1996 | 33.83 | 316.01 | 3.21 | 353.05 |
| 1997 | 33.81 | 326.81 | 20.35 | 380.97 |
| 1998 | 28.52 | 333.86 | 0.75 | 363.12 |
| 1999 | 37.70 | 472.05 | 5.78 | 515.53 |
| 2000 | 54.07 | 660.28 | 9.79 | 724.14 |
| 2001 | 38.45 | 388.61 | 11.49 | 438.54 |
| 2002 | 57.41 | 367.04 | 10.11 | 434.56 |
| 2003 | 18.41 | 222.56 | 29.43 | 270.40 |
| 2004 | 29.05 | 231.66 | 61.83 | 322.53 |
| 2005 | 41.14 | 262.50 | 96.98 | 400.61 |
| 2006 | 26.46 | 372.44 | 58.79 | 457.70 |
| 2007 | 49.59 | 258.71 | 16.66 | 324.97 |
| 2008 | 33.87 | 300.36 | 0.02 | 334.24 |
| 2009 | 27.36 | 299.97 | 18.20 | 345.53 |
| 2010 | 30.16 | 333.31 | 13.93 | 377.40 |
| 2011 | 22.89 | 348.40 | 25.46 | 396.75 |
| 2012 | 108.36 | 424.72 | 28.38 | 561.47 |
| 2013 | 75.04 | 489.61 | 16.04 | 580.69 |
| 2014 | 175.52 | 522.17 | 7.01 | 704.70 |

Table 7.16. Estimated time series of landings in numbers (1000 fish) for commercial handlines (cH), headboat (cL), and recreational ( $r A$ ).

| Year | cH | cL | rA | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1962 | 0.05 | 0.29 | 0.00 | 0.33 |
| 1963 | 0.04 | 0.27 | 0.00 | 0.31 |
| 1964 | 0.01 | 0.08 | 0.00 | 0.10 |
| 1965 | 0.31 | 1.95 | 0.00 | 2.27 |
| 1966 | 0.06 | 0.37 | 0.00 | 0.43 |
| 1967 | 0.14 | 0.87 | 0.00 | 1.01 |
| 1968 | 0.09 | 0.53 | 0.00 | 0.62 |
| 1969 | 0.07 | 0.43 | 0.00 | 0.50 |
| 1970 | 0.14 | 0.86 | 0.00 | 0.99 |
| 1971 | 0.25 | 1.59 | 0.00 | 1.84 |
| 1972 | 0.15 | 0.95 | 0.00 | 1.10 |
| 1973 | 0.53 | 3.31 | 0.00 | 3.84 |
| 1974 | 1.21 | 7.52 | 0.00 | 8.73 |
| 1975 | 2.10 | 12.95 | 0.00 | 15.05 |
| 1976 | 2.13 | 12.55 | 0.00 | 14.68 |
| 1977 | 2.50 | 6.07 | 0.00 | 8.57 |
| 1978 | 8.92 | 8.92 | 0.00 | 17.83 |
| 1979 | 5.44 | 11.06 | 0.00 | 16.50 |
| 1980 | 14.49 | 17.22 | 0.00 | 31.71 |
| 1981 | 32.65 | 76.20 | 0.05 | 108.90 |
| 1982 | 58.76 | 274.57 | 0.00 | 333.33 |
| 1983 | 26.45 | 163.63 | 0.08 | 190.16 |
| 1984 | 20.92 | 113.83 | 0.63 | 135.39 |
| 1985 | 15.20 | 104.30 | 9.00 | 128.50 |
| 1986 | 13.35 | 107.87 | 0.03 | 121.24 |
| 1987 | 2.75 | 26.65 | 0.04 | 29.44 |
| 1988 | 5.93 | 53.10 | 0.41 | 59.45 |
| 1989 | 11.23 | 89.72 | 0.00 | 100.95 |
| 1990 | 10.78 | 94.50 | 0.07 | 105.35 |
| 1991 | 10.96 | 108.74 | 0.10 | 119.79 |
| 1992 | 12.27 | 131.16 | 1.19 | 144.62 |
| 1993 | 28.91 | 144.45 | 0.00 | 173.37 |
| 1994 | 17.82 | 116.95 | 1.73 | 136.50 |
| 1995 | 13.46 | 96.83 | 0.00 | 110.29 |
| 1996 | 5.38 | 49.91 | 0.80 | 56.09 |
| 1997 | 5.32 | 51.00 | 5.01 | 61.33 |
| 1998 | 4.48 | 51.96 | 0.17 | 56.61 |
| 1999 | 5.91 | 73.31 | 1.27 | 80.48 |
| 2000 | 8.44 | 102.35 | 2.20 | 113.00 |
| 2001 | 5.96 | 59.81 | 2.83 | 68.60 |
| 2002 | 8.96 | 56.85 | 2.65 | 68.46 |
| 2003 | 2.94 | 35.28 | 7.54 | 45.76 |
| 2004 | 4.63 | 36.60 | 14.85 | 56.08 |
| 2005 | 6.40 | 40.47 | 21.90 | 68.77 |
| 2006 | 3.96 | 55.31 | 12.75 | 72.02 |
| 2007 | 7.14 | 36.94 | 3.49 | 47.57 |
| 2008 | 4.70 | 41.35 | 0.00 | 46.05 |
| 2009 | 3.69 | 40.08 | 3.43 | 47.21 |
| 2010 | 3.96 | 43.40 | 2.54 | 49.91 |
| 2011 | 2.93 | 44.24 | 4.62 | 51.79 |
| 2012 | 13.58 | 52.77 | 5.17 | 71.52 |
| 2013 | 9.32 | 60.25 | 2.96 | 72.54 |
| 2014 | 21.99 | 64.73 | 1.32 | 88.05 |

Table 7.17. Estimated landings at age in gutted weight (1000 lb)




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Table 7.18. (Continued) Estimated total landings at age in numbers (1000 fish)


Table 7.19. Estimated status indicators, benchmarks, and related quantities from the Beaufort catch-age model, conditional on estimated current selectivities averaged across fisheries. Rate estimates $(F)$ are in units of $\mathrm{y}^{-1}$; status indicators are dimensionless; and biomass estimates are in units of metric tons or gutted pounds, as indicated. Spawning stock biomass (SSB) and minimum stock size threshold (MSST) are measured by total gonad weight of mature females. Symbols, abbreviations, and acronyms are listed in Appendix A.

| Quantity | Units | Estimate |
| :--- | :--- | :--- |
| MSST | mt | 16.45 |
| $F_{\mathrm{MSY}}$ | $\mathrm{y}^{-1}$ | 0.24 |
| $85 \% F_{\mathrm{MSY}}$ | $\mathrm{y}^{-1}$ | 0.20 |
| $75 \% F_{\mathrm{MSY}}$ | $\mathrm{y}^{-1}$ | 0.18 |
| $65 \% F_{\mathrm{MSY}}$ | $\mathrm{y}^{-1}$ | 0.15 |
| MSY | 1000 lb | 560 |
| $Y a t 85 \% F_{M S Y}$ | 1000 lb | 557 |
| $Y a t 75 \% F_{M S Y}$ | 1000 lb | 551 |
| $Y a t 65 \% F_{M S Y}$ | 1000 lb | 540 |
| $D_{M S Y}$ | 1000 fish | 70 |
| $B_{\mathrm{MSY}}$ | mt | 2574 |
| SSB | MSY | mt |
| $R_{\mathrm{MSY}}$ | 1000 age | 21.93 |
| $F_{30 \%}$ | $\mathrm{y}^{-1}$ | 327 |
| $F_{40 \%}$ | $\mathrm{y}^{-1}$ | 0.17 |
| $F_{50 \%}$ | $\mathrm{y}^{-1}$ | 0.11 |
| $F_{2012-2014}$ | - | 0.08 |
| $F_{2012-2014} / F_{\mathrm{MSY}}$ | - | 0.29 |
| $F_{2014}$ | $\mathrm{y}^{-1}$ | 1.22 |
| $F_{2014} / F_{\mathrm{MSY}}$ | - | 0.36 |
| $S S B_{2014}$ | mt | 1.53 |
| $S S B_{2014} / \mathrm{MSST}$ | - | 18.65 |
| $S S B_{2014} / \mathrm{SSB}_{\mathrm{MSY}}$ | - | 1.13 |

Table 7.20. Results from sensitivity runs of the Beaufort catch-at-age model. Current F represented by geometric mean of last three assessment years. See text for full description of sensitivity runs. Fleet abbreviations include rA for recreational, cH for commercial handline, and cL for

Table 7.21. Projection results with fishing mortality rate fixed at $F=$ Fcurrent starting in 2017. $R=$ number of age- 1 recruits (in 1000 s), $N$ $=$ total stock abundance (1000 fish), $F=$ fishing mortality rate (per year), $S=$ spawning stock ( mt ), $B=$ total stock biomass (mt), $L=$ landings expressed in numbers (1000 fish) and gutted weight ( $w$, in 1000 lb ), and pr. $75=$ proportion of stochastic projection replicates with $\mathrm{SSB} \geq \mathrm{MSST}$ using the $75 \%$ definition of MSST. The extension b indicates expected values (deterministic) from the base run; the extension med indicates median values from the stochastic projections.

| Year | R.b | R.med | N.b | N.med | F.b | F.med | S.b(mt) | S.med (mt) | B.b(mt) | B.med(mt) | L.b(n) | L.med(n) | L.b(w) | L.med(w) | pr. 75 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 316 | 313 | 1457 | 1517 | 0.2889 | 0.2632 | 18 | 18 | 2238 | 2294 | 69 | 66 | 537 | 524 | 0.5012 |
| 2016 | 314 | 307 | 1460 | 1514 | 0.2889 | 0.2632 | 18 | 18 | 2248 | 2290 | 69 | 66 | 534 | 509 | 0.4988 |
| 2017 | 313 | 309 | 1462 | 1512 | 0.2889 | 0.2632 | 18 | 18 | 2259 | 2283 | 70 | 66 | 537 | 506 | 0.4976 |
| 2018 | 314 | 305 | 1464 | 1509 | 0.2889 | 0.2632 | 18 | 18 | 2270 | 2282 | 71 | 66 | 543 | 508 | 0.4956 |
| 2019 | 314 | 309 | 1465 | 1513 | 0.2889 | 0.2632 | 18 | 18 | 2277 | 2278 | 72 | 67 | 548 | 511 | 0.4938 |
| 2020 | 314 | 305 | 1466 | 1502 | 0.2889 | 0.2632 | 18 | 18 | 2281 | 2274 | 72 | 67 | 552 | 513 | 0.4939 |
| 2021 | 314 | 306 | 1466 | 1500 | 0.2889 | 0.2632 | 18 | 18 | 2283 | 2268 | 72 | 67 | 554 | 512 | 0.4927 |
| 2022 | 314 | 303 | 1466 | 1496 | 0.2889 | 0.2632 | 18 | 18 | 2284 | 2261 | 72 | 67 | 555 | 511 | 0.4918 |
| 2023 | 314 | 305 | 1466 | 1500 | 0.2889 | 0.2632 | 18 | 18 | 2284 | 2255 | 72 | 67 | 555 | 509 | 0.4898 |
| 2024 | 314 | 301 | 1467 | 1490 | 0.2889 | 0.2632 | 18 | 18 | 2284 | 2248 | 72 | 66 | 555 | 506 | 0.4877 |

Table 7.22. Projection results with fishing mortality rate fixed at $F=F_{\text {MSY }}$ starting in 2017. $R=$ number of age-1 recruits (in 1000 s), $N=$ total stock abundance (1000 fish), $F=$ fishing mortality rate (per year), $S=$ spawning stock ( $m t$ ), $B=$ total stock biomass ( $m$ t), $L=$ landings expressed in numbers (1000 fish) and gutted weight ( $w$, in 1000 lb ), and pr. $75=$ proportion of stochastic projection replicates with $\mathrm{SSB} \geq \mathrm{MSST}$ using the $75 \%$ definition of MSST. The extension b indicates expected values (deterministic) from the base run; the extension med indicates median values from the stochastic projections.

| Year | R.b | R.med | N.b | N.med | F.b | F.med | S.b(mt) | S.med(mt) | B.b(mt) | B.med(mt) | L.b(n) | L.med(n) | L.b(w) | L.med(w) | pr. 75 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 316 | 314 | 1457 | 1519 | 0.2889 | 0.2652 | 18 | 18 | 2238 | 2313 | 69 | 66 | 537 | 522 | 0.5012 |
| 2016 | 314 | 309 | 1460 | 1520 | 0.2889 | 0.2652 | 18 | 18 | 2248 | 2303 | 69 | 65 | 534 | 510 | 0.4964 |
| 2017 | 313 | 308 | 1462 | 1524 | 0.2358 | 0.2157 | 18 | 18 | 2259 | 2302 | 58 | 48 | 448 | 376 | 0.5138 |
| 2018 | 315 | 313 | 1476 | 1525 | 0.2358 | 0.2157 | 19 | 18 | 2315 | 2303 | 61 | 52 | 469 | 404 | 0.5471 |
| 2019 | 317 | 315 | 1488 | 1537 | 0.2358 | 0.2157 | 19 | 18 | 2359 | 2309 | 63 | 55 | 487 | 427 | 0.5816 |
| 2020 | 319 | 319 | 1499 | 1547 | 0.2358 | 0.2157 | 20 | 18 | 2392 | 2322 | 65 | 57 | 502 | 444 | 0.6154 |
| 2021 | 320 | 317 | 1507 | 1555 | 0.2358 | 0.2157 | 20 | 18 | 2418 | 2330 | 66 | 59 | 513 | 459 | 0.6446 |
| 2022 | 321 | 323 | 1515 | 1564 | 0.2358 | 0.2157 | 20 | 19 | 2439 | 2338 | 66 | 59 | 520 | 468 | 0.6702 |
| 2023 | 322 | 324 | 1522 | 1568 | 0.2358 | 0.2157 | 21 | 19 | 2455 | 2351 | 67 | 60 | 525 | 473 | 0.6923 |
| 2024 | 323 | 325 | 1527 | 1579 | 0.2358 | 0.2157 | 21 | 19 | 2470 | 2361 | 67 | 60 | 529 | 477 | 0.7127 |

Table 7.23. Projection results with fishing mortality rate fixed at $F=75 \% F_{\mathrm{MSY}}$ starting in 2017. $R=$ number of age-1 recruits (in 1000 s), $N$ $=$ total stock abundance (1000 fish), $F=$ fishing mortality rate (per year), $S=$ spawning stock ( $m$ t), $B=$ total stock biomass ( mt ), $L=$ landings expressed in numbers ( 1000 fish) and gutted weight ( $w$, in 1000 lb ), and pr. $75=$ proportion of stochastic projection replicates with $\mathrm{SSB} \geq \mathrm{MSST}$ using the $75 \%$ definition of MSST. The extension b indicates expected values (deterministic) from the base run; the extension med indicates median values from the stochastic projections.

| Year | R.b | R.med | N.b | N.med | F.b | F.med | S.b(mt) | S.med (mt) | B.b(mt) | B.med(mt) | L.b(n) | L.med(n) | L.b(w) | L.med(w) | pr. 75 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 316 | 311 | 1457 | 1507 | 0.2889 | 0.2644 | 18 | 18 | 2238 | 2285 | 69 | 66 | 537 | 522 | 0.5016 |
| 2016 | 314 | 303 | 1460 | 1506 | 0.2889 | 0.2644 | 18 | 18 | 2248 | 2286 | 69 | 66 | 534 | 511 | 0.4976 |
| 2017 | 313 | 306 | 1462 | 1505 | 0.1768 | 0.1611 | 19 | 18 | 2259 | 2283 | 45 | 37 | 344 | 286 | 0.5204 |
| 2018 | 316 | 310 | 1490 | 1530 | 0.1768 | 0.1611 | 20 | 19 | 2367 | 2353 | 48 | 41 | 374 | 323 | 0.5741 |
| 2019 | 320 | 315 | 1516 | 1551 | 0.1768 | 0.1611 | 21 | 19 | 2457 | 2392 | 51 | 45 | 402 | 356 | 0.6300 |
| 2020 | 324 | 320 | 1538 | 1574 | 0.1768 | 0.1611 | 22 | 20 | 2532 | 2428 | 54 | 48 | 426 | 380 | 0.6798 |
| 2021 | 326 | 324 | 1558 | 1590 | 0.1768 | 0.1611 | 23 | 21 | 2593 | 2464 | 56 | 49 | 445 | 398 | 0.7208 |
| 2022 | 329 | 328 | 1576 | 1606 | 0.1768 | 0.1611 | 23 | 21 | 2645 | 2500 | 57 | 51 | 460 | 411 | 0.7573 |
| 2023 | 330 | 326 | 1592 | 1621 | 0.1768 | 0.1611 | 24 | 22 | 2689 | 2537 | 58 | 52 | 471 | 420 | 0.7911 |
| 2024 | 332 | 332 | 1605 | 1634 | 0.1768 | 0.1611 | 24 | 22 | 2728 | 2569 | 59 | 52 | 480 | 428 | 0.8219 |

Table 7.24. Projection results with fishing mortality rate fixed at $P *=0.35$ starting in 2017. $R=$ number of age- 1 recruits (in 1000 s), $N=$ total stock abundance (1000 fish), $F=$ fishing mortality rate (per year), $S=$ spawning stock ( $m$ t), $B=$ total stock biomass ( $m$ t), $L=$ landings expressed in numbers (1000 fish) and gutted weight ( $w$, in 1000 lb ), and pr. $75=$ proportion of stochastic projection replicates with $\mathrm{SSB} \geq \mathrm{MSST}$ using the $75 \%$ definition of MSST. All values except year and probabilities are medians from the stochastic projections.
Table 7.25. Projection results with fishing mortality rate fixed at $P *=0.5$ starting in 2017. $R=$ number of age- 1 recruits (in 1000 s), $N=$ total stock abundance (1000 fish), $F=$ fishing mortality rate (per year), $S=$ spawning stock ( $m t$ ), $B=$ total stock biomass ( $m$ t), $L=$ landings expressed in numbers (1000 fish) and gutted weight ( $w$, in 1000 lb ), and pr. $75=$ proportion of stochastic projection replicates with $\mathrm{SSB} \geq \mathrm{MSST}$ using the $75 \%$ definition of MSST. All values except year and probabilities are medians from the stochastic projections.

## 8 Figures

Figure 8.1. Mean length at age (mm) and estimated 95\% confidence interval of the population.


Figure 8.2. Observed (open circles) and estimated (line, solid circles) commercial handline landings (1000 lb gutted weight).


Figure 8.3. Observed (open circles) and estimated (line, solid circles) commercial longline landings (1000 lb gutted weight).


Figure 8.4. Observed (open circles) and estimated (line, solid circles) recreational landings (1000 whole lbs).


Figure 8.5. Observed (open circles) and estimated (line, solid circles) index of abundance from the commercial longline fishery.


Figure 8.6. Observed (open circles) and estimated (solid circles) index of abundance from the MARMAP horizontal longline survey.


Figure 8.7. Observed (open circles) and estimated (solid line) annual length and age compositions by fleet or survey. In panels indicating the data set, lcomp refers to length compositions, acomp to age compositions, cH to commercial handline, cL to commercial longline, rA to recreational, and mm to MARMAP. $N$ indicates the number of trips from which individual fish samples were taken.


Figure 8.7. (Continued) Observed (open circles) and estimated (solid line) annual length and age compositions by fleet or survey. In panels indicating the data set, lcomp refers to length compositions, acomp to age compositions, $c H$ to commercial handline, cL to commercial longline, rA to recreational, and mm to MARMAP. $N$ indicates the number of trips from which individual fish samples were taken.


Figure 8.7. (Continued) Observed (open circles) and estimated (solid line) annual length and age compositions by fleet or survey. In panels indicating the data set, lcomp refers to length compositions, acomp to age compositions, $c H$ to commercial handline, cL to commercial longline, rA to recreational, and mm to MARMAP. $N$ indicates the number of trips from which individual fish samples were taken.


Figure 8.7. (Continued) Observed (open circles) and estimated (solid line) annual length and age compositions by fleet or survey. In panels indicating the data set, lcomp refers to length compositions, acomp to age compositions, $c H$ to commercial handline, cL to commercial longline, rA to recreational, and mm to MARMAP. $N$ indicates the number of trips from which individual fish samples were taken.


Figure 8.7. (Continued) Observed (open circles) and estimated (solid line) annual length and age compositions by fleet or survey. In panels indicating the data set, lcomp refers to length compositions, acomp to age compositions, $c H$ to commercial handline, cL to commercial longline, rA to recreational, and mm to MARMAP. $N$ indicates the number of trips from which individual fish samples were taken.


Figure 8.8. Estimated abundance at age at start of year.


Figure 8.9. Top panel: Estimated recruitment of age-1 fish. Horizontal dashed line indicates $R_{\text {MSY }}$. Bottom panel: log recruitment residuals.



Figure 8.10. Estimated biomass at age at start of year.


Figure 8.11. Top panel: Estimated total biomass (metric tons) at start of year. Horizontal dashed line indicates $B_{\mathrm{MSY}}$. Bottom panel: Estimated spawning stock (gonad biomass of mature females) at time of peak spawning.


Figure 8.12. Selectivities of commercial fleets, 1962-2014. Top panel: commercial handline, Bottom panel: commercial longline.


Figure 8.13. Selectivities of the recreational fleet and MARMAP survey 1962-2014. Top panel: recreational, Bottom panel: MARMAP longline survey.


Figure 8.14. Average selectivity from the terminal assessment year weighted by geometric mean Fs from the last three assessment years, and used in computation of benchmarks and central-tendency projections.


Figure 8.15. Estimated fully selected fishing mortality rate (per year) by fishery. cL refers to commercial longline, $c H$ to commercial handline, and rA to recreational.


Figure 8.16. Estimated landings in gutted weight by fishery from the catch-age model. cL refers to commercial longline, $c H$ to commercial handline, and $r A$ to recreational. Horizontal dashed line in the top panel corresponds to the point estimate of MSY.




Figure 8.17. Estimated landings in numbers by fishery from the catch-age model. cL refers to commercial longline, $c H$ to commercial handline, and $r A$ ro recreational.



|  |  |
| :---: | :---: |

Figure 8.18. Top panel: Beverton-Holt spawner-recruit curves, with and without lognormal bias correction. The expected (upper) curve was used for computing management benchmarks. Years within panel indicate year of recruitment generated from spawning biomass one year prior. Bottom panel: log of recruits (number age-1 fish) per spawner (mature female gonad weight) as a function of spawners.


Figure 8.19. Probability densities of spawner-recruit quantities R0 (unfished recruitment of age-1 fish), steepness, unfished spawners per recruit, and standard deviation of recruitment residuals in log space. Solid vertical lines represent point estimates or values from the base run; dashed vertical lines represent medians from the MCB runs.


Figure 8.20. Estimated time series of static spawning potential ratio, the annual equilibrium spawners per recruit relative to that at the unfished level. Horizontal dashed line indicates the equilibrium MSY level.


Figure 8.21. Top panel: yield per recruit. Bottom panel: spawning potential ratio (spawning biomass per recruit relative to that at the unfished level), from which the $y \%$ levels provide $F_{y \%}$. Both curves are based on average selectivity from the end of the assessment period.



Figure 8.22. Top panel: equilibrium landings. The peak occurs where fishing rate is $F_{\mathrm{MSY}}=0.24$ and equilibrium landings are MSY $=560$ (1000 lb gutted weight). Bottom panel: equilibrium spawning biomass. Both curves are based on average selectivity from the end of the assessment period.


Figure 8.23. Equilibrium landings as a function of equilibrium biomass, which itself is a function of fishing mortality rate. The peak occurs where equilibrium biomass is $B_{\mathrm{MSY}}=2574 \mathrm{mt}$ and equilibrium landings are MSY $=560$ (1000 lb gutted weight).


Figure 8.24. Probability densities of MSY-related benchmarks from MCB analysis of the Beaufort Assessment Model. Vertical lines represent point estimates from the base run.


Figure 8.25. Estimated time series relative to benchmarks. Solid line indicates estimates from base run of the Beaufort Assessment Model; gray error bands indicate $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of the MCB trials. Top panel: spawning biomass relative to the minimum stock size threshold (MSST). Bottom panel: F relative to $F_{\mathrm{MSY}}$.



Figure 8.26. Probability densities of terminal status estimates from MCB analysis of the Beaufort Assessment Model. Vertical lines represent point estimates from the base run.




Figure 8.27. Phase plot of terminal status estimates from MCB analysis of the Beaufort Assessment Model. The intersection of crosshairs indicates estimates from the base run; lengths of crosshairs defined by $5^{\text {th }}$ and $95^{\text {th }}$ percentiles.


Figure 8.28. Top panel: Sensitivity of F/Fmsy to model configuration changes made during this update. Bottom panel: Sensitivity of SSB/MSST to model configuration changes made during this update. The solid line indicates sensitivity run $S 1$ (effect of using SEDAR-25 multinomial likelihood), S2 (setting recreational selectivity equal to commercial handline), and S3 (inclusion of biased age compositions from 1990s longline samples) as described in 4.1.6.


Figure 8.29. Retrospective analyses. Sensitivity to terminal year of data (sensitivity runs S4-S11 as described in 4.1.6). Fishing mortality rate, where solid circles show geometric mean of terminal three years, as used to compute fishing status.


Figure 8.30. Retrospective analyses. Sensitivity to terminal year of data (sensitivity runs S4-S11 as described in 4.1.6). Biomass time series.


Figure 8.31. Retrospective analyses. Sensitivity to terminal year of data (sensitivity runs S4-S11 as described in 4.1.6). Spawning stock biomass time series.


Figure 8.32. Retrospective analyses. Sensitivity to terminal year of data (sensitivity runs S4-S11 as described in 4.1.6). Recruitment time series.


Figure 8.33. Retrospective analyses. Sensitivity to terminal year of data (sensitivity runs S4-S11 as described in 4.1.6). Relative fishing mortality rate time series.


Figure 8.34. Retrospective analyses. Sensitivity to terminal year of data (sensitivity runs S4-S11 as described in 4.1.6). Relative spawning stock biomass time series


Figure 8.35. Comparison of stock status indicators among base and continuity runs described in Section 4.2.6.


Figure 8.36. Projection results under scenario 1 -fishing mortality rate fixed at $F=F_{\text {current }}$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{t h}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians.


Figure 8.37. Projection results under scenario 1 -fishing mortality rate fixed at $F=F_{\text {current }}$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians. In the bottom panel, the curve represents the proportion of projection replicates for which SSB exceeds the replicate-specific MSST. Horizontal lines drawn at 0.5 and 0.7 for reference.


Figure 8.38. Projection results under scenario 2-fishing mortality rate fixed at $F=F_{\text {MSY }}$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians.


Figure 8.39. Projection results under scenario 2—fishing mortality rate fixed at $F=F_{\text {MSY }}$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians. In the bottom panel, the curve represents the proportion of projection replicates for which SSB exceeds the replicate-specific MSST. Horizontal lines drawn at 0.5 and 0.7 for reference.


Figure 8.40. Projection results under scenario 3-fishing mortality rate fixed at $F=75 \% F_{\text {MSY }}$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians.


Figure 8.41. Projection results under scenario 3-fishing mortality rate fixed at $F=75 \% F_{\text {MSY }}$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{t h}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians. In the bottom panel, the curve represents the proportion of projection replicates for which SSB exceeds the replicate-specific MSST. Horizontal lines drawn at 0.5 and 0.7 for reference.


Figure 8.42. Projection results under scenario 4-fishing mortality rate fixed at $P *=0.35$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians.


Figure 8.43. Projection results under scenario 4-fishing mortality rate fixed at $P *=0.35$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians. In the bottom panel, the curve represents the proportion of projection replicates for which SSB exceeds the replicate-specific MSST. Horizontal lines drawn at 0.5 and 0.7 for reference.


Figure 8.44. Projection results under scenario 5-fishing mortality rate fixed at $P *=0.5$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians.


Figure 8.45. Projection results under scenario 5-fishing mortality rate fixed at $P *=0.5$. Expected values (base run) represented by dotted solid lines, medians represented by dashed lines with open circles, and uncertainty represented by thin lines corresponding to $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of replicate projections. Solid horizontal lines mark MSY-related quantities; dashed horizontal lines represent corresponding medians. In the bottom panel, the curve represents the proportion of projection replicates for which SSB exceeds the replicate-specific MSST. Horizontal lines drawn at 0.5 and 0.7 for reference.


## Appendix A Abbreviations and symbols

Table A.1. Acronyms and abbreviations used in this report

| Symbol | Meaning |
| :---: | :---: |
| ABC | Acceptable Biological Catch |
| AW | Assessment Workshop (here, for tilefish) |
| ASY | Average Sustainable Yield |
| $B$ | Total biomass of stock, conventionally on January 1 |
| BAM | Beaufort Assessment Model (a statistical catch-age formulation) |
| CPUE | Catch per unit effort; used after adjustment as an index of abundance |
| CV | Coefficient of variation |
| DW | Data Workshop (here, for tilefish) |
| $F$ | Instantaneous rate of fishing mortality |
| $F_{\text {MSY }}$ | Fishing mortality rate at which MSY can be attained |
| FL | State of Florida |
| GA | State of Georgia |
| GLM | Generalized linear model |
| K | Average size of stock when not exploited by man; carrying capacity |
| kg | Kilogram(s); 1 kg is about 2.2 lb . |
| klb | Thousand pounds; thousands of pounds |
| lb | Pound(s); 1 lb is about 0.454 kg |
| m | Meter(s); 1 m is about 3.28 feet. |
| M | Instantaneous rate of natural (non-fishing) mortality |
| MARMAP | Marine Resources Monitoring, Assessment, and Prediction Program, a fishery-independent data collection program of SCDNR |
| MCB | Monte Carlo/Bootstrap, an approach to quantifying uncertainty in model results |
| MFMT | Maximum fishing-mortality threshold; a limit reference point used in U.S. fishery management; often based on $F_{\mathrm{MSY}}$ |
| mm | Millimeter(s); 1 inch $=25.4 \mathrm{~mm}$ |
| MRFSS | Marine Recreational Fisheries Statistics Survey, a data-collection program of NMFS, predecessor of MRIP |
| MRIP | Marine Recreational Information Program, a data-collection program of NMFS, descended from MRFSS |
| MSST | Minimum stock-size threshold; a limit reference point used in U.S. fishery management. The SAFMC has defined MSST for tilefish as $(1-M) \mathrm{SSB}_{\mathrm{MSY}}=0.7 \mathrm{SSB}_{\mathrm{MSY}}$. |
| MSY | Maximum sustainable yield (per year) |
| mt | Metric ton(s). One mt is 1000 kg , or about 2205 lb . |
| $N$ | Number of fish in a stock, conventionally on January 1 |
| NC | State of North Carolina |
| NMFS | National Marine Fisheries Service, same as "NOAA Fisheries Service" |
| NOAA | National Oceanic and Atmospheric Administration; parent agency of NMFS |
| OY | Optimum yield; SFA specifies that OY $\leq$ MSY. |
| PSE | Proportional standard error |
| $R$ | Recruitment |
| SAFMC | South Atlantic Fishery Management Council (also, Council) |
| SC | State of South Carolina |
| SCDNR | Department of Natural Resources of SC |
| SDNR | Standard deviation of normalized residuals |
| SEDAR | SouthEast Data Assessment and Review process |
| SFA | Sustainable Fisheries Act; the Magnuson-Stevens Act, as amended |
| SL | Standard length (of a fish) |
| SRHS | Southeast Region Headboat Survey, conducted by NMFS-Beaufort laboratory |
| SPR | Spawning potential ratio |
| SSB | Spawning stock biomass; mature biomass of males and females |
| $\mathrm{SSB}_{\text {MSY }}$ | Level of SSB at which MSY can be attained |
| TIP | Trip Interview Program, a fishery-dependent biodata collection program of NMFS |
| TL | Total length (of a fish), as opposed to FL (fork length) or SL (standard length) |
| VPA | Virtual population analysis, an age-structured assessment |
| WW | Whole weight, as opposed to GW (gutted weight) |
| yr | Year(s) |

# Appendix B Final SEDAR South Atlantic Tilefish Assessment Terms of Reference Approved 9-28-2015 



SouthEast Data, Assessment, and Review

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## SEDAR South Atlantic Tilefish Assessment* <br> Terms of Reference

1. Update the approved SEDAR 25 South Atlantic Tilefish ("Golden Tilefish") base model with data through 2014.
2. Document any changes or corrections made to the model and input datasets and provide updated input data tables. Provide commercial and recreational landings and discards in pounds and numbers.
3. Update model parameter estimates and their variances, model uncertainties, and estimates of stock status and management benchmarks. Identify other sources of scientific uncertainty that are not already included in the model uncertainties.
4. Provide stock projections, including a pdf for biological reference point estimates and yield separated for landings and discards reported in pounds and numbers. Projection outputs shall include relevant population parameters including recruitment, spawning and stock biomass, population abundance, exploitation rates and the probability that biomass and exploitation exceed reference values for MFMT and MSST. Projection criteria:

- Probability of overfishing $\left(\mathrm{P}^{*}\right)=50 \%$ and $35 \%$.
- If the stock is determined to be overfished, provide the probability of rebuilding within mandated time periods based on $F=0, F=F c u r r e n t, F=F r e b u i l d, ~ a n d ~ F=F m s y$. For this analysis, Frebuild is defined as the maximum exploitation that rebuilds the stock in the maximum time allowed; evaluate at Prebuild $=50 \%$ and $65 \%$. In addition to reporting yield and stock status as described above, for this projection also report the probability of SSB>SSBmsy.

5. Develop a stock assessment update report to address these TORS and fully document the input data and results of the stock assessment update.

NOTE: The intent of the assessment update approach is to expedite appraisals of stock status by using only the methods and data sets used in the base model and approved during the preceding SEDAR assessment of that stock. Accordingly, it is not the intent of this update to resolve any outstanding issues identified in the previous SEDAR 25 assessment.
*This assessment is following the update assessment approach.


## Appendix C Parameter estimates from the Beaufort Assessment Model

## \%

\# Number of parameters $=188$ Objective function value $=-3030.83$ Maximum gradient component $=8.39261 \mathrm{e}-005$
\# Linf:
825.100000000
\# K:
0.189000000000
\# to:
-0.470000000000
\# Linf_f:
806.300000000
\# K_f:
0.167000000000
\# to_f:
\# to_f:
-0.470000000000
\# len_cv_val:
\# len_cv_val:
0.150819142821
\# log_Nage_dev
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
0.000000000000 .00000000000
\# log_R0:
12.800534442
\# steep:
\# steep:
0.840000000000
\# rec_sigma:
\# rec_sigma:
0.330013121980
\# R_autocorr:
0.00000000000
\# log_rec_dev:
-0.188330436602-0.155115349017
$-0.176686408015-0.330819241480$
$-0.439357031221-0.416747816119$
$-0.159855627911-0.302683327635$
$-0.321406902021-0.554602748140$
$-0.2434032446660 .0796627275268$
0.7587671494790 .331513078878
$-0.305076606272-0.256129115945$
$-0.171617546922-0.0448728410623$
$0.456362914022 \quad 0.0422065542115$
$-0.274948972357-0.0784469698168$
0.04126369491890 .528290177897

$\begin{array}{lll}0.523848934528 & 0.472448874850 \\ 0.424599061416 & 0.373935776134\end{array}$
0.4245990614160 .373935776134
0.3914629691550 .243396511634
$-0.0681572688094-0.179500970636$
\# selpar_L50_cH
6.99231613396
\# selpar_slope_cH

1. 36760587934
\# selpar_L50_cL:
7.14721289359
\# selpar_slope_cL:
2. 32462982627
\# selpar_L50_ra
3.24863183885
\# selpar_slope_rA:
2.84629573669
\# selpar_L50_mm
6.45592314032
\# selpar_slope_mm
2.58969298498
\# log_q_cL:
-6.77307997818
\# log_q_mm:
-6.94553772716
\# M_constant:
0.10830000000
\# log_avg_F_cH
-4.79993246435
\# log_F_dev_ch:
5.22354117794-5.28781207116
$-6.46558020569-3.32348039021$
$-5.00237262903-4.14675499552$
$-4.64331155471-4.85094852090$
$-4.17114406915-3.55857409447$
$-4.07910454532-2.82740467323$
$-2.00545231549-1.44759156619$
$-1.42251535014-1.25713600316$
$0.0186593537473-0.465985138499$
. 528986451848 1. 40296122646
2.231187143701 .76641098029
1.774804459661 .67192458804
1.779018406440 .286353286001
$1.02619426128 \quad 1.75061922533$
$1.89937447585 \quad 2.16818840735$
$\begin{array}{ll}1.89937447585 & 2.16818840735\end{array}$
2.542830721393 .54824145130
3.08942420536
2.91801227791
.089424053 2.91801227791
2.068610519821 .99674689904
.726100727151 .92022229115
2.353452631362 .13785400201
2.625913517891 .41984243524
. 1010874145577 1.45003384
1.004741455771 .45004164958
0.9052586337440 .562296757630
$0.556022008378 \quad 0.207157954238$
1.753854107131 .43700294164
2.36857620822
\# log_avg_F_cL
-2.89682218885
\# log_F_dev_cL:
-5.28285905990-5.34420997566
$-6.52587728437-3.38123290566$
$-5.06113552861-4.20505619352$
$-4.70086996392-4.90817296906$
$-4.22992642144-3.61684626065$
$-4.22992642144-3.61684626065$
. $06362169918-1.51417769570$
$2.06362169918-1.51417769570$
$-1.53718787106-2.25864996323$
$-1.86808399836-1.64335095243$
3. 887111858 寝 1.706
1.887111858991 .70646586941
1.590259029641 .72307396662
1.997947289790 .692673100832
1.355719962571 .96735699170
2.210748934062 .61092870165
3.080828353073 .34906792735
3.166531720763 .07380769629
2.470175266702 .42893340127
2.350543379702 .61491335020
3.025364570022 .61473272054
2.640375554022 .07596152737
1.854365151961 .69070470652
$1.80798080482 \quad 1.25153613795$
$\begin{array}{ll}1.80798080482 & 1.25153613795 \\ 1.22970614604 & 1.09420074963\end{array}$
1.229706146041 .09420074963
. 092438075701.06011027149
1.245941215641 .43765586024
1.58667823441
\# log_avg_F_rA:
\# log_F_dev_rA
-3.17059795603-6.07494736049
$-2.27512113437-0.0385896390333$
$2.71698654487-2.95388448022$
$-2.56511523758-0.191306200127$
$-5.26955263330-1.97551142004$
$-1.813086566760 .763117260887$
$-4.591852598741 .53220254089$
-4.24484902766 0.832369028904
$2.56744293945-0.795132784843$
1.276732825191 .93497616059
$2.15263688160 \quad 1.92564896479$
2.827723868483 .41126812859
3.75721111066 3.19161281460
$1.85611069710-4.94362853672$
$1.85611069710-4.94362853672$
1.824184444231 .52810492474
2.102225411612 .20410090518
1.64738121530

F_init:
0.0100000000000


[^0]:    ${ }^{1}$ Abbreviations and acronyms used in this report are defined in Appendix A

